

**Tonga Island Marine Reserve,
Abel Tasman National Park
update of biological monitoring,
1993 – 2007**

Research, Survey and Monitoring Report Number 484

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Davidson Environmental Ltd.
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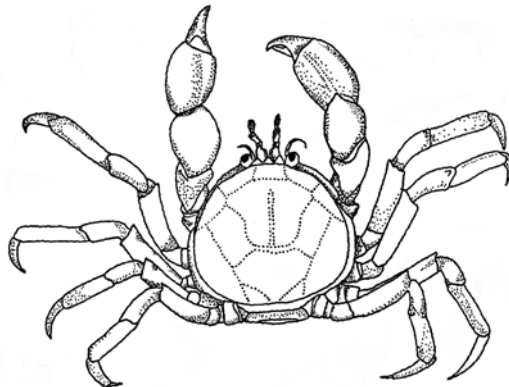
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1.0 INTRODUCTION

A number of studies have documented the impacts of marine reserve establishment in temperate areas of the world (Bell, 1983; McCormick and Choat, 1987; Buxton and Smale, 1989; Garcia-Rubies and Zabala, 1990; Bennett and Attwood, 1991; MacDiarmid and Breen, 1993; Dufour *et al.*, 1995; Edgar and Barrett, 1997; Kelly *et al.*, 2000; Willis *et al.*, 2000, Davidson *et al.*, 2002, Shears *et al.*, 2006). Some studies have reported the recovery of populations of particular species (Bennett and Attwood, 1993; Cole and Keuskamp, 1998; Kelly *et al.*, 2000), while others have reported mixed responses to the establishment of a marine reserve (Cole *et al.*, 1990). Particular studies have considered implications of marine reserve for fisheries management (eg. Hilborn *et al.* 2004). Several reasons have been suggested for the variation in responses to marine reservation. Kelly *et al.* (2000) suggested that failure to demonstrate the effects of protection did not mean that change had not occurred, but may reflect limitations in the sampling methodology and analysis. These authors suggested that limitations in methodology and survey design rendered conclusions from many of the published studies in temperate marine reserves questionable.

Most studies investigating temperate marine reserves have been conducted over relatively short time scales, while others have investigated only a small number of sites inside and outside these reserves. Few studies have extended beyond three years, with on-going funding being the principal problem. The time frame of university-based student studies may also limit the duration of such marine reserve investigations.

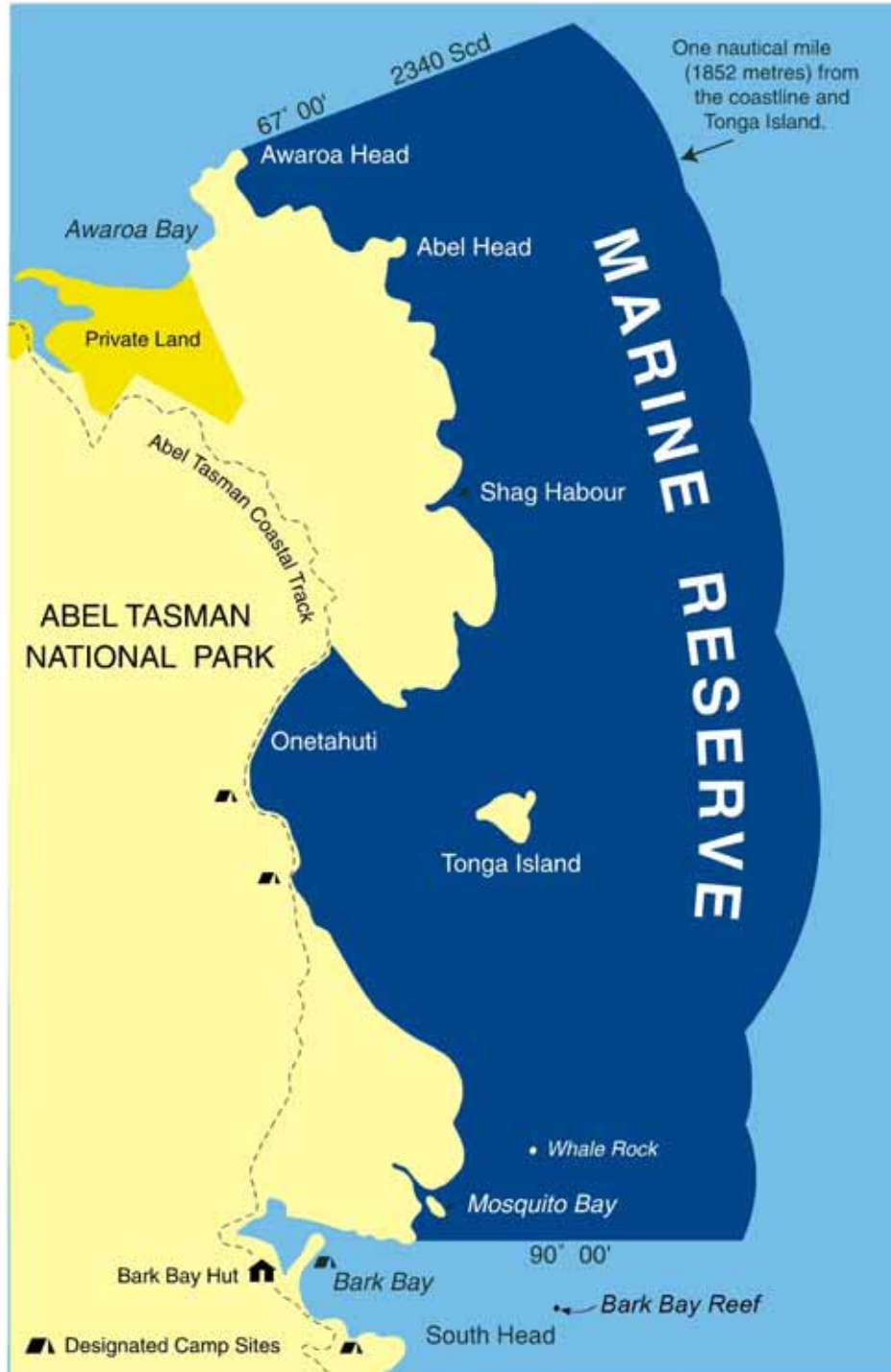
Problems in detecting the impacts of marine reserve establishment have been further compounded by the mixed responses that occur in species or communities (Cole, 1994; Willis *et al.*, 2000). Based on temperate marine reserve studies, it is clear that the same changes will not necessarily occur in all reserves, or perhaps any two reserves. For example, MacDiarmid and Breen (1993) reported conflicting results for spiny lobsters (*Jasus edwardsii*) at two marine reserves and three control locations in north-eastern New Zealand. Results from the marine reserve areas revealed highest densities in one marine reserve and lowest densities in the other. The authors concluded that low abundance in the second reserve was probably due to a lack of suitable habitat and its location with respect to the continental shelf and major currents.

The present study is the continuation of an on-going monitoring programme investigating the impact of a marine reserve (i.e. removal of fishing) at Tonga Island Marine Reserve (located centrally along the Abel Tasman National Park coastline) (see Map 1). To date, data has been collected over a period of 15 years from Tonga Island Marine Reserve and adjacent control sites (Davidson 1999, 2001, present study). The baseline monitoring study was conducted in December 1993, with another sampling event in December 1994. These results were presented



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by Davidson in 1999, with an additional monitoring update report produced by Davidson in 2001. The present biological report presents results from all years (1993 – 2007).



Map 1. Tonga Island Marine Reserve.



Blue cod (*Parapercis colias*) was one of the key species selected for study, primarily because it is a widespread and important recreational fishery resource and it could be studied using established underwater visual methods. In addition, blue cod have been the focus of movement studies locally (Mace and Johnson, 1983; Cole *et al.*, 2000) and in southern New Zealand (Carbines, 1998, 1999). Blue cod have also been the focus of fisheries-related research in the Marlborough Sounds (Blackwell, 1997, 1998).

Spiny lobster (*Jasus edwardsii*) was also specifically selected as many studies within and outside New Zealand have shown that this species often responds to protection (MacDiarmid and Breen, 1993; Kelly *et al.*, 2000; Davidson *et al.*, 2002, Shears *et al.*, 2006).

2.0 STUDY AREA

The Abel Tasman coastline and the Tonga Island Marine Reserve are located centrally within Tasman and Golden Bays, Nelson. Tonga Island Marine Reserve was established in November 1993 (Map 1). The reserve is 1835 hectares in size and extends one nautical mile, or 1.852 km, offshore from mean high water. The marine reserve boundaries are from the headland immediately north of Bark Bay to Awaroa Head, and includes the shoreline of all islands and stacks within its boundaries.

The coastline within the marine reserve and to the north and south is next to the Abel Tasman National Park (see Dennis, 1985, for review). This coastline is sheltered from large ocean swells and is predominantly influenced by wind-generated waves which quickly subside with a drop or change in wind direction (Davidson 1992, Davidson and Chadderton 1994). High sediment input from the hill catchments both within and adjacent to this coast (e.g. Motueka River catchment), combined with regular sea-breezes and large tides (4.7 m extreme high tide), maintain water clarity at relatively low levels (approximately 2-8 m horizontal distance). Water temperatures range from 10°C to 22°C (Dix, 1970).

The Abel Tasman intertidal coastline is comprised of rocky shores and headlands interspersed by coarse sand beaches. A number of estuaries are located along this coastline (Davidson 1990). Subtidally, rocky reefs may extend to a depth of 14 m and are bordered by a low gradient soft sediment benthos. Subtidal soft shores are primarily characterised by broken shell and coarse sands with a silt component. Granite boulder and bedrock substrata dominate the Abel Tasman rocky coast. Less than 1% of rocky shores along the National Park coast are composed of limestone (Davidson, 1992, Davidson and Chadderton 1994). Davidson and Chadderton (1994) reported that subtidal communities found on limestone were dramatically different to communities inhabiting granite shores. Sample sites within the present study were confined to granite rock and adjacent offshore sediment habitats located in comparable shore aspects and depths.



3.0 SAMPLING SUMMARY SINCE 2001

Data collected in relation to the Tonga Island Marine Reserve has been presented in two previous reports (Davidson, 1999, 2001). This report presents data from 2001 through to 2007 and provides comparisons over the entire sampling period since 1993. The location of sites sampled since 2001 are presented in Table 1 and Figures 1-6. Sampling history is summarised in Table 2. Coordinates for sites sampled prior to 2001 but not sampled since 2001 are presented in Davidson (1999, 2001).

Table 1. Reef fish and lobster sites sampled since 2001 (shore profiles sampled 2001 only).

| Type | Coordinates | Location | Treatment | Substratum |
|---------------|---------------------------|---------------------------|-----------|-------------------------------|
| Reef fish | 40 47.05806, 172 59.90864 | Separation Point, 1 | Control | Boulder, cobble |
| Reef fish | 40 48.22062, 173 00.59191 | Totaranui north, 2 | Control | Bedrock |
| Reef fish | 40 48.92921, 173 01.00318 | Totaranui Reef, 3 | Control | Bedrock |
| Reef fish | 40 51.12408, 173 02.62163 | Awaroa, 4 | Control | Bedrock, boulder |
| Reef fish | 40 51.15995, 173 02.77066 | Canoe Bay, 5 | Reserve | Boulder, cobble |
| Reef fish | 40 51.39304, 173 03.44764 | Abel Head, 6 | Reserve | Boulder, cobble, bedrock |
| Reef fish | 40 51.72980, 173 03.67184 | Cottage Loaf, 7 | Reserve | Boulder, cobble, bedrock |
| Reef fish | 40 53.10127, 173 03.51041 | Reef Pt. 8 | Reserve | Bedrock, boulder |
| Reef fish | 40 53.45573, 173 04.13343 | Tonga Is. 9 | Reserve | Bedrock, boulder |
| Reef fish | 40 54.23512, 173 03.74018 | Foul Pt. 10 | Reserve | Bedrock, boulder |
| Reef fish | 40 54.58073, 173 04.14347 | Whale Rock, 11 | Reserve | Bedrock, boulder |
| Reef fish | 40 55.10164, 173 04.30913 | Bark Bay Reef, 12 | Control | Boulder, cobble, bedrock |
| Reef fish | 40 56.35684, 173 03.70098 | Totara Rocks, 13 | Control | Bedrock, boulder |
| Lobster | 40 47.05806, 172 59.90864 | Separation Point, 1 | Control | Boulder, cobble |
| Lobster | 40 51.12408, 173 02.62163 | Awaroa, 2 | Control | Bedrock, boulder |
| Lobster | 40 51.27420, 173 02.94736 | Canoe Bay, 3 | Reserve | Bedrock, boulder |
| Lobster | 40 51.70603, 173 03.66512 | Cottage Loaf, 4 | Reserve | Boulder, cobble, bedrock |
| Lobster | 40 53.43768, 173 04.13356 | Tonga Is. 5 | Reserve | Bedrock, boulder |
| Lobster | 40 54.23333, 173 03.75072 | Foul Pt. 6 | Reserve | Bedrock, boulder |
| Lobster | 40 54.58073, 173 04.14347 | Whale Rock, 7 | Reserve | Bedrock, boulder |
| Lobster | 40 55.10164, 173 04.30913 | Bark Bay Reef, 8 | Control | Boulder, cobble, bedrock |
| Lobster | 40 56.28063, 173 03.74358 | Totara Rocks, 9 | Control | Bedrock, boulder |
| Shore profile | 40 46.92579, 172 59.81331 | Separation Pt. (north), 1 | Control | Bedrock, boulder, sand, shell |
| Shore profile | 40 47.04179, 172 59.90323 | Separation Pt. 2 | Control | Bedrock, boulder, sand |
| Shore profile | 40 48.67368, 173 00.81487 | Totaranui, 3 | Control | Bedrock, cobble, sand |
| Shore profile | 40 51.09813, 173 02.66039 | Awaroa Head, 4 | Control | Bedrock, boulder, sand, shell |
| Shore profile | 40 51.45515, 173 03.48456 | Brereton Cove, 5 | Reserve | Bedrock, boulder, sand, shell |
| Shore profile | 40 53.33155, 173 03.98949 | Tonga Is. 6 | Reserve | Boulder, cobble, sand, shell |
| Shore profile | 40 53.92518, 173 03.27818 | Onetahuti, 7 | Reserve | Boulder, sand |
| Shore profile | 40 54.57867, 173 04.16220 | Whale Rock, 8 | Reserve | Bedrock, boulder, sand, shell |
| Shore profile | 40 55.09445, 173 04.30222 | Bark Bay Reef, 9 | Reserve | Bedrock, boulder, sand, shell |



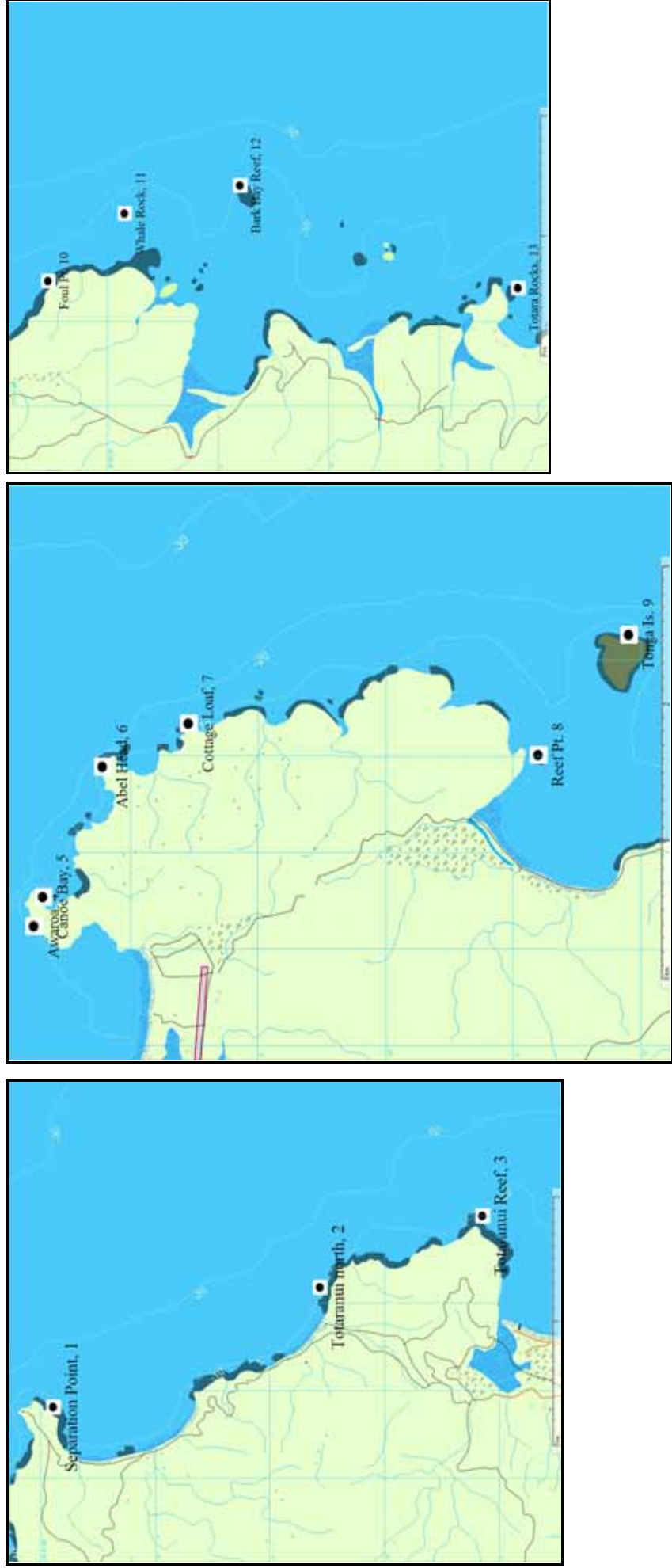
Table 2. Sampling history along the Abel Tasman coastline from 1993 to 2007.

| Group | 1993 | 1994 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
|--------------------------|------|------|------|------|------|------|------|------|------|------|------|------|
| Reef fish density | + | + | | + | + | + | + | | + | + | + | + |
| Reef fish size | | | | | + | + | + | | + | + | + | + |
| Lobster density | | + | | + | + | | + | | + | + | + | + |
| Lobster size and sex | | | + | + | + | | + | | + | | + | + |
| Benthic quadrats | + | | | | | + | | | | | | |
| Kina density | + | | | | | | + | | | | | |
| Kina size | + | | | | | | + | | | | | |
| Cooks turban density | + | | | | | | + | | | | | |
| Cooks turban size | + | | | | | | + | | | | | |
| Topshell density | | | | | | | + | | | | | |
| Cats eye density | + | | | | | | + | | | | | |
| Cats eye size | + | | | | | | | | | | | |
| Limpet density | | | | | | | + | | | | | |
| Scallop size and density | | + | | + | | | | + | | | + | |
| Horse mussel density | | + | | + | | | | + | | | + | |
| Shore profiles | | | | | | + | | | | | | |



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Figure 1. Location of reef fish sites sampled from 2001-2007 for reserve (RF) and control (CF) sites along the Abel Tasman coastline.





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Figure 2. Location of lobster sites sampled from 2001-2007 for reserve (RL) and control (CL) sites along the Abel Tasman coastline.

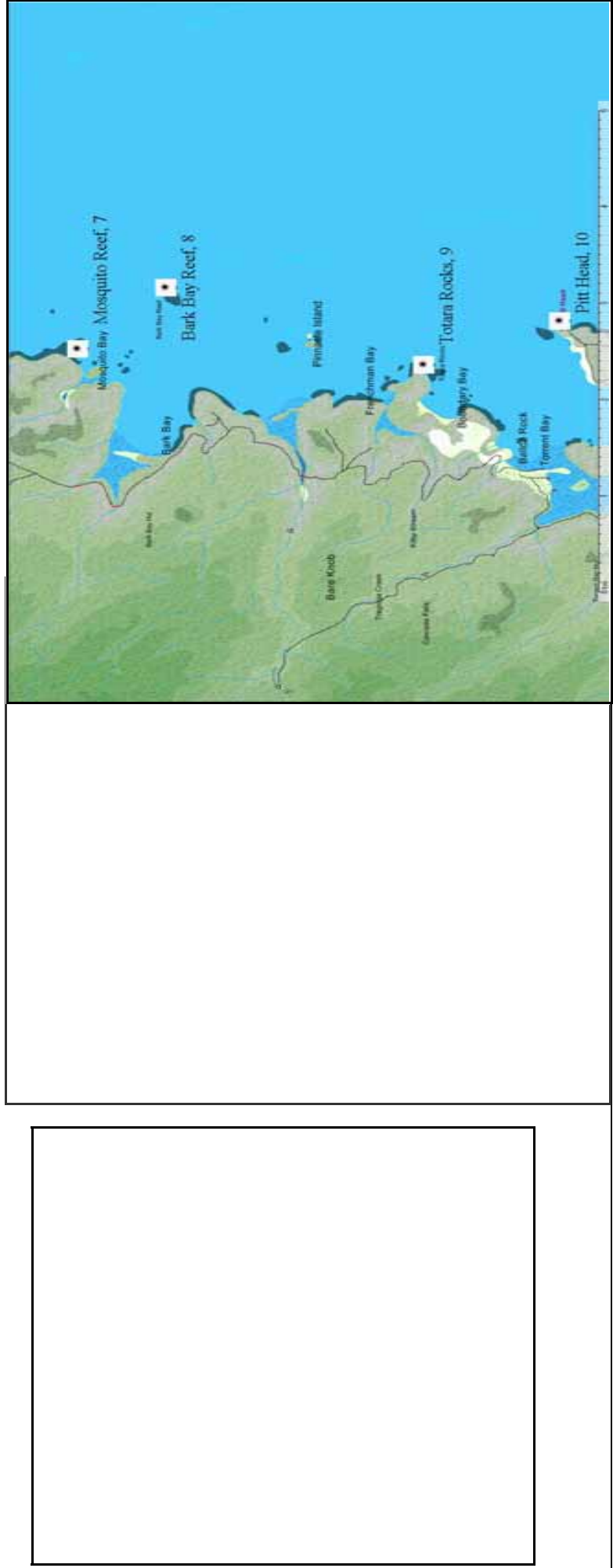




Figure 3. Location of shore profiles from reserve (RP) and control (CP) sites collected in March 2001 along the Abel Tasman coastline.





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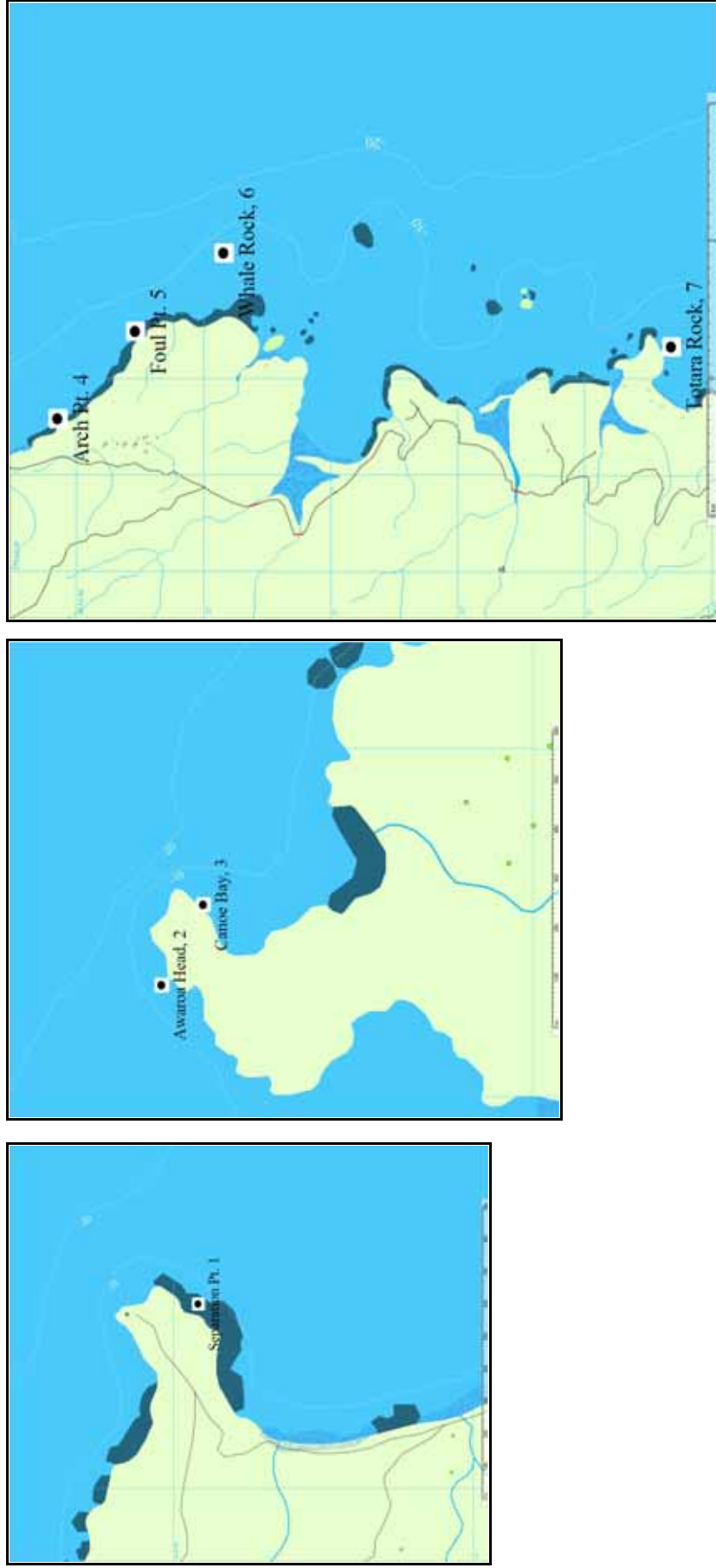
Figure 4. Location of benthic quadrat samples from reserve (RB) and control (CB) in March 2001 along the Abel Tasman coastline.





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Figure 5. Location of quadrat samples for key species density (kina, Cook's turban, topshell, cats-eye, and limpet) and size (kina) collected in April 2002 along the Abel Tasman coastline.





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Figure 6. Location of scallop and horse mussel sample sites collected in March 2003 and April 2006 along the Abel Tasman coastline. Note: T2 reserve site was not sampled in 2006.





4.0 MATERIALS AND METHODS

4.1 Fish density derived from underwater visual transects

Blue cod and other reef fish abundance were investigated using established underwater visual transect methods (Bell, 1983; McCormick and Choat, 1987; Choat *et al.*, 1988; Buxton and Smale, 1989; Cole *et al.*, 1990; Cole, 1994; Willis *et al.*, 2000).

All transects were established parallel to shore in boulder and reef habitat at depths from 5 m to 12 m. Blue cod sizes were estimated by divers to the nearest centimetre of body length. In most years, lengths of other reef fish such as blue moki, red moki, tarakihi and butterflyfish were also recorded. Divers recorded the presence and density of other reef fish, excluding triplefins and cave- and crevice-dwelling species. Diver estimation of blue cod size was standardised using plastic model fish, the sizes of which were validated underwater prior to fieldwork.

At each site, a lead weight at the start of the transect line was dropped onto the substrate within the designated depth range. The line was automatically reeled off a spool as the diver holding the spool swam away from the lead weight. At a distance of 5 m from the weight (indicated by a marker on the line), the diver started counting fish present within an estimated 2 m wide x 2 m high x 30 m long “tunnel”. A total of 12 replicates were collected on each occasion. Transects were swum at a constant slow speed, but fast enough to ensure that swimming fish did not overtake the divers. Underwater visibility was at least 4.5 m horizontal distance for the collection of fish transect data.

From 2000 to 2007, divers estimated the length of particular reef fish during the collection of fish density transects (i.e. blue cod, tarakihi, blue moki, red moki, butterflyfish, and magpie moki). In 2005, additional measurement of reef fish were collected from two reserve and two control sites. At each of these additional samples, two divers estimated reef fish length during 30 minute free swims. Where possible, all fish sighted during the swim were estimated for length.

4.2 Spiny lobster density, sex and size

Spiny lobster density was investigated from a variety of sites located within Tonga Island Marine Reserve and adjacent control sites in December 1994, December 1998, February 1999, May 1999, November 1999, March 2000, December 2000, December 2002, February 2004, April 2006 and February 2007 (Figure 2, Table 1). Lobster data collected between February 1999 and December 2000 were presented in Davidson (2001).



In 1994, density of spiny lobsters was sampled from 30 x 4 m quadrats, with between 3-13 quadrats sampled per site depending on the number of divers available to collect data.

From 1998 onwards, lobster density, size and sex were sampled using 25 x 4 m quadrats (100 m²), with a total of 10 quadrats per site. Since 1998, quadrats were also depth-stratified so that five quadrats were counted at a depth of 4-7 m and five were counted at 9-11 m at each site. In most years five reserve and five control sites were sampled.

Lobster size estimate methodology also changed over the duration of the monitoring programme. In the 1994 survey, total body length was estimated and grouped into 4 size classes: juvenile (< 150 mm), small (150 – 250 mm), medium (250 – 350 mm), and large (> 350 mm). From 1998 onwards, carapace length (CL) replaced total body length estimates and was estimated to the nearest 5 mm.

Lobster quadrats were haphazardly placed and oriented within the depth stratum. Underwater visibility was at least 2.5 m during all lobster counts. Two divers independently searched all crevices, caves and cracks within each quadrat using a dive torch. The size and sex of lobsters encountered were recorded. A core group of five divers were involved in most of the surveys. The size and sex of some lobsters could not be measured because they were deeply concealed beneath boulders or within caves. As a result, the number of lobsters in density and size data do not correspond.

4.3 Kina density and size

Kina (*Evechinus chloroticus*) density and size data have been collected on two sample occasions (1993 and 2002). Kina were sampled from either rock or rubble habitat free of foliose macroalgae. All kina encountered within quadrats were measured *in situ* using callipers to the nearest 5 mm length.

In 1993 control kina density and size were sampled from a total of nine stations across six sites while reserve kina were sampled from 14 stations across 12 reserve sites. At each station, kina were counted and measured from 34 to 66 haphazard 1 m² quadrats collected from predetermined depth ranges (see Davidson 1999 for methods).

In 2002, kina were sampled from a reduced number of sites (i.e. four reserve and three control). Kina data were again collected from within a predetermined depth range (see Appendix 16 for depths), but only one station was sampled at each site. Kina density was determined from a total of sixty quadrats at each site. Where insufficient kina were encountered in quadrats, additional



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size-frequency samples were collected from adjacent areas by divers thoroughly searching rocky habitats. The maximum width of between 79 to 125 kina were measured at each site.

4.4 Cook's turban size and density

Cook's turban snails (*Cookia sulcata*) density and size measurements were collected in 1993 and 2002.

Snail maximum length (mm) and density were collected from predetermined depth ranges. Snail data were collected from rock or rubble substratum without foliose macroalgae. All snails encountered within the haphazardly deployed 1m² quadrats were either measured and counted. Where sufficient numbers for measurement purposes were not collected from quadrats, additional individuals were measured from adjacent areas by divers thoroughly searching each site. Each Cook's turban was measured using callipers to the nearest 1 mm.

In 1993, snails were measured from a total of 10 stations across six control sites while reserve samples were collected from 12 stations across 11 sites. Between 34 and 176 individuals were measured in 1993.

In 2002, a reduced number of sites were sampled (four reserve and three control). Snail sizes and density were collected from within a predetermined depth range (see Appendix 17 for depths), but only one station was sampled at each site. Snail density was determined from a total of 60 quadrats at each site.

4.5 Cats-eye, topshell and limpet

Cats-eye snail (*Turbo smaragdus*), topshell (*Trochus* sp.) and limpet (*Cellana* sp.) abundance were collected in 1993 and 2002. Invertebrates were sampled from rock or rubble barrens (i.e. no cover of foliose macroalgae).

In 1993 the density of these species were established from 21 to 46 haphazard 1m² quadrats sampled from predetermined depth ranges (see Appendix 18-20 for depths). In 2002, the density of these species were established from 60 quadrats sampled from four reserve and three control sites.

Cats eye snail sizes were collected in 1993 (Davidson, 1999). Control invertebrates were sampled from eight stations across six sites and seven reserve stations across six sites (see Davidson 1999). In 1993, all cats-eyes encountered within quadrats were measured with



additional individuals being measured from adjacent areas when sufficient numbers were not present in quadrats. Cats-eyes were measured *in situ* using callipers to the nearest 1 mm.

4.6 Scallop density and size and horse mussel density

Scallops and horse mussels were counted and the maximum width of scallops measured in March 2003 and April 2006. These data had been previously sampled in December 1994 and September 1999 (see Davidson 2001). The same methodology was used in all years, however in 2003 an additional reserve site was sampled 2003 (Figure 6).

In 2003 and 2006, two control sites located in Bark Bay (Bark Bay centre and Bark Bay north). In 2006 two reserve sites were located in Tonga Roadstead (Tonga north-west and Tonga south) (Figure 2, Table 1). In 2003 an additional reserve site was sampled in Tonga Roadstead (Tonga north T1).

At each site 10 replicate sets, each consisting of 50 contiguous 1 m² quadrats were sampled. Each set of 50 quadrats were deployed by divers instructed to swim at least 10 m distance from the previous set of quadrats. For each set of quadrats divers used a different compass bearing to the previous set to ensure divers remained within approximately 150 m distance of the start point.

5.0 RESULTS AND DISCUSSION

5.1 Underwater visual surveys

Visual fish count data collected from December 2000 to February 2007 have been presented in the present report in Appendices 1 to 7, while fish count data collected prior to December 2000 have been presented in previous reports (Davidson, 1999, 2001).

Divers observed a total of 15 species of reef fish from cobble, boulder and bedrock habitats in the Tonga Island Marine Reserve (Figure 7). Magpie moki (*Cheilodactylus nigripes*) was the only species recorded from control and reserve sites in 2007 but not in 2003. Spotty was the most abundant reef fish on all sample occasions from both reserve and control treatments (Appendices 1-7). Tarakihi abundance varied dramatically between years but was often the next most abundant species after spotty. Blue cod, banded wrasse, scarlet wrasse, sweep, blue moki and goatfish were regularly recorded from both reserve and control sites (Appendices 1-7). Leatherjacket, marblefish, red moki, magpie moki, butterfly and sea perch were recorded sporadically as individual adults. No butterfly were recorded or observed by divers during the study.



Blue cod (*Parapercis colias*) were recorded from pooled reserve and control groups on all sample occasions (Figure 8). Pooled reserve versus control mean density for blue cod (< 300 mm total length) was significantly different between the reserve and control treatments only in April-May 2006 primarily due to a drop in this size of cod at control sites during that year (Table 3). The mean density of small blue cod for the pooled control treatment generally remained low (i.e. below 0.3 individuals per 60 m²) over the 15 years except for a one sample peak in February 2005.

The density of small blue cod from the pooled reserve treatment increased over the duration of the study and was significantly higher in 2006 compared to at the start of the study (i.e. 1993) (T = -2.67, P < 0.008). Low numbers of large blue cod (≥300 mm length) were recorded on four sample occasions at control sites (December 2000, March 2001, February 2005, April-May 2006). No large blue cod were recorded from the reserve treatment in the first two sample years (Figure 8). The density of blue cod ≥300 mm TL has remained significantly higher at reserve sites compared to the control treatment since December 2000 (Table 3).

Table 3. T-test of density data for two size groups of blue cod for all sample years along the Abel Tasman coast.

| Reserve versus control | < 300 blue cod (TL) | | | | > 300 blue cod (TL) | | | |
|------------------------|---------------------|---------|-----|------|---------------------|---------|-----|------|
| | T value | P value | Df | Sig. | T value | P value | Df | Sig. |
| December 1993 | -0.7243 | 0.4698 | 169 | NS | +Inf | 0.0 | 169 | NS |
| December 1994 | 0.67107 | 0.50634 | 37 | NS | +Inf | 0.0 | 37 | NS |
| September 1999 | 0.6726 | 0.5022 | 155 | NS | 1.5173 | 0.1312 | 155 | NS |
| December 2000 | 0.3731 | 0.7095 | 154 | NS | 1.7276 | 0.0861 | 154 | Sig. |
| March 2001 | 1.0358 | 0.3019 | 154 | NS | 1.9722 | 0.0504 | 154 | Sig. |
| April 2002 | 1.1576 | 0.2488 | 154 | NS | 1.6587 | 0.09921 | 154 | Sig. |
| February 2004 | 1.5077 | 0.13336 | 154 | NS | 4.3700 | 0.00002 | 154 | Sig. |
| February 2005 | 0.8525 | 0.3952 | 154 | NS | 3.5593 | 0.00049 | 154 | Sig. |
| April-May 2006 | -2.3388 | 0.0206 | 154 | Sig. | -2.4779 | 0.01429 | 154 | Sig. |
| February 2007 | -1.1498 | 0.2519 | 154 | NS | -3.74409 | 0.00026 | 154 | Sig. |

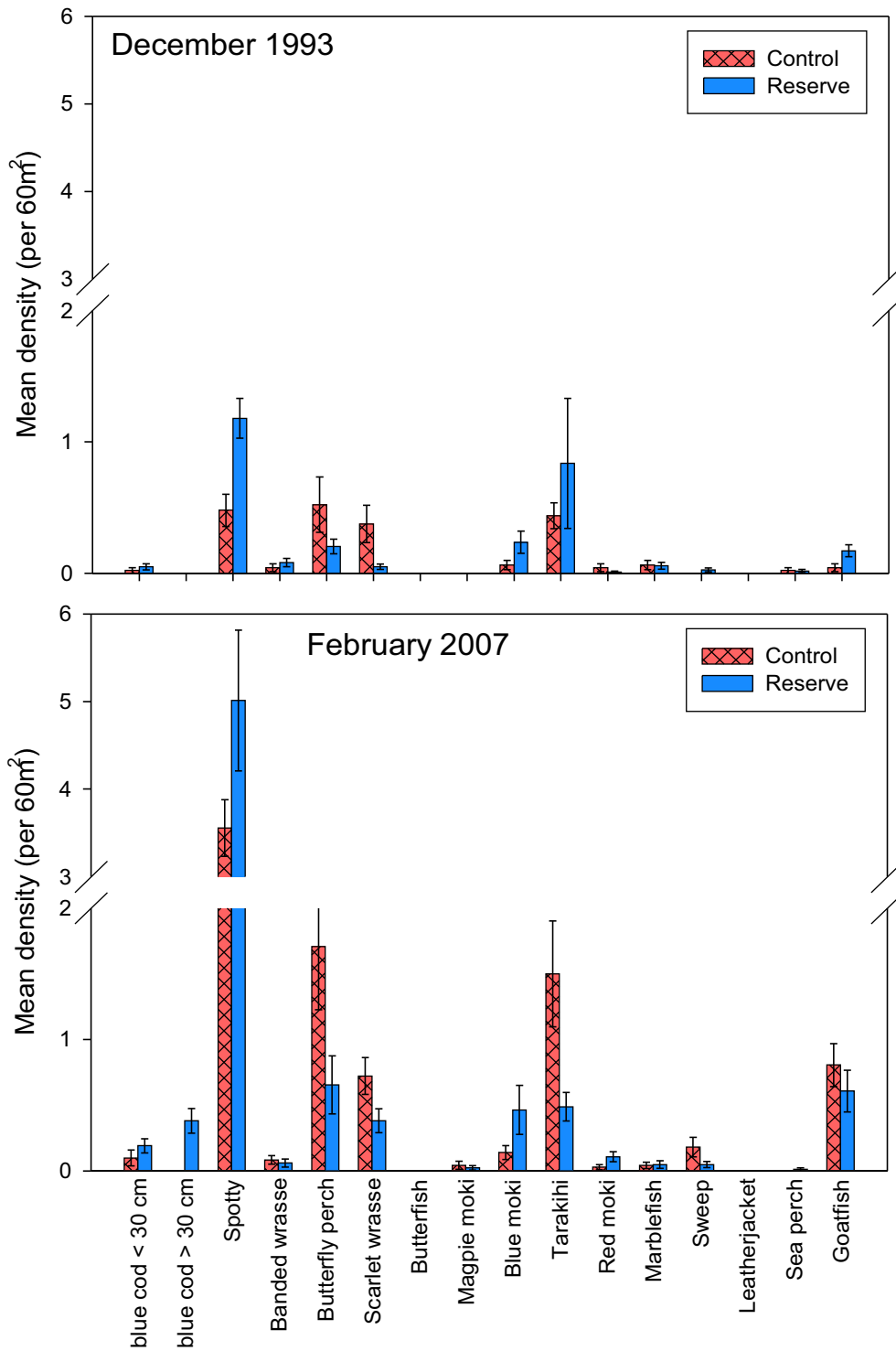


Figure 7. Mean densities of all reef fish sampled in December 1993 and February 2007 from pooled reserve (blue) and control sites (red hatched). Error bars are +/- 1 s.e.

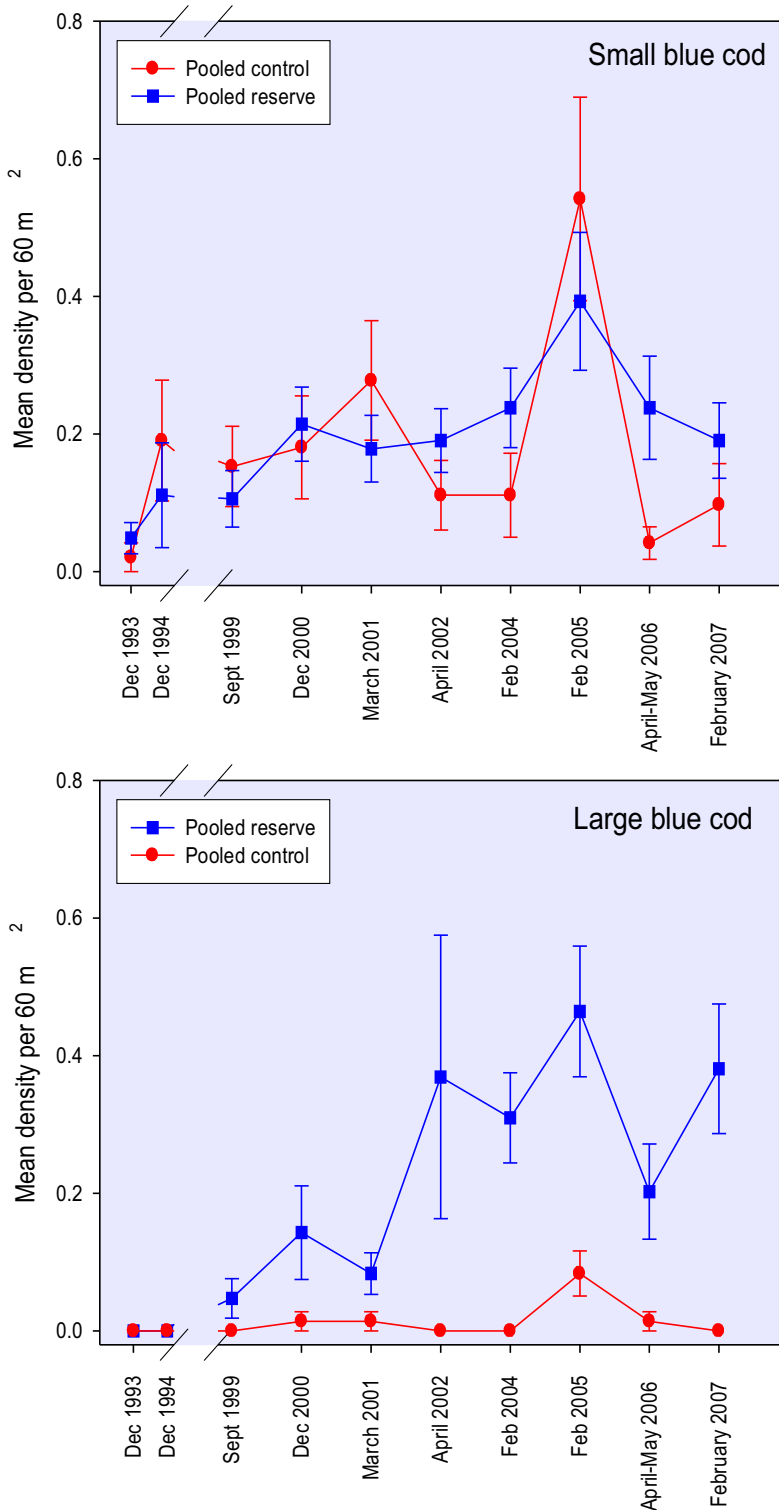


Figure 8. Pooled mean density of blue cod (< 300 mm (small) and ≥ 300 mm (large)) from all control and reserve sites sampled from 1993 to 2007. Error bars are +/- 1s.e. Large blue cod include 300 mm individuals. Note x-axis time sale is not regular.



Spotty (*Notolabrus celidotus*) were recorded from both treatments in all years (Figure 9). Little difference between the pooled reserve and control treatments was recorded in each sample year. Spotty density, however, varied between years generally increasing at both treatments over the duration of the study, peaking in April-May 2006. Significantly more spotty were recorded in April-May 2006 than December 1993 (reserve, $T = 7.024$, $P < 0.00001$; control, $T = 5.415$, $P < 0.00001$). The reason for this increase is unknown and unlikely to be related to the reserve establishment as it has occurred in both treatments.

Tarakihi (*Nemadactylus macropterus*) were recorded from both treatments in all years (Figure 9). Tarakihi abundance varied considerably between years, with peaks recorded in December 1994, December 2000, February 2005 and February 2007 (Figure 9). Very low densities were recorded for both treatments in September 1999, April 2002 and April-May 2006. Both pooled reserve and control treatments followed similar trends over the duration of the study with relatively small differences between the treatments. During the peaks, however, the pooled mean density of tarakihi was consistently higher at the control treatment compared to the reserve treatment. Tarakihi abundance showed no overall increase or decrease over the duration of the study, instead it was characterised by highs and lows. This pattern appeared independent of reservation and occurred in both reserve and control areas.

Apart from December 1994, the abundance of blue moki (*Latridopsis ciliaris*) has been higher from the pooled reserve group compared to the control treatment throughout the study. Blue moki abundance was significantly higher for the pooled reserve treatment from December 2000 onwards, apart from April 2002 and February 2005 (Figure 10, Table 4). The higher density of blue moki from the reserve treatment is probably due to the impact of reservation. The density of blue moki in the reserve treatment at the end of the study was higher than at the start of the study but this difference was not significant due the relatively large statistical variation around the mean ($T=-1.14$, $P=0.256$) (Figure 10).

Red moki (*Cheilodactylus spectabilis*) have not shown any increasing or decreasing trend at either reserve or control treatment (Figure 10). Their densities for both treatments were very low with occasional individuals being recorded at reserve and control sites. No change in their abundance can be attributed to reservation.

The abundance of all other reef fish species were very low and have not been plotted. Of note, however, was the occurrence of magpie moki (*Cheilodactylus nigripes*) in reserve transects in February 2005 and in both reserve and control transects in February 2007 (Figure 7, Appendix 5 and 7). This species seems to be more widespread than in the past and their increased incidence in transects suggests their abundance is also increasing.



Table 4. T-test of density data for blue moki for all sample years along the Abel Tasman coast.

| Reserve versus control | Blue moki (TL) | | | |
|------------------------|----------------|---------|-----|------|
| | T value | P value | Df | Sig. |
| December 1993 | 1.5136 | 0.1319 | 169 | NS |
| December 1994 | 1.2350 | 0.2248 | 36 | NS |
| September 1999 | 1.5181 | 0.1310 | 155 | NS |
| December 2000 | 2.8505 | 0.0049 | 154 | Sig. |
| March 2001 | 2.1170 | 0.0358 | 154 | Sig. |
| April 2002 | 1.1370 | 0.2573 | 154 | NS |
| February 2004 | 2.5776 | 0.0109 | 154 | Sig. |
| February 2005 | 1.0688 | 0.2868 | 154 | NS |
| April-May 2006 | 2.0820 | 0.0389 | 154 | Sig. |
| February 2007 | -1.567 | 0.1192 | 154 | NS |

Since December 2000, divers have estimated blue cod length during underwater visual fish counts (Figure 11). Blue cod mean size at pooled reserve sites has gradually climbed over the six year period from an average of 27.06 cm in December 2000 to 31.3 cm in February 2007. The mean size of blue cod from the pooled control treatment remained relatively consistent starting at 22.7 cm and ending at 22.8 cm but with a noticeable drop in February 2004. On all sample occasions, the mean size of blue cod was higher from the reserve treatment than the control treatment (Figure 11). Size-frequency histograms of blue cod from pooled reserve and control treatments show that large cod were very noticeable in the reserve from April 2002 onwards (Figure 12). These large individuals (>34 cm) were rare at the control sites, with only one > 40 cm cod recorded from a control site (February 2005). The existence of these large individuals in the population at reserve sites and not control sites is probably due to the exclusion of fishing from the marine reserve.

The average size of blue moki for the pooled reserve treatment fluctuated over the sample period, but ended significantly higher in 2007 compared to December 2000 ($P = 0.039$, $df = 96$) (Figure 11). At control sites, the average size of blue moki also increased, but always remained below the mean size recorded from the pooled reserve group. Size-frequency histograms of blue moki from pooled reserve and control treatments show that large moki were much more often encountered in the reserve compared to the control area (Figure 13) with few large individuals > 40 cm encountered from control sites. The existence of these large individuals in the population at reserve sites with less at control sites is probably due to the exclusion of fishing from the marine reserve.



Red moki mean size increased at pooled control and reserve treatments over the period that measurements were collected (2000 to 2007). Apart from March 2001 and February 2007, mean size at control sites was below the mean size recorded from the pooled reserve treatment (Figure 14). In December 2000 and February 2004, red moki size was significantly larger in pooled reserve sites than control sites. In March 2001, February 2005 and February 2007, there was little difference between the pooled reserve and control treatments. Overall there has been an increase in red moki size over the duration of the study but as this trend occurred at both reserve and control treatments there is no evidence to attribute these changes to marine reserve protection.

Most tarakihi at both reserve and control sites fell into two size classes (small 7-14 cm and medium 24-30 cm). Small tarakihi were observed in schools from 4 to 50 individuals at depths of 5-7 m for most years. The number of medium size class individuals, however, fluctuated between years with relatively few being observed by divers in some years. Although data were collected in 2000 to 2002, insufficient numbers were measured to include these data in the graph.

Both pooled reserve and control tarakihi mean lengths followed similar trends (Figure 14). Highest tarakihi densities for both reserve and control treatments were recorded from April-May 2006 and February 2007. A small increase in mean size of tarakihi occurred for both reserve and control treatments between February 2004 and 2007.

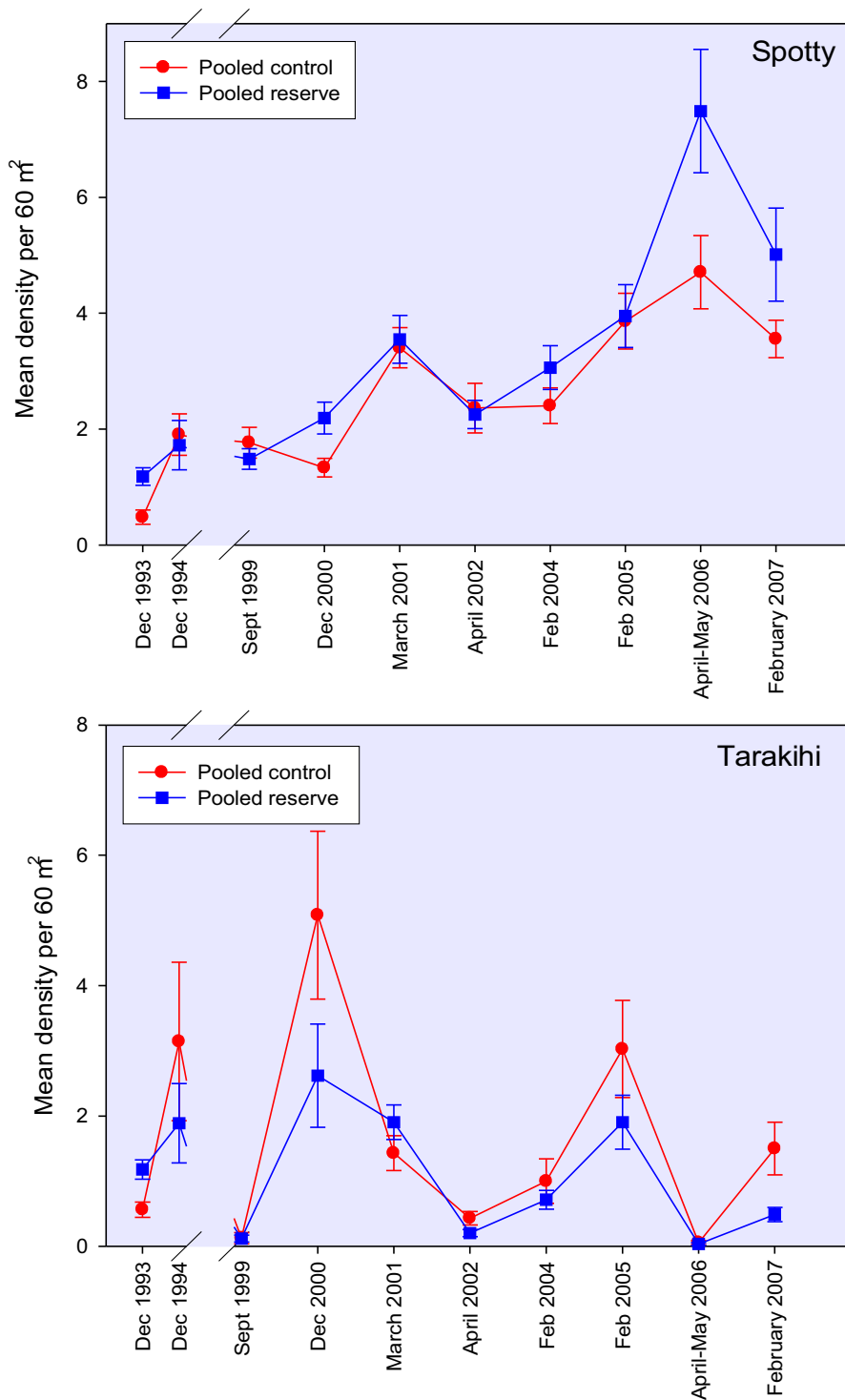


Figure 9. Pooled mean density of spotty and tarakihi from all control and reserve sites sampled from 1993 to 2007. Error bars are +/- 1s.e. Note x-axis time sale is not regular.

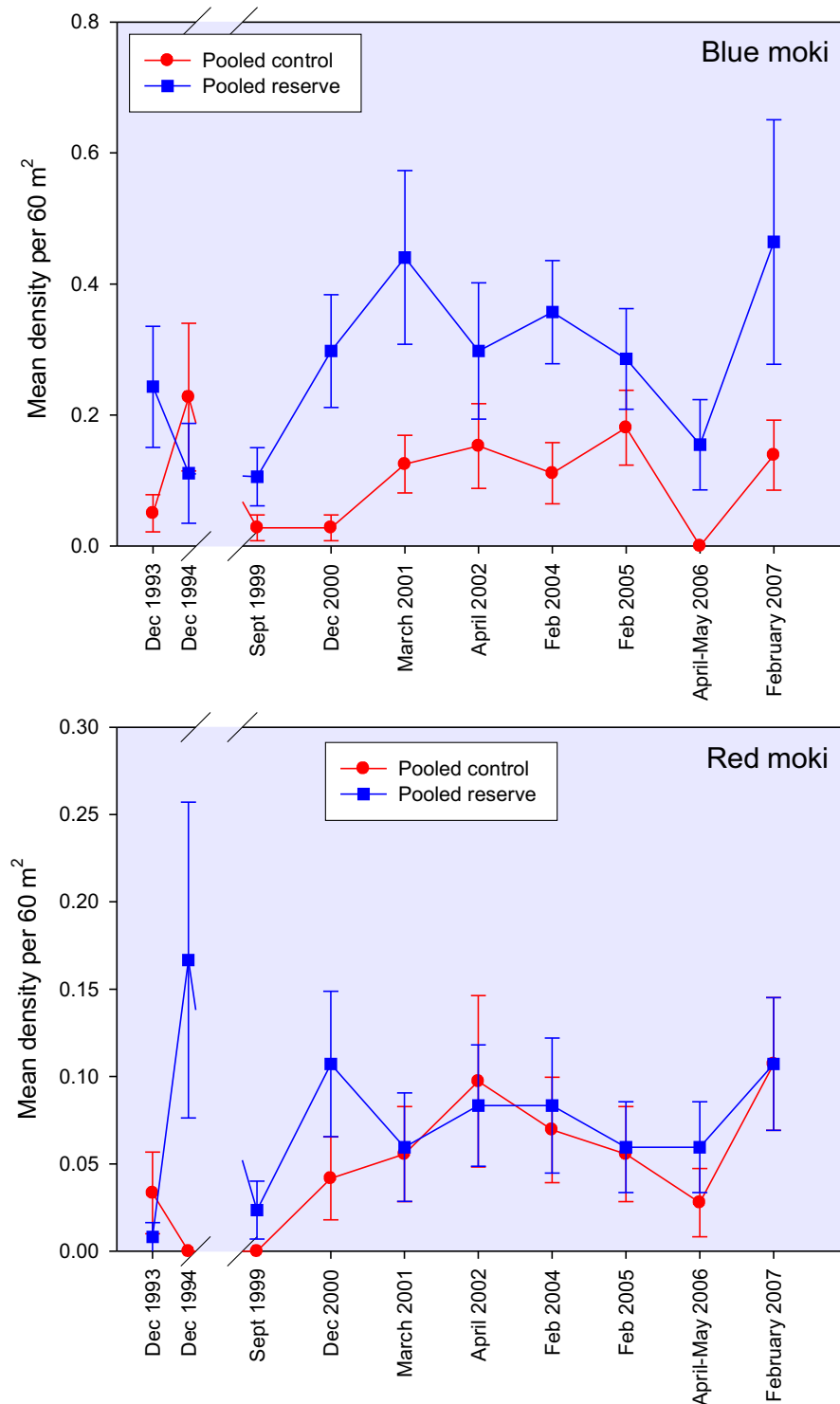


Figure 10. Pooled mean density of blue moki and red moki from all control and reserve sites sampled from 1993 to 2007. Error bars are +/- 1s.e. Note x-axis time sale is not regular.

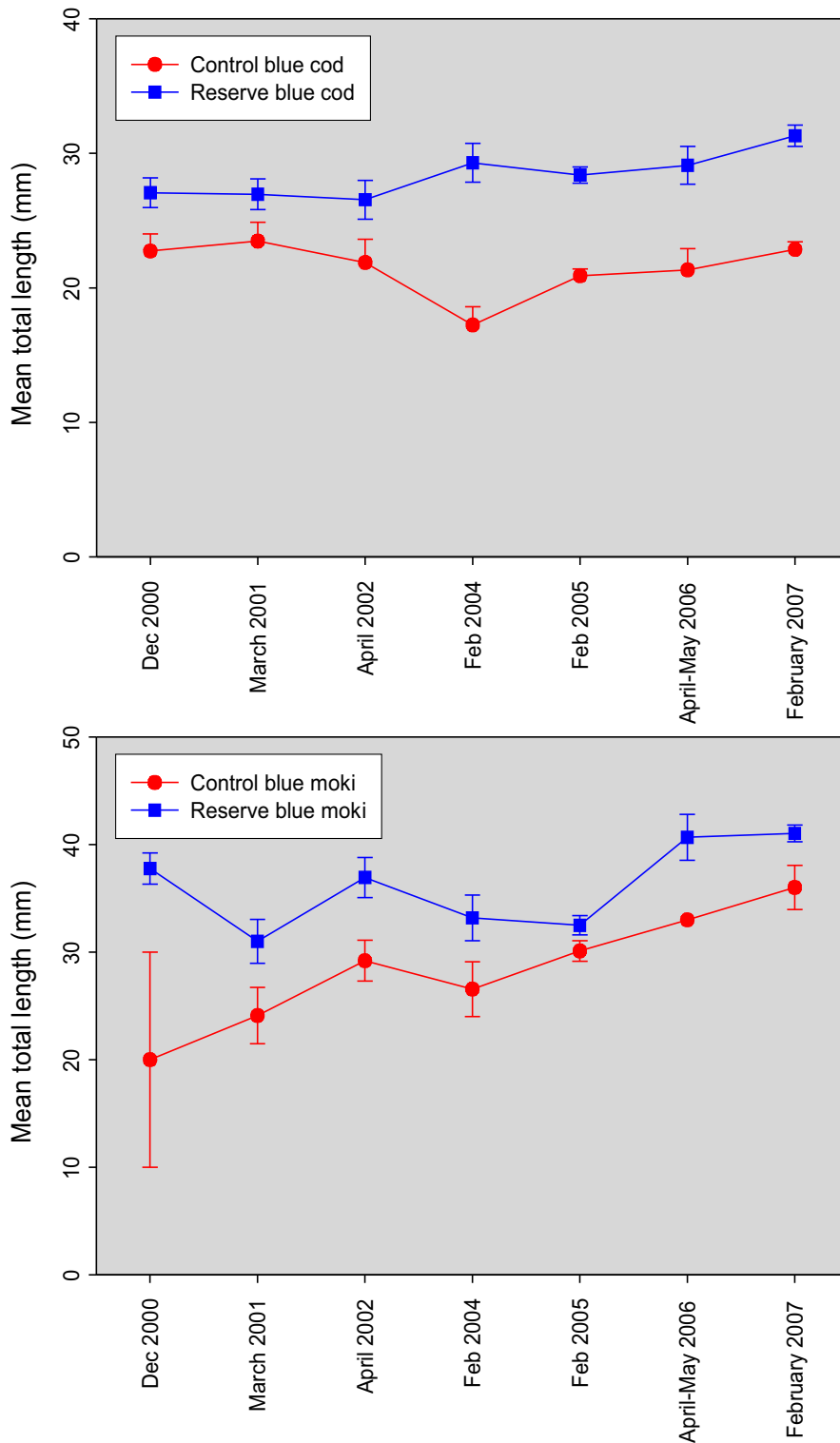


Figure 11. Pooled mean size of blue cod and blue moki from all control and reserve sites sampled from 2000 to 2007. Error bars are +/- 1s.e.

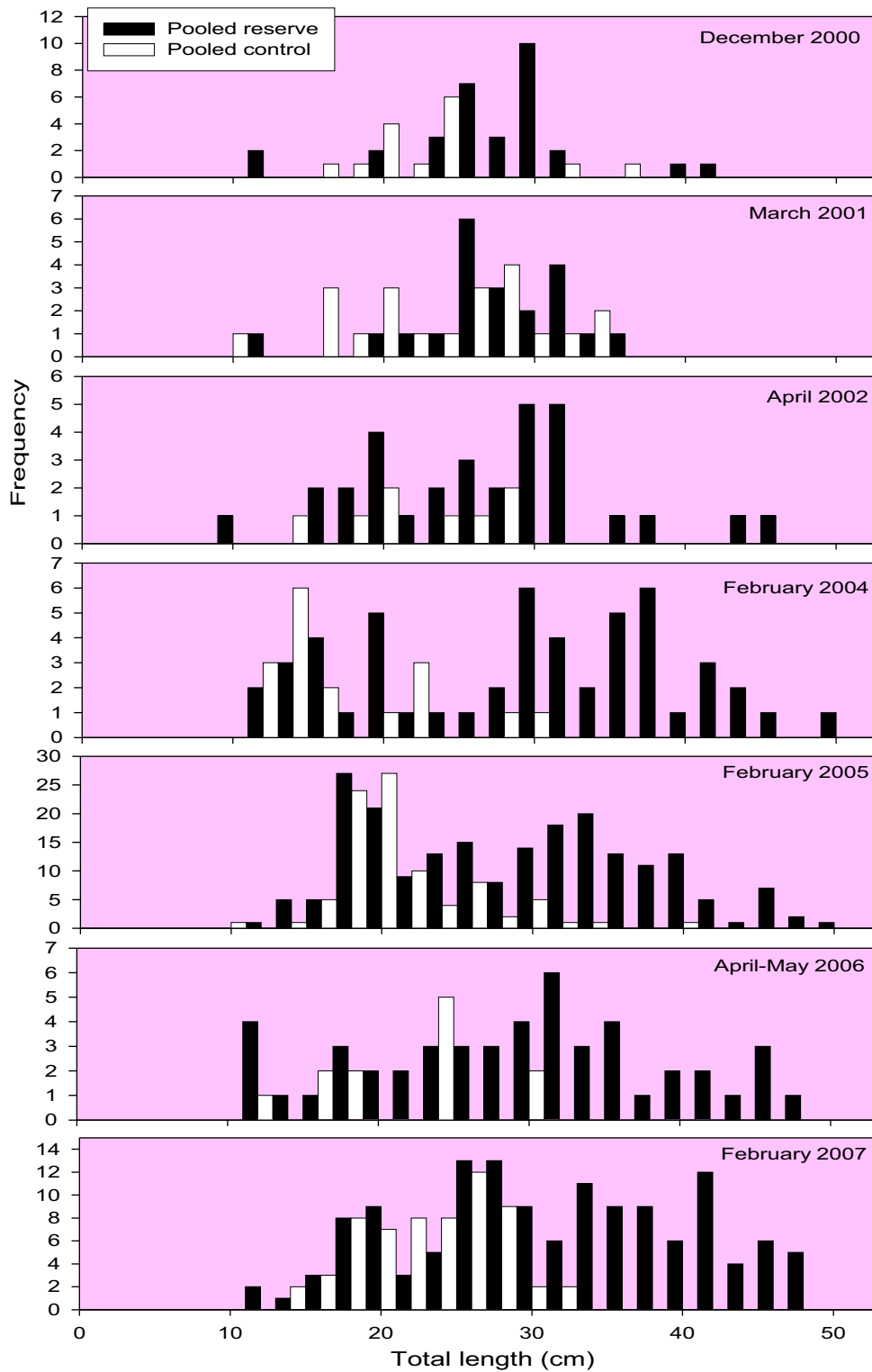


Figure 12. Pooled size frequency of blue cod from pooled control and reserve sites sampled from 2000 to 2007.

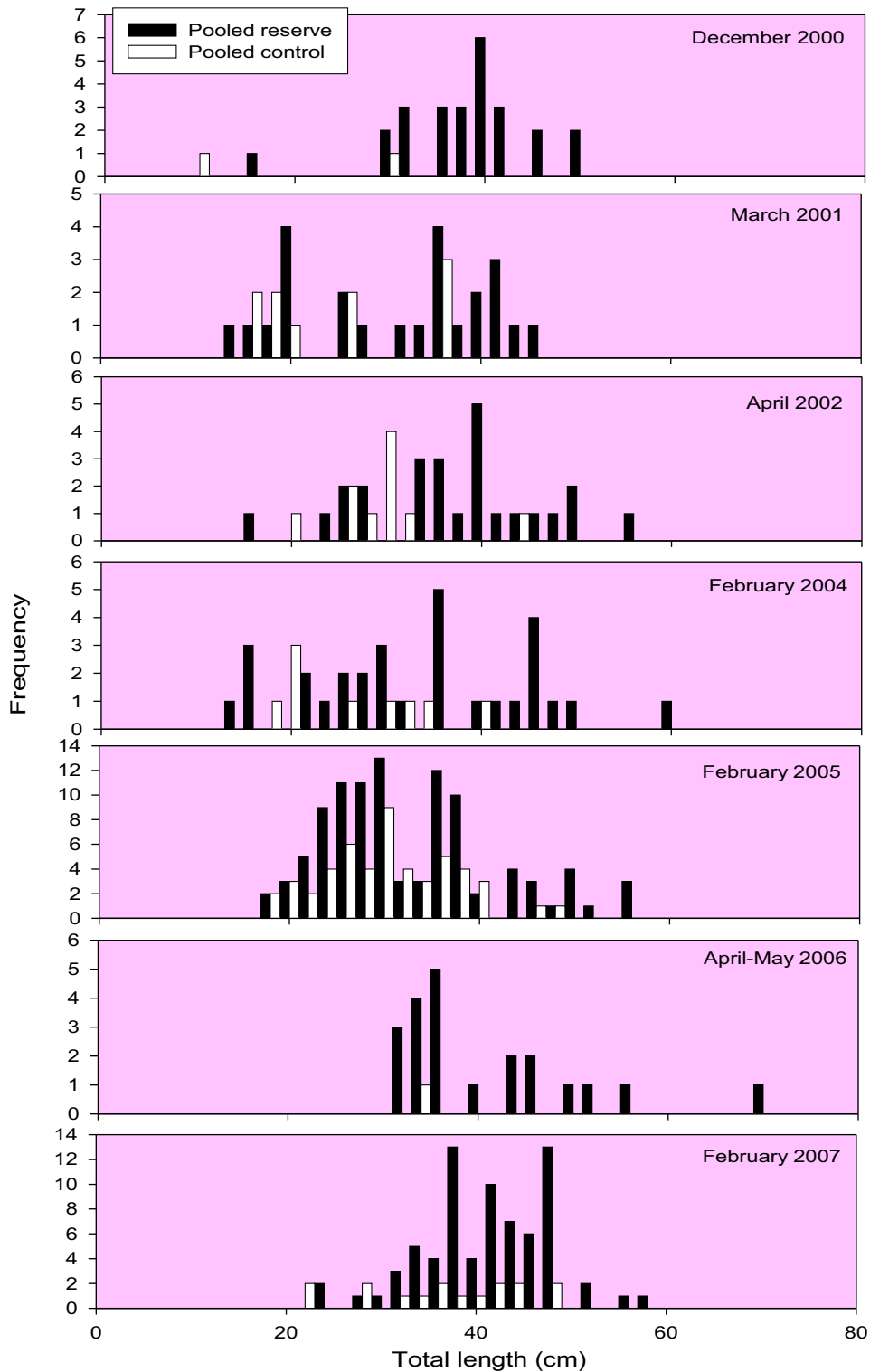


Figure 13. Pooled size frequency of blue moki from pooled control and reserve sites sampled from 2000 to 2007.

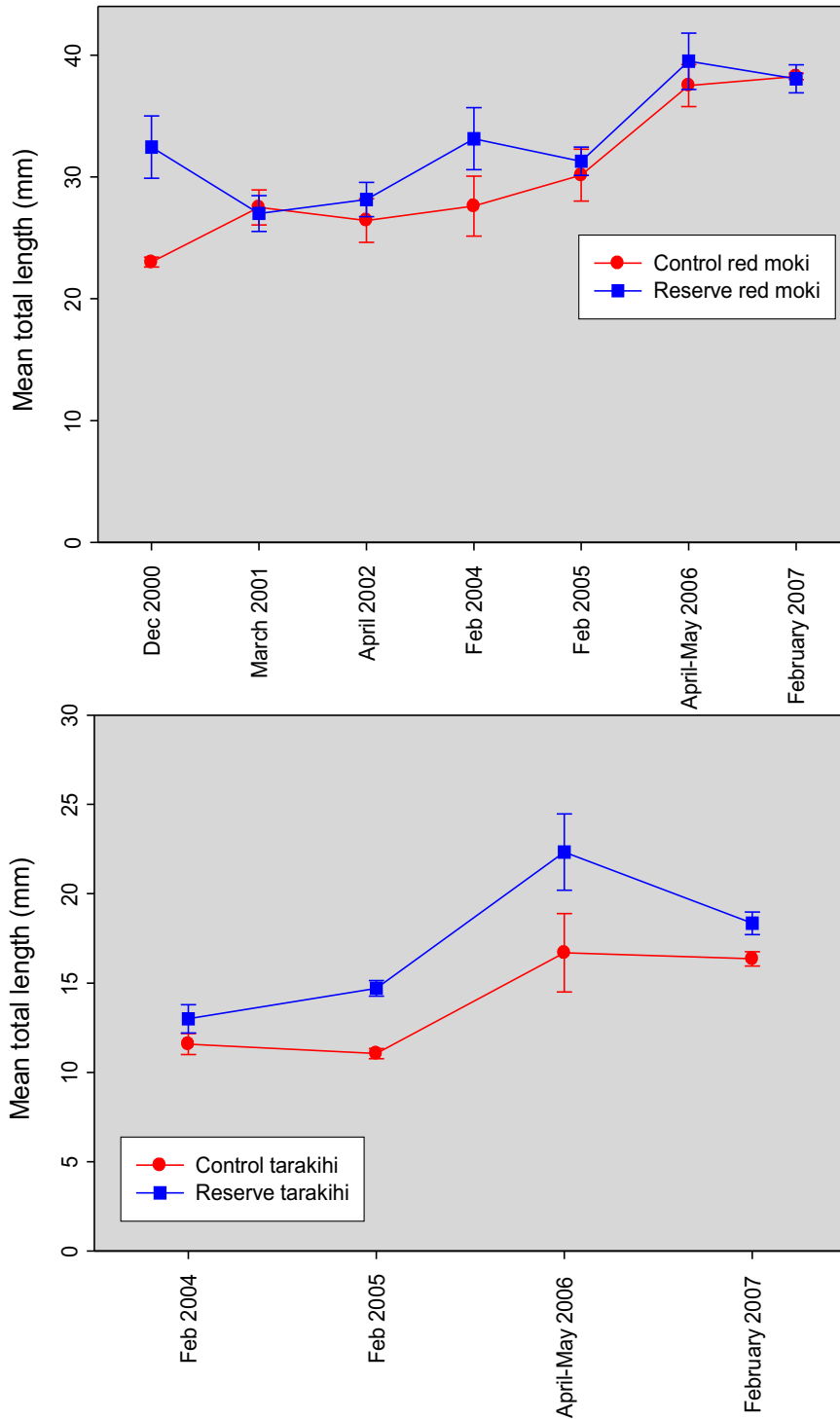


Figure 14. Pooled mean size of red moki and tarakihi from all control and reserve sites sampled from 2000 to 2007 for red moki and 2004 to 2007 for tarakihi. Error bars are +/- 1s.e.



5.2 Spiny lobster density, sex, and size

Lobsters were recorded from boulder or bedrock habitats at reserve and control sites. Rocky habitat was most often located in water < 12 m depth along the length of the Abel Tasman coastline. In some areas, however, rocky substratum ended by < 3 m depth. These very shallow areas were not sampled in the present study due to a lack of suitable lobster habitat.

In shallow quadrats (4-7m), lobster density was consistently higher in the reserve compared to the pooled control sites (Figure 15). Lobster density from the shallow control treatment gradually increased from December 1998, peaking in February 2007. Lobster density from the shallow reserve treatment also increased throughout the study peaking in December 2002 and again in April 2006 (Figure 15).

Similarly, the density of lobsters surveyed from deeper areas (9-11 m) increased in both pooled reserve and control samples from 1994 to 2002. In contrast to the control pooled average densities that declined from December 2002, the average lobster density from the pooled reserve treatment continued to increase to an all time high in February 2007 (Figure 15).

For most lobster size classes a comparison cannot be made between data collected before and after 1998 as the measurement methodology changed from total body length to carapace length. The exception was for juvenile lobsters where the cut-off size for lobster length corresponded to the 85 cm carapace length for juveniles (Figure 16 and Table 5).

The numbers of juvenile lobsters (< 85 cm CL) were comparable for most years between control and reserve treatments (Figure 16). In December 2000 the density of juvenile lobsters peaked in both reserve and control treatments. Juvenile lobsters recorded from this peak grew through to the larger size classes resulting in an increase in mature females and non reproductive males in the reserve from 2004 to 2007 (Figure 16). This increase was also recorded for non reproductive males outside the reserve but was not so pronounced for mature females.

More large reproductive males were recorded in most years from the reserve treatment compared to the control group where they remained relatively uncommon (Figure 16). A dip in the numbers of these large males was recorded from the reserve treatment in November 1999 and December 2002. The reason for this dip is unknown but may be related to movements of large lobsters onshore and offshore. This phenomenon has been recorded for lobsters in other marine reserves in New Zealand (Kelly 2001, Freeman in prep.). Since this dip in numbers, large male lobsters have increased to an all time high in February 2007.



Lobsters > 110 mm CL for 2002, 2004, 2006 and 2007 represented a greater proportion of the population within the reserve compared to the control treatment (Figures 17, 18, 19, 20 and 21). In February 2007, the mean size of lobsters in the pooled reserve group was 18.4 mm larger than the control treatment mean. In 2007, no lobsters > 160 mm CL were recorded from the control treatment. In comparison, lobsters up to 200 mm CL were regularly recorded from the reserve treatment (Figure 21).

Table 5. Sex composition of spiny lobsters sampled in Tonga Island Marine Reserve and control sites (1994 to 2007). Data from quadrats and from additional lobsters observed outside quadrats.

| Date | Reserve sites | | | Control sites | | |
|---------------|---------------|-------|---------|---------------|-------|---------|
| | Juveniles | Males | Females | Juveniles | Males | Females |
| | N | N | N | N | N | N |
| December 1994 | 26 | NA | NA | 13 | NA | NA |
| December 1998 | 14 | 22 | 6 | 0 | 2 | 4 |
| February 1999 | 17 | 48 | 26 | 12 | 7 | 4 |
| May 1999 | 2 | 39 | 19 | 9 | 2 | 0 |
| November 1999 | 23 | 36 | 11 | 15 | 4 | 7 |
| March 2000 | 24 | 47 | 40 | 21 | 6 | 3 |
| December 2000 | 57 | 37 | 23 | 22 | 11 | 5 |
| December 2002 | 83 | 33 | 38 | 51 | 18 | 10 |
| February 2004 | 29 | 88 | 36 | 40 | 26 | 13 |
| April 2006 | 17 | 137 | 115 | 28 | 33 | 24 |
| February 2007 | 41 | 120 | 102 | 20 | 11 | 31 |

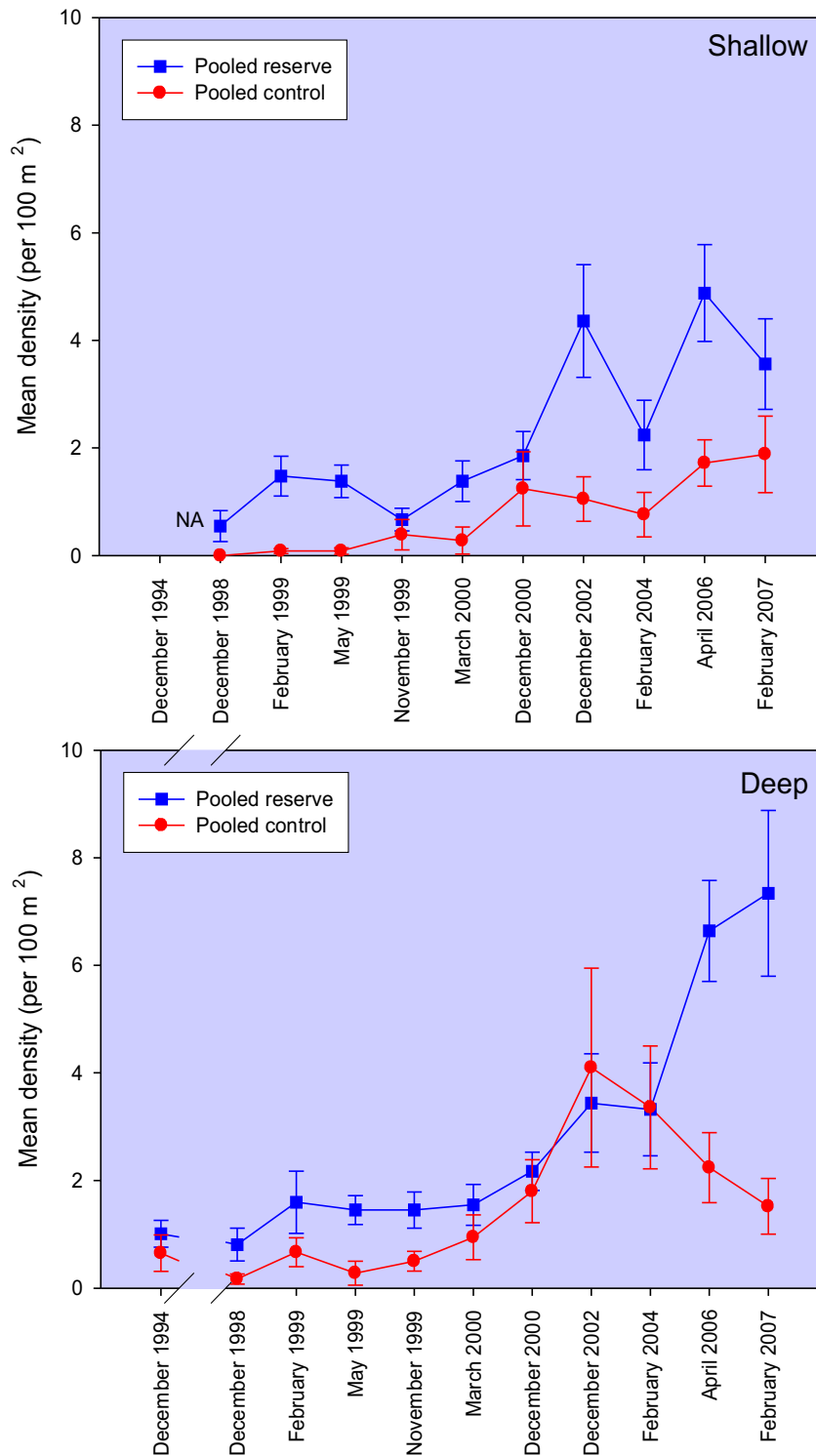


Figure 15. Pooled mean density of lobsters from all control and reserve sites sampled from 1994 to 2007 from deep (10-11m) and shallow (6-7 m) strata. Error bars are +/- 1s.e.

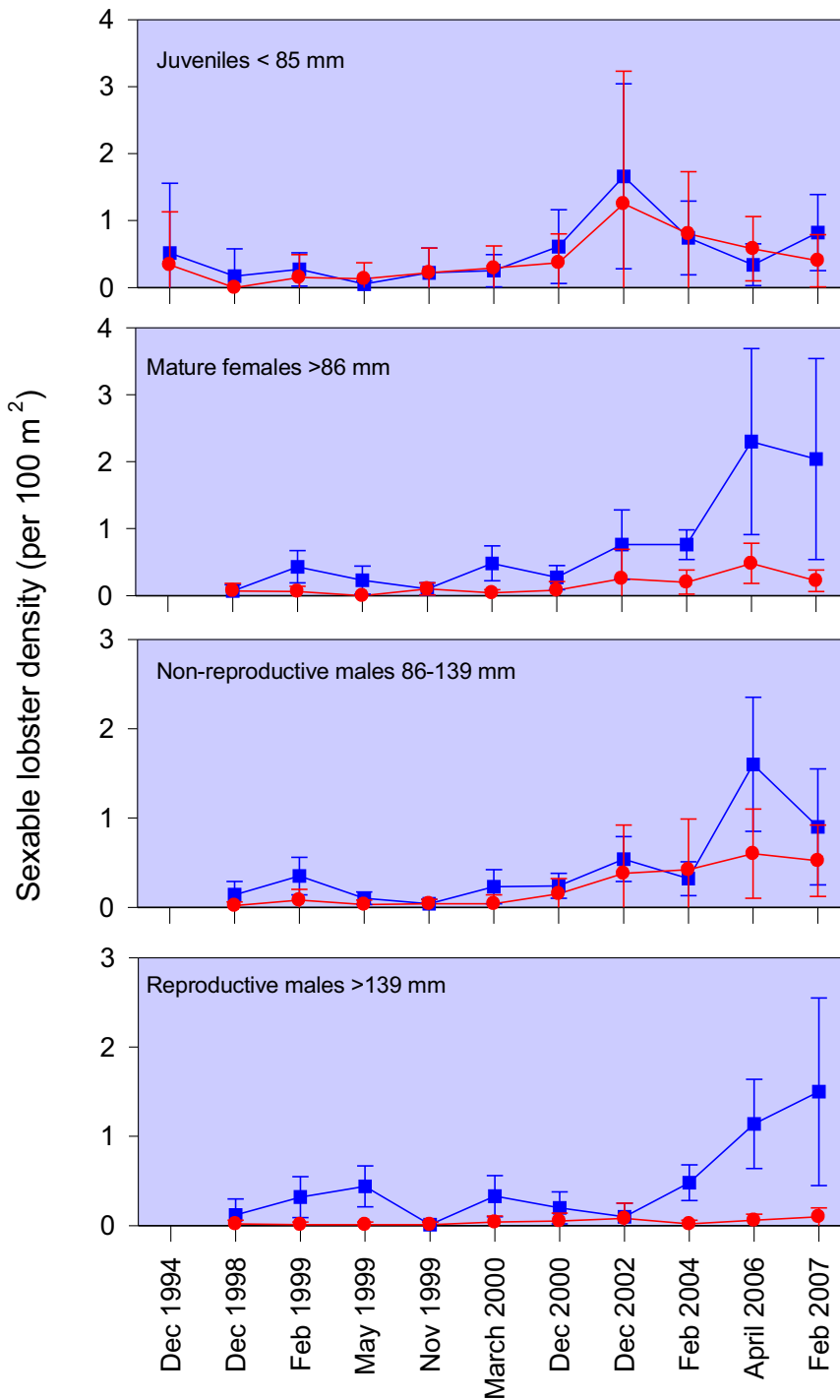


Figure 16. Density of spiny lobster individuals that could be sexed from all control (red circles) and reserve (blue squares) sites sampled from 1994 to 2007. Lobsters have been divided into reproductive classes according to MacDiarmid (1989). Sizes are estimated carapace length (mm).

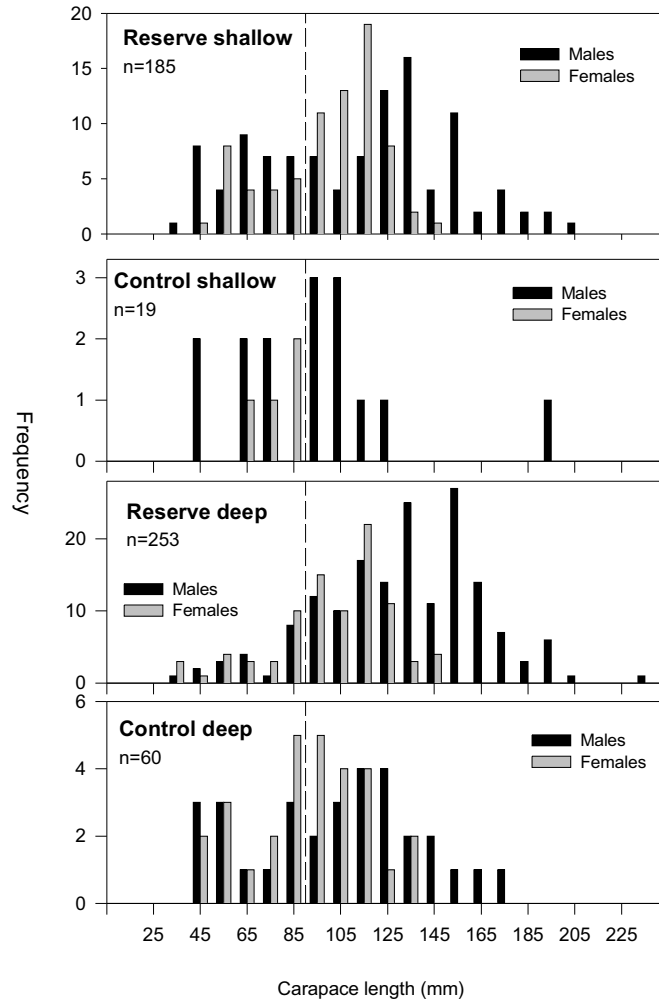


Figure 17. (from Davidson *et al.* 2002) Size-frequency distributions of spiny lobster for shallow and deep quadrats pooled across reserve and control treatments from December 1998 to December 2000. Sizes are estimated carapace length (mm).

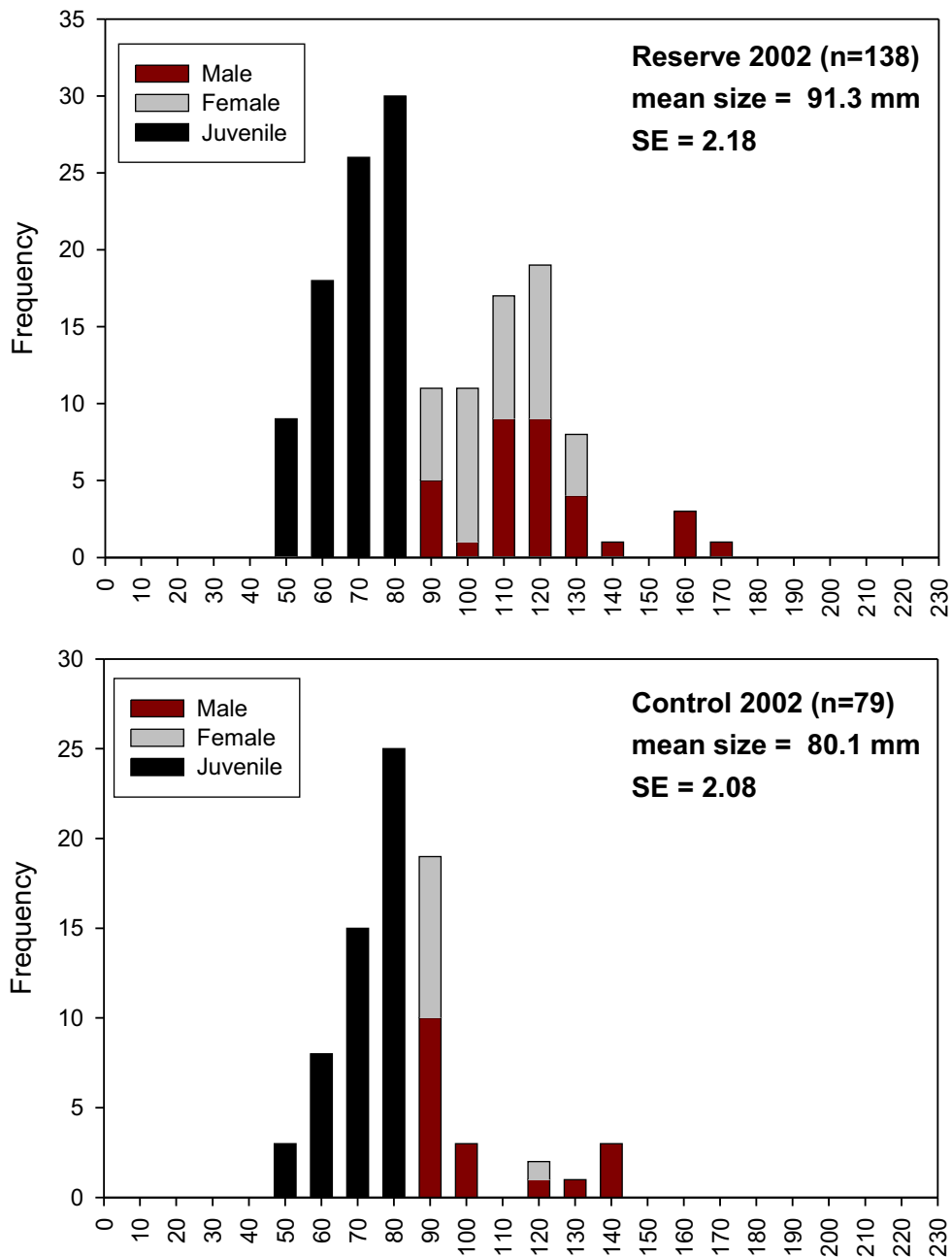


Figure 18. Size-frequency distributions of spiny lobster pooled across reserve and control treatments for December 2002. Sizes are estimated carapace length (mm).

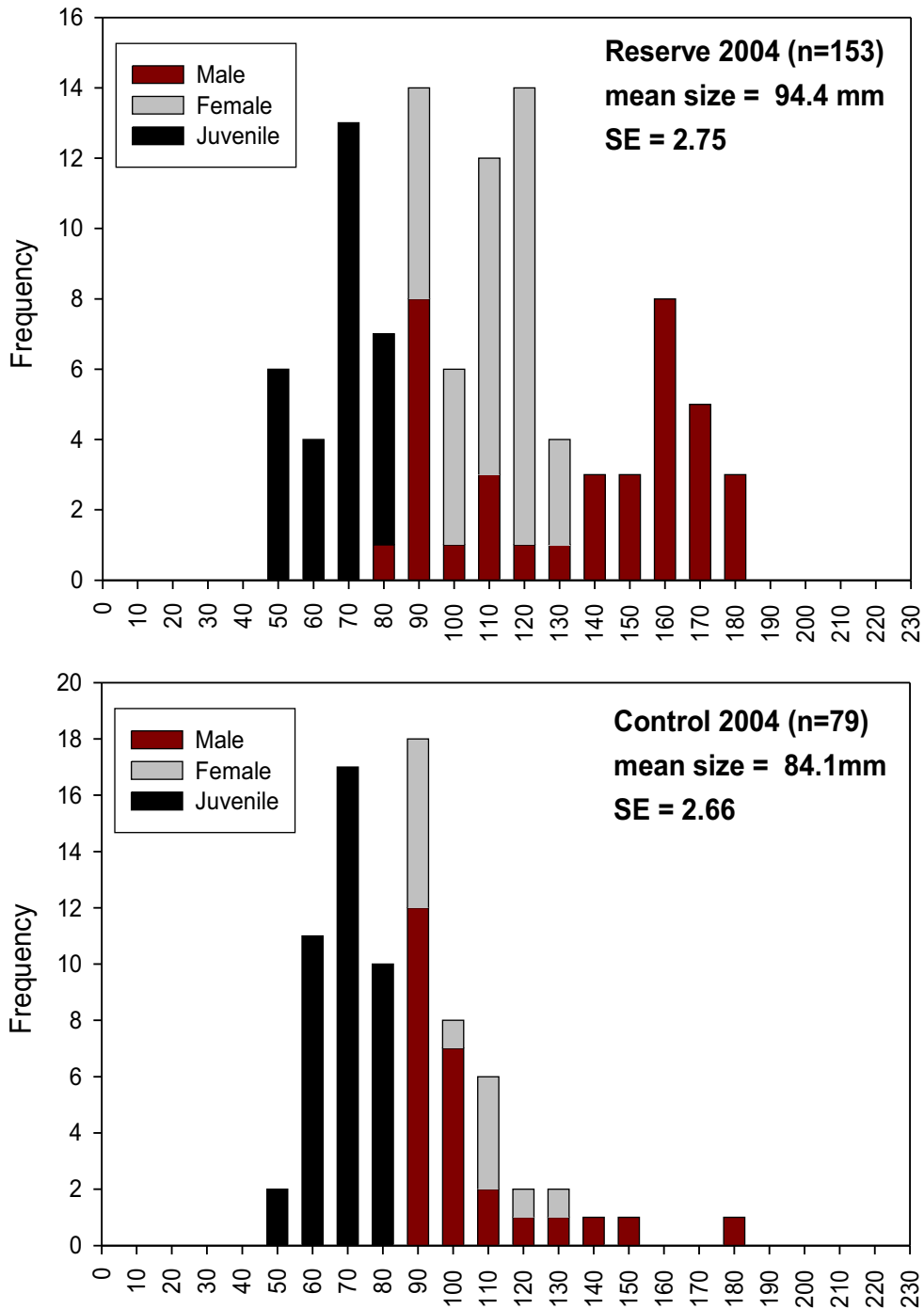


Figure 19. Size-frequency distributions of spiny lobster pooled across reserve and control treatments for February 2004. Sizes are estimated carapace length (mm).

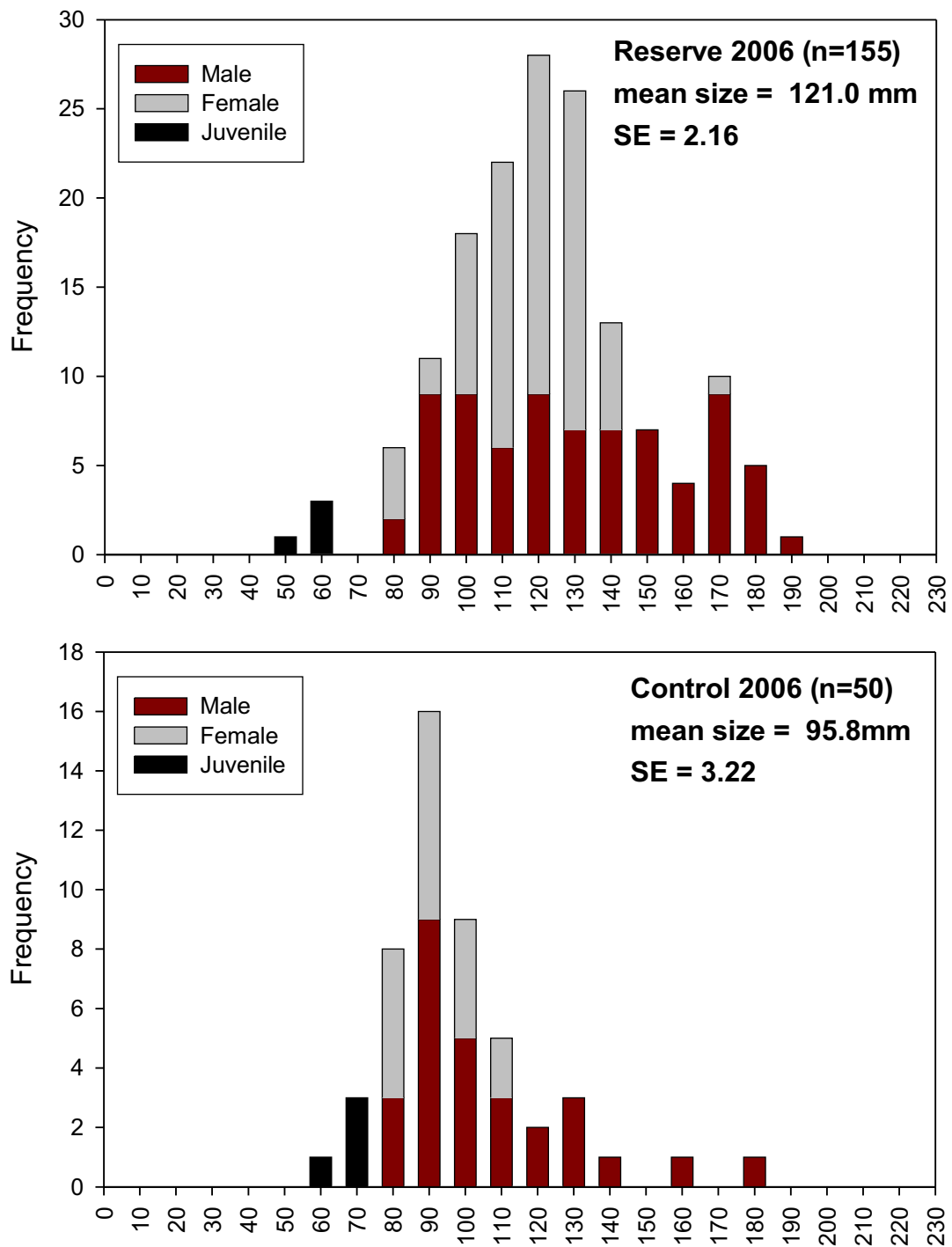


Figure 20. Size-frequency distributions of spiny lobster pooled across reserve and control treatments for April 2006. Sizes are estimated carapace length (mm).

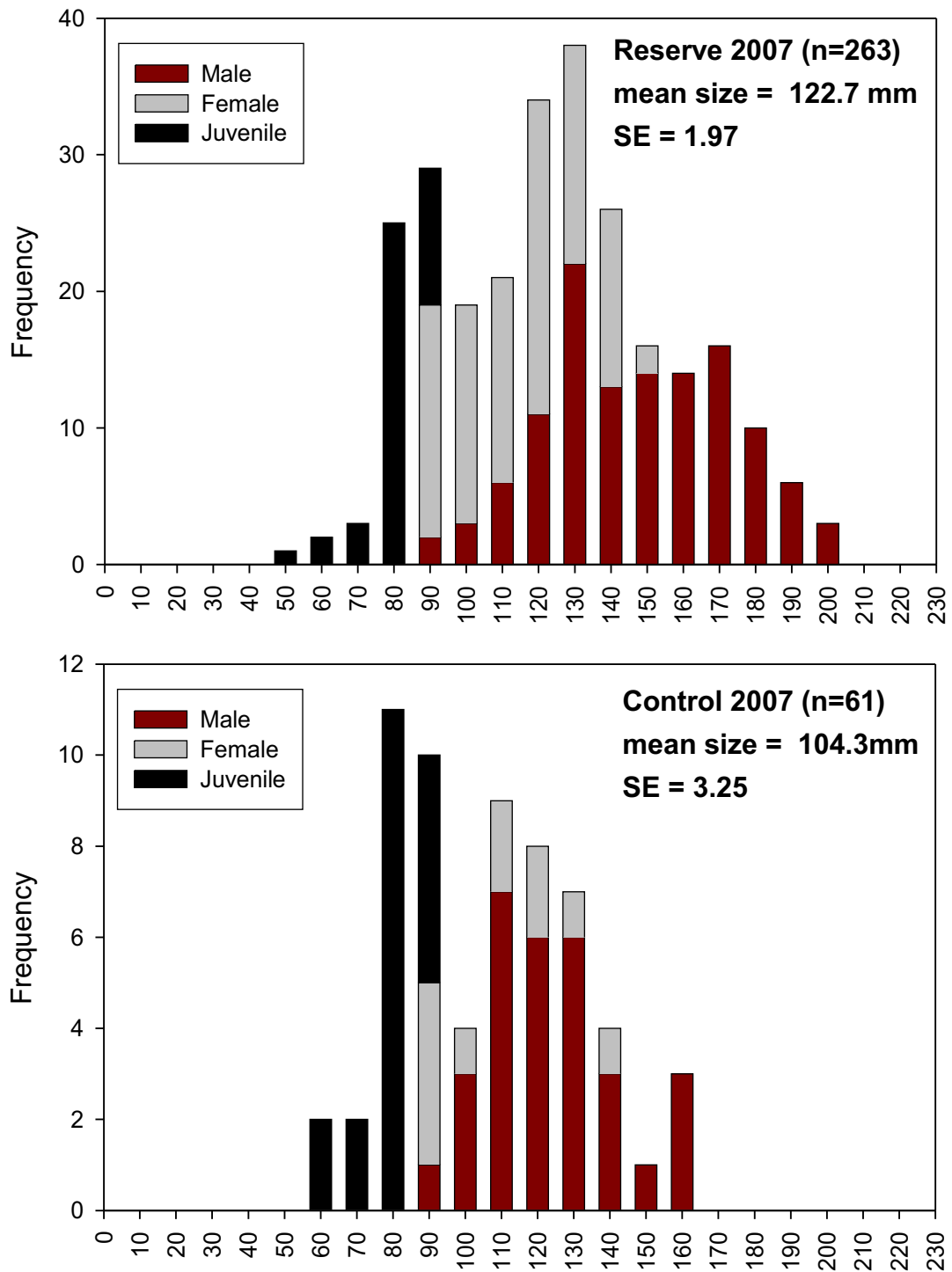


Figure 21. Size-frequency distributions of spiny lobster pooled across reserve and control treatments for February 2007. Sizes are estimated carapace length (mm).



5.3 Shore profiles

Shore profile data was collected from nine sites in March 2001 (Figure 3, Appendix 14). No obvious changes in habitat composition or macroalgae percentage cover were observed between these samples and profiles collected in December 1993.

5.4 Benthic quadrats

Benthic quadrat data were collected in 1993-94 (Davidson, 1999) and in March 2001 (Appendix 15, Figures 22 and 23). In March 2001, a total of nine shallow rocky sites (five reserve and four control) were sampled. Few large scale changes in species abundance were observed between the two sample occasions. The most notable was the anemone, *Culicea rubeola*. In 1993, this species was more abundant at both reserve and control sites than in 2001 (Table 6, Figures 22 and 23). The tubeworm (*Galeolaria hystrix*) also exhibited a decrease similar to *C. rubeola*. Of note, was an increase in the percentage cover and number of stipes of the alga *Carpophyllum flexuosum* in the reserve treatment between 1993 and 2001. Most abundant species at both reserve and control sites were coralline algae, box anemones, limpets, window oyster, top shells, tubeworms, kina, cushion stars and Cook's turban shells.

5.5 Kina density and size

Raw kina density data collected in April 2002 have been presented in Appendix 16, while size data have been presented in Appendix 21. Previous kina density data collected in December 1993 was presented in Davidson (1999).

Kina density declined at the pooled reserve treatment and increased at the control treatment between the two sample occasions (Figure 24). Of the five key rock dwelling invertebrates sampled, this is the only species where the reserve density did not follow the same trend as the control treatment (Figures 24 and 25). The decline in kina density at the reserve site was not large (i.e. 1.57 to 1.36 individuals per m²) with error bars from each treatment overlapping.

Mean kina size was higher in the reserve compared to the control treatment. Between 1993 and 2002 mean kina size increased slightly in the reserve but declined by a similar margin outside the reserve (Figure 26, Table 7). These differences were, however, relatively small with error bars overlapping between the treatments.

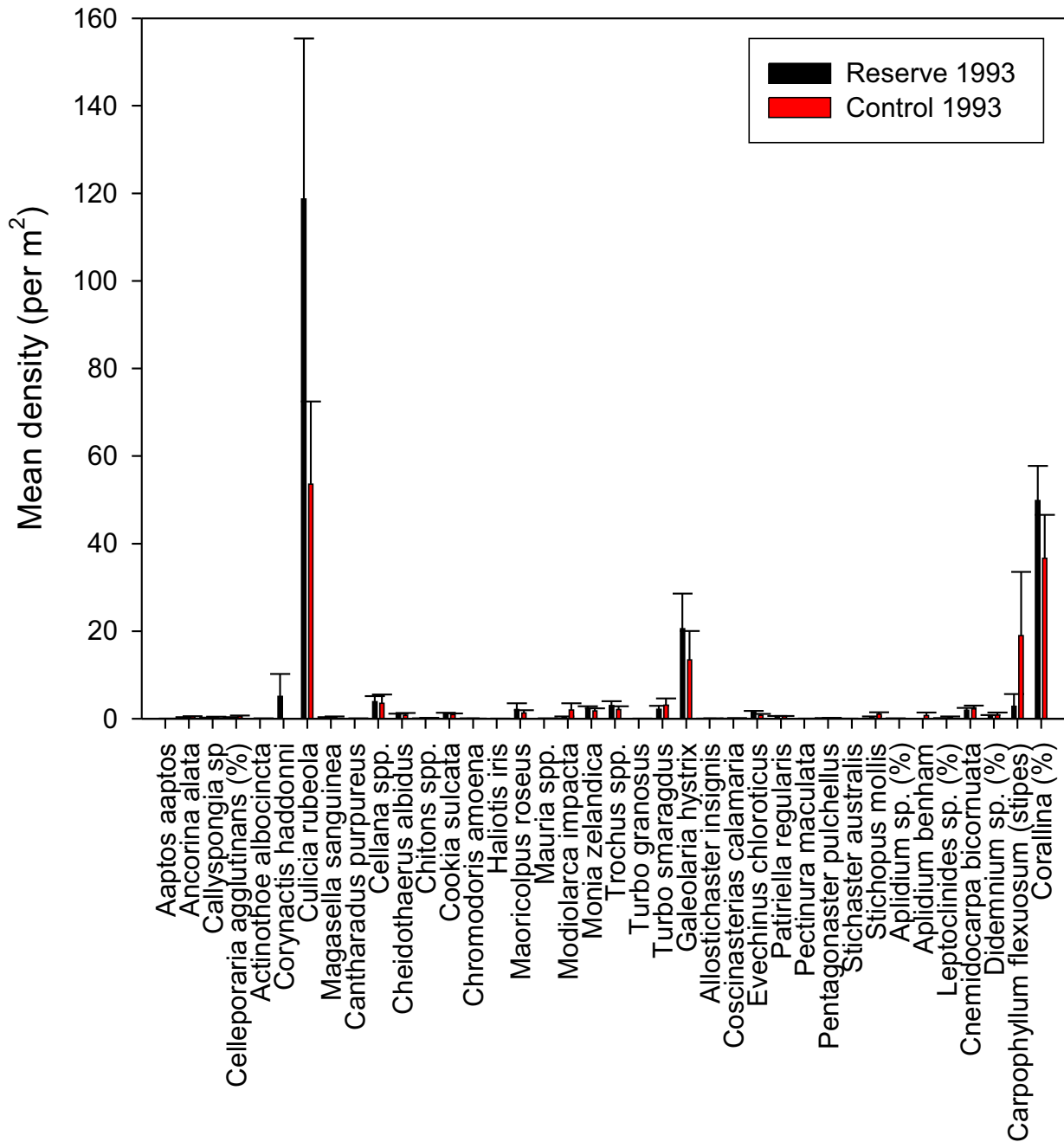


Figure 22. Mean density of benthic invertebrates and algae recorded from pooled reserve and control sites in 1993. Error +/- 1 s.e.

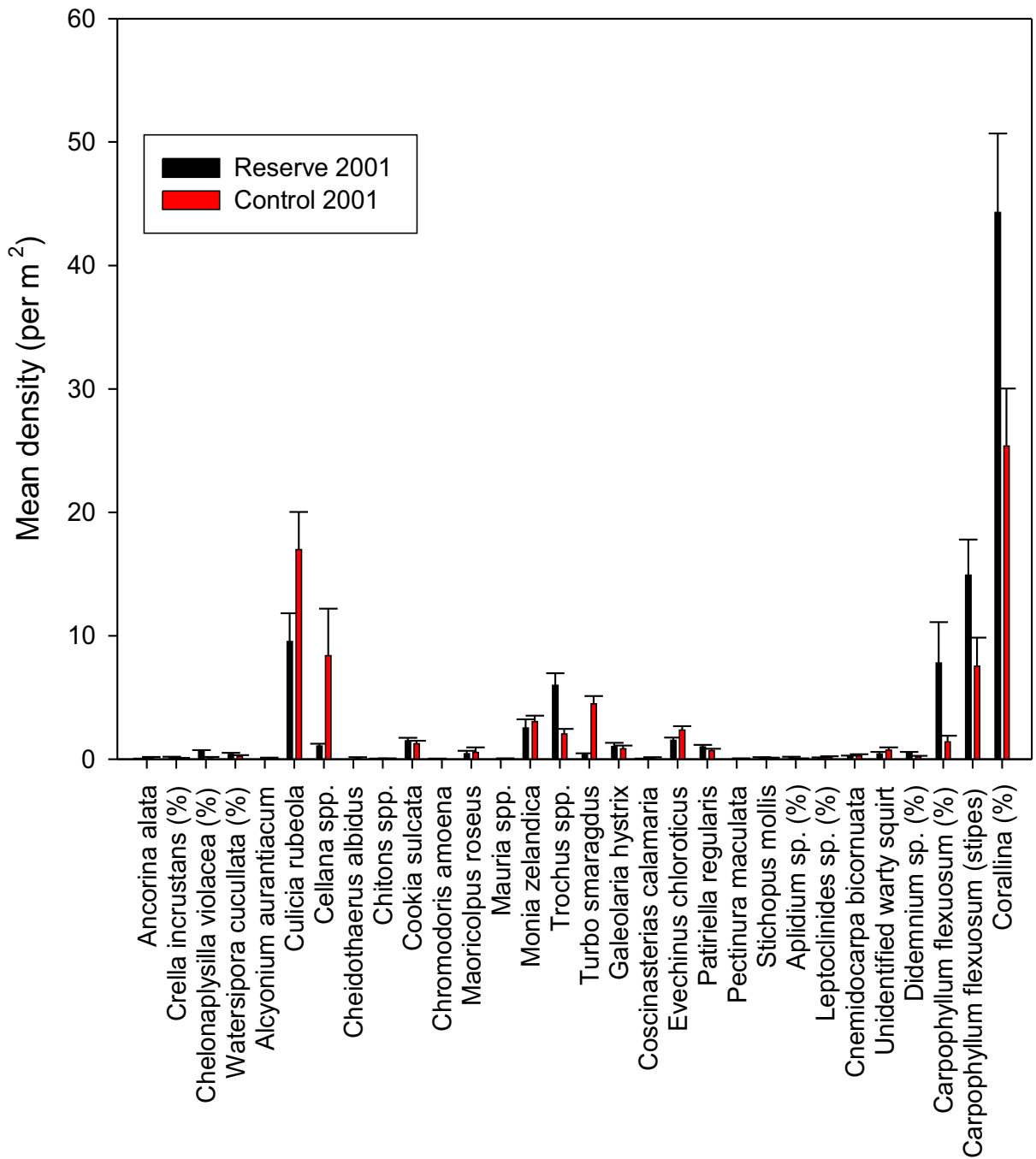


Figure 23. Mean density of benthic invertebrates and algae recorded from pooled reserve and control sites in 2001. Error +/- 1 s.e.



Table 6. Mean densities (per m²) for selected benthic species sampled from pooled rocky reserve and control treatments in 1993 and 2001.

| Date Treatment | 1993 | 1993 | 1993 | 1993 | 1993 | 1993 | 2001 | 2001 | 2001 | 2001 | 2001 | 2001 |
|--|--------------|------------|------------|--------------|------------|------------|--------------|------------|------------|--------------|------------|------------|
| | Reserve Mean | Reserve SD | Reserve SE | Control Mean | Control SD | Control SE | Reserve Mean | Reserve SD | Reserve SE | Control Mean | Control SD | Control SE |
| <i>Aaptos aaptos</i> | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Ancorina alata</i> | 0.22 | 0.39 | 0.11 | 0.42 | 0.43 | 0.14 | 0.02 | 0.15 | 0.02 | 0.12 | 0.43 | 0.08 |
| <i>Crella incrustans</i> (%) | NA | NA | NA | NA | NA | NA | 0.13 | 0.47 | 0.07 | 0.07 | 0.25 | 0.04 |
| <i>Callyspongia</i> sp | 0.14 | 0.36 | 0.10 | 0.23 | 0.52 | 0.17 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Chelonaplysilla violacea</i> (%) | NA | NA | NA | NA | NA | NA | 0.63 | 0.69 | 0.11 | 0.13 | 0.35 | 0.06 |
| <i>Celleporaria agglutinans</i> (%) | 0.26 | 0.32 | 0.09 | 0.49 | 0.58 | 0.19 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Watersipora cucullata</i> (%) | NA | NA | NA | NA | NA | NA | 0.38 | 0.91 | 0.14 | 0.24 | 0.58 | 0.10 |
| <i>Actinothoe albocincta</i> | 0.00 | 0.00 | 0.00 | 0.04 | 0.07 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Alcyonium aurantiacum</i> | NA | NA | NA | NA | NA | NA | 0.00 | 0.00 | 0.00 | 0.06 | 0.35 | 0.06 |
| <i>Corynactis haddonii</i> | 5.10 | 19.08 | 5.10 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Culicia rubeola</i> | 118.74 | 137.15 | 36.65 | 53.57 | 56.69 | 18.90 | 9.55 | 14.78 | 2.28 | 17.00 | 17.12 | 3.03 |
| <i>Magasella sanguinea</i> | 0.21 | 0.47 | 0.12 | 0.26 | 0.60 | 0.20 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Cantharadus purpureus</i> | 0.00 | 0.00 | 0.00 | 0.04 | 0.11 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Cellana</i> spp. | 3.86 | 4.84 | 1.29 | 3.50 | 6.11 | 2.04 | 1.06 | 1.35 | 0.21 | 8.39 | 21.53 | 3.81 |
| <i>Cheidothaerus albidus</i> | 0.83 | 1.20 | 0.32 | 0.75 | 1.57 | 0.52 | 0.00 | 0.00 | 0.00 | 0.10 | 0.40 | 0.07 |
| <i>Chitons</i> spp. | 0.04 | 0.09 | 0.02 | 0.10 | 0.17 | 0.06 | 0.02 | 0.15 | 0.02 | 0.03 | 0.18 | 0.03 |
| <i>Cookia sulcata</i> | 1.07 | 0.91 | 0.24 | 0.91 | 0.92 | 0.31 | 1.50 | 1.50 | 0.23 | 1.27 | 1.31 | 0.23 |
| <i>Chromodoris amoena</i> | 0.05 | 0.10 | 0.03 | 0.02 | 0.05 | 0.02 | 0.03 | 0.16 | 0.03 | 0.00 | 0.00 | 0.00 |
| <i>Haliotis iris</i> | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Maoricolpus roseus</i> | 2.05 | 5.48 | 1.47 | 1.27 | 1.96 | 0.65 | 0.45 | 1.54 | 0.24 | 0.56 | 2.33 | 0.41 |
| <i>Mauria</i> spp. | 0.00 | 0.00 | 0.00 | 0.03 | 0.07 | 0.02 | 0.00 | 0.00 | 0.00 | 0.03 | 0.18 | 0.03 |
| <i>Modiolarca impacta</i> | 0.24 | 0.89 | 0.24 | 1.94 | 4.73 | 1.58 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Monia zelandica</i> | 2.13 | 2.61 | 0.70 | 1.73 | 1.80 | 0.60 | 2.53 | 4.49 | 0.69 | 3.05 | 2.72 | 0.48 |
| <i>Trochus</i> spp. | 2.97 | 3.83 | 1.02 | 2.11 | 2.09 | 0.70 | 6.00 | 6.24 | 0.96 | 2.04 | 2.37 | 0.42 |
| <i>Turbo granosus</i> | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Turbo smaragdus</i> | 2.13 | 2.98 | 0.80 | 3.10 | 4.45 | 1.48 | 0.37 | 0.75 | 0.12 | 4.48 | 3.57 | 0.63 |
| <i>Galeolaria hystrix</i> | 20.55 | 30.05 | 8.03 | 13.41 | 19.88 | 6.63 | 1.03 | 1.98 | 0.30 | 0.85 | 1.46 | 0.26 |
| <i>Allostichaster insignis</i> | 0.01 | 0.04 | 0.01 | 0.03 | 0.07 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Coscinasterias calamaria</i> | 0.01 | 0.04 | 0.01 | 0.06 | 0.17 | 0.06 | 0.03 | 0.16 | 0.02 | 0.10 | 0.40 | 0.07 |
| <i>Evechinus chloroticus</i> | 1.45 | 1.26 | 0.34 | 0.80 | 0.74 | 0.25 | 1.53 | 1.46 | 0.22 | 2.36 | 1.84 | 0.33 |
| <i>Patiriella regularis</i> | 0.52 | 0.40 | 0.11 | 0.45 | 0.51 | 0.17 | 1.00 | 1.12 | 0.17 | 0.71 | 0.86 | 0.15 |
| <i>Pectinura maculata</i> | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.02 | 0.00 | 0.00 | 0.00 | 0.03 | 0.18 | 0.03 |
| <i>Pentagonaster pulchellus</i> | 0.07 | 0.18 | 0.05 | 0.11 | 0.28 | 0.09 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Pseudechinus albocinctus</i> | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Stichaster australis</i> | 0.02 | 0.06 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Stichopus mollis</i> | 0.37 | 0.34 | 0.09 | 1.04 | 1.10 | 0.37 | 0.11 | 0.31 | 0.05 | 0.07 | 0.27 | 0.05 |
| <i>Aplidium</i> sp. (%) | 0.04 | 0.13 | 0.04 | 0.00 | 0.00 | 0.00 | 0.15 | 0.36 | 0.06 | 0.03 | 0.18 | 0.03 |
| <i>Aplidium benham</i> | 0.00 | 0.00 | 0.00 | 0.69 | 2.07 | 0.69 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| <i>Leptoclinides</i> sp. (%) | 0.08 | 0.16 | 0.04 | 0.31 | 0.60 | 0.20 | 0.10 | 0.30 | 0.05 | 0.15 | 0.46 | 0.08 |
| <i>Cnemidocarpa bicornuata</i> | 1.92 | 2.07 | 0.55 | 2.23 | 2.15 | 0.72 | 0.21 | 0.53 | 0.08 | 0.28 | 0.70 | 0.12 |
| Unidentified warty squirt | NA | NA | NA | NA | NA | NA | 0.41 | 1.19 | 0.18 | 0.73 | 1.31 | 0.23 |
| <i>Didemnum</i> sp. (%) | 0.66 | 0.67 | 0.18 | 0.85 | 1.51 | 0.50 | 0.48 | 0.62 | 0.10 | 0.21 | 0.41 | 0.07 |
| <i>Carpophyllum flexuosum</i> (stipes) | 2.82 | 10.47 | 2.80 | 18.93 | 43.86 | 14.62 | 7.81 | 21.37 | 3.30 | 1.41 | 2.85 | 0.50 |
| <i>Carpophyllum flexuosum</i> (%) | NA | NA | NA | NA | NA | NA | 14.91 | 18.72 | 2.89 | 7.53 | 13.17 | 2.33 |
| <i>Corallina</i> (%) | 49.82 | 29.68 | 7.93 | 36.67 | 29.58 | 9.86 | 44.31 | 41.31 | 6.37 | 25.38 | 26.33 | 4.65 |

5.6 Cook's turban density and size

Cook's turban density data collected in April 2002 are presented in Appendix 14. Previous Cook's turban density data collected in December 1993 were presented in Davidson (1999). Mean size for pooled reserve and control treatments for 1993 and 2002 are given in Table 7 (see also Appendix 22).

Cook's turban density increased at both pooled reserve and control treatments between the two sample occasions (Figure 24). In 2002, the density of Cook's turban was comparable between control and reserve treatments. Cook's turban size was comparable between pooled reserve and control treatments and varied little between sample occasions (Figure 26, Table 7).

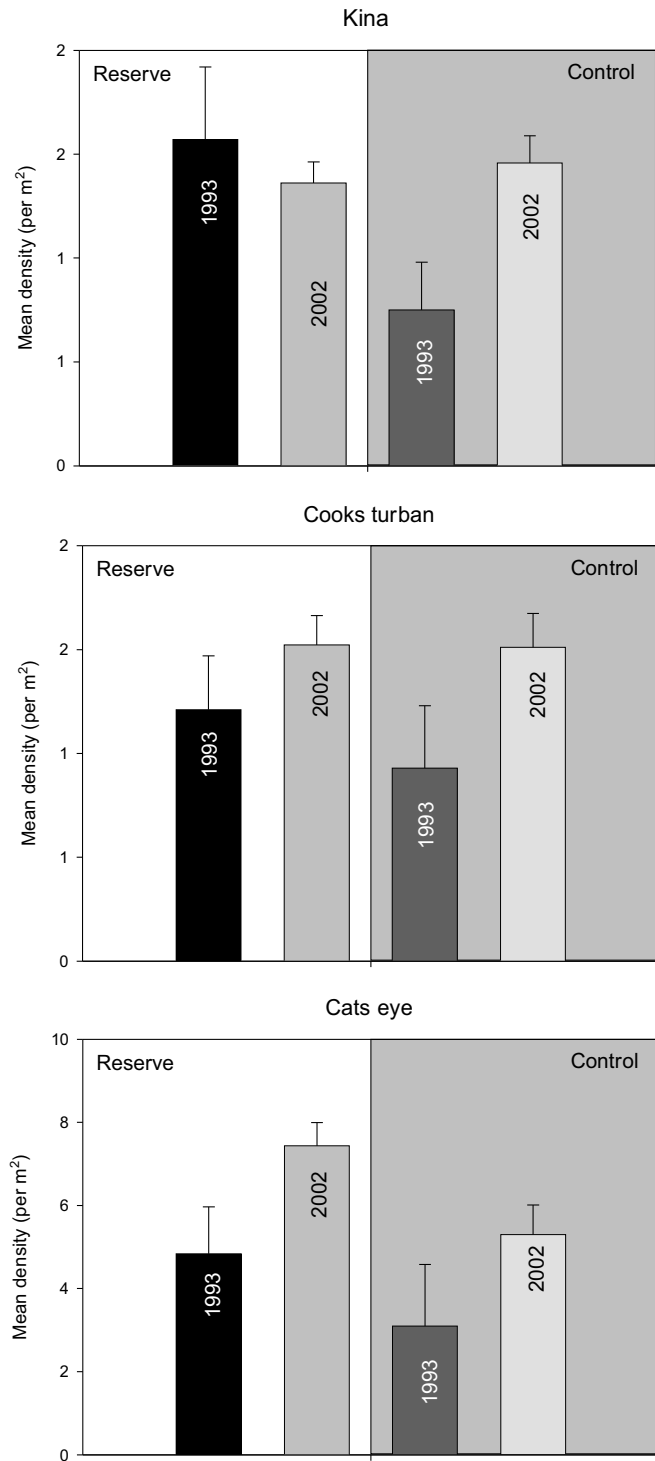


Figure 24. Mean density of kina, Cook's turban and cats-eye from pooled reserve and control sites in 1993 and 2002. Error bars are +/- 1 s.e.



Table 7. Mean size of invertebrates from reserve and control treatments in 1993 and 2002.

| | Kina | | | | Cook's Turban | | | | Cats eye | |
|------|--------------|--------------|--------------|--------------|---------------|--------------|--------------|--------------|--------------|--------------|
| | 1993 Reserve | 1993 Control | 2002 Reserve | 2002 Control | 1993 Reserve | 1993 Control | 2002 Reserve | 2002 Control | 1993 Reserve | 1993 Control |
| Mean | 56.62 | 53.95 | 59.80 | 51.45 | 42.38 | 42.80 | 42.03 | 42.87 | 29.90 | 30.46 |
| SD | 6.34 | 9.60 | 9.57 | 8.92 | 10.60 | 9.47 | 9.39 | 9.44 | 4.50 | 3.87 |
| SE | 0.34 | 0.52 | 0.52 | 0.48 | 0.52 | 0.49 | 0.39 | 0.49 | 0.23 | 0.19 |

5.7 Cats-eye snail density

Cats-eye snail raw density data collected in April 2002 are presented in Appendix 18. Previous cats-eye density data collected in December 1993 were presented in Davidson (1999). Mean size for pooled reserve and control treatments for 1993 are given in Table 7. No size data collected in 2002. Cats-eye density increased at both pooled reserve and control treatments between the two sample occasions (Figure 24). In both 1993 and 2002, the pooled density was higher in the reserve treatment compared to the control.

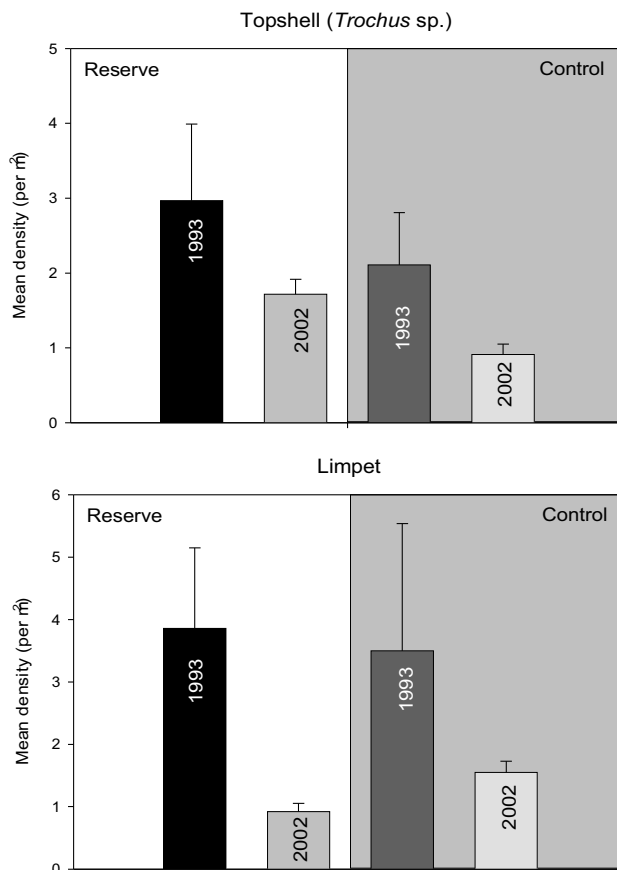


Figure 25. Mean density of topshell and limpet from pooled reserve and control sites in 1993 and 2002. Error bars are +/- 1 s.e.



5.8 Topshell density

Topshell (*Trochus* sp.) raw density data collected in April 2002 are provided in Appendix 19. Previous topshell density data collected in December 1993 were presented in Davidson (1999). Topshell density declined at both pooled reserve and control treatments between the two sample occasions (Figure 25). In both years, the pooled density was higher for the reserve treatment.

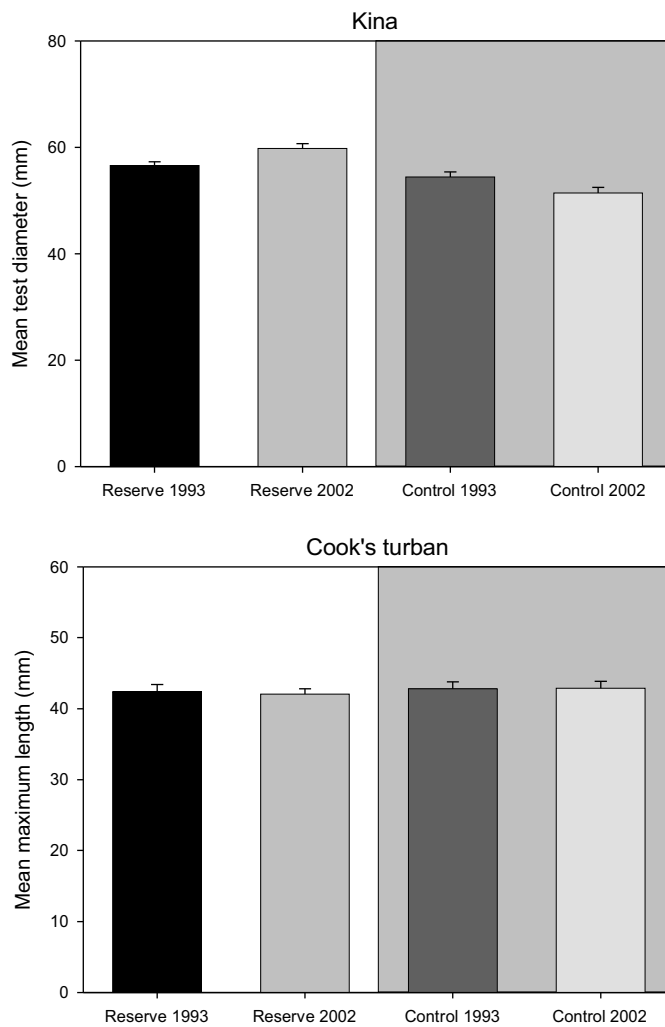


Figure 26. Mean size of kina and Cook's turban from pooled reserve and control sites in 1993 and 2002. Error bars are +/- 95% confidence.



5.9 Limpet density

Limpet (*Cellana* sp.) raw density data collected in April 2002 are given in Appendix 20. Previous limpet density data collected in December 1993 were presented in Davidson (1999). Limpet density declined at both pooled reserve and control treatments between the two sample occasions (Figure 25). The pooled density was higher for the reserve treatment in 1993 but lower in 2002 compared to the control treatment.

5.10 Scallop size and density

Scallop size and density data collected in March 2003 are presented in Appendices 23, 24, 30 and 31. Scallop data collected in 1994 and 1999 were presented in Davidson (1999, 2001).

Mean density of scallops in both reserve and control sites were comparable in 1994 and 1999 (Figure 27). In 2003, scallop density increased in the reserve. At control sites, the density of scallops remained relatively stable and significantly lower than at the reserve for the 2003 and 2006 samples (t-test, $P < 0.007$, $df = 47$).

Scallop size showed comparable patterns at both reserve and control treatments over the duration of the study (Figures 28 and 29). To date scallops size data suggest that reservation has little impact on the size of scallops in the population.

5.11 Horse mussel density

Horse mussel density data collected in March 2003 are presented in Appendices 25 and 26. Horse mussel density data collected in 1994 and 1999 were presented in Davidson (1999, 2001).

Mean horse mussels density at both reserve and control sites were very low at the start of the study in 1994 (Figure 27). At control sites, the density of horse mussels increased dramatically in 1999 and 2003, but declined back to low levels in 2006 (Figure 27). At reserve sites, densities were also low at the start of the study in 1994 and 1999, but increased in 2003 and 2006. Reserve densities at the end of the study were significantly higher than horse mussel densities recorded in 1993 (t-test, $P < 0.00001$, $df = 36$).

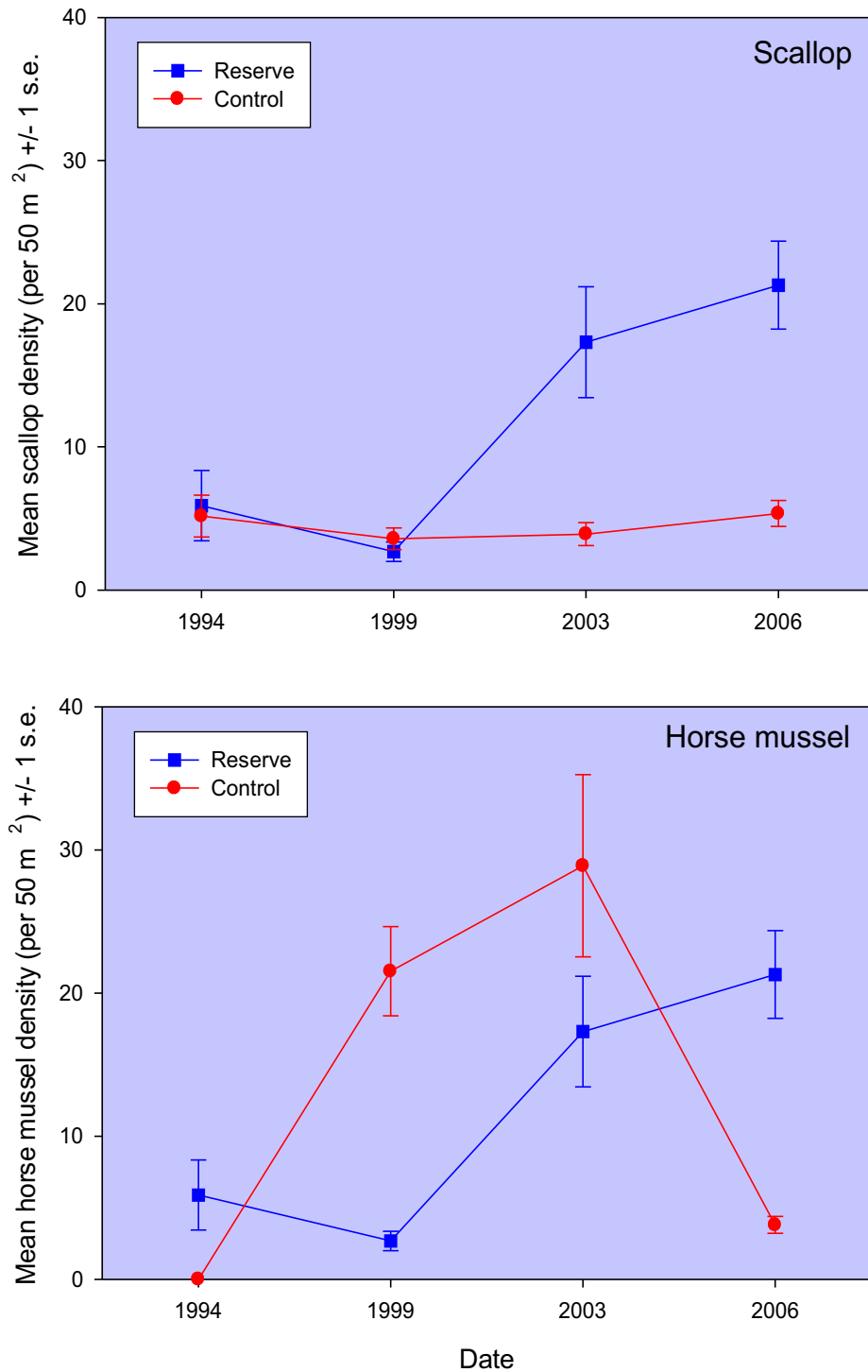


Figure 27. Mean density of scallops and horse mussels from pooled reserve and control sites. Error bars are +/- 1s.e. Note: x axis intervals are unequal time scales.

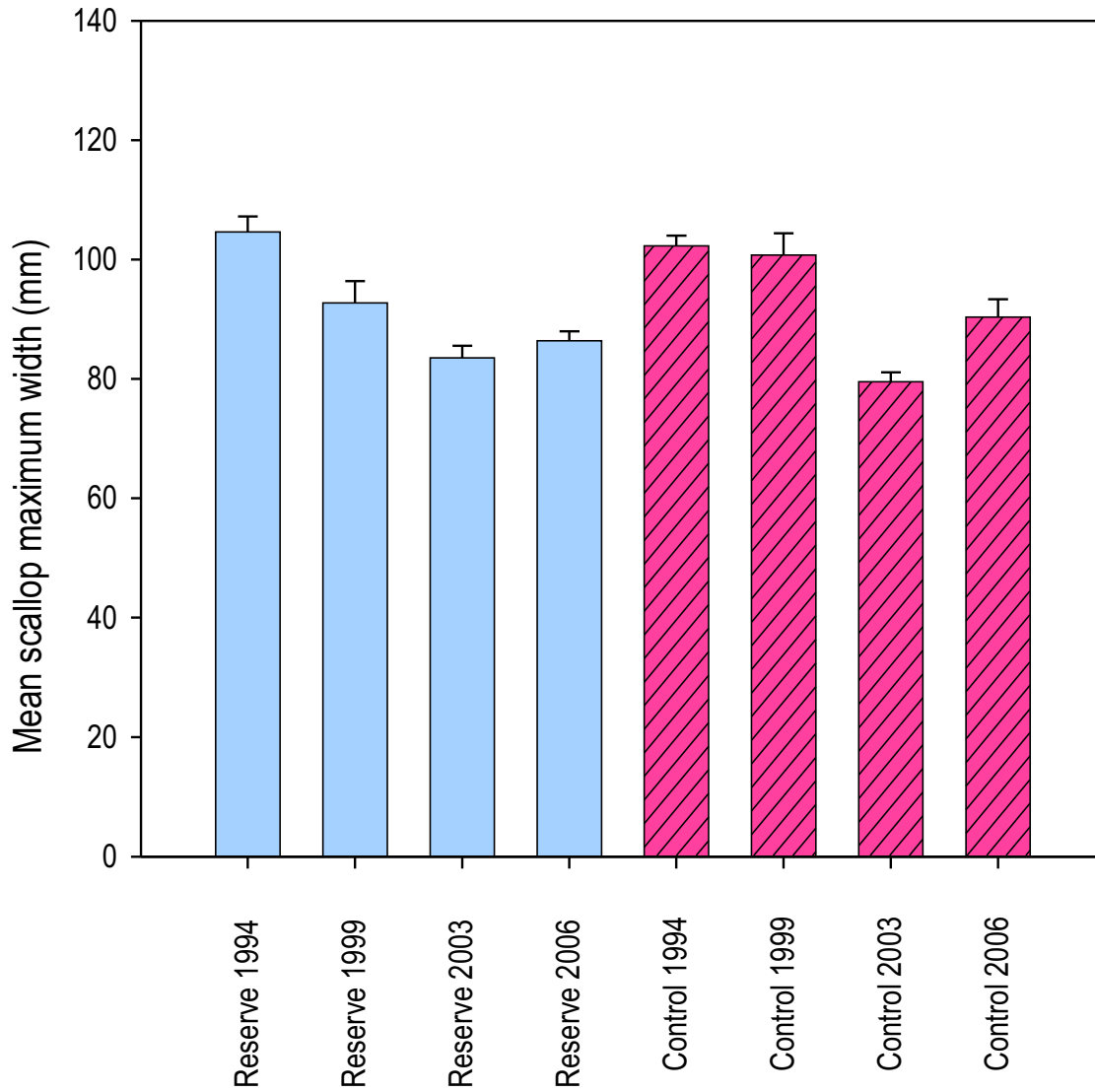


Figure 28. Mean size of scallops from individual reserve and control sites in 1994, 1999, 2003 and 2006. Error bars are 95% confidence. Blue bars = reserve, red hatched bars = control.

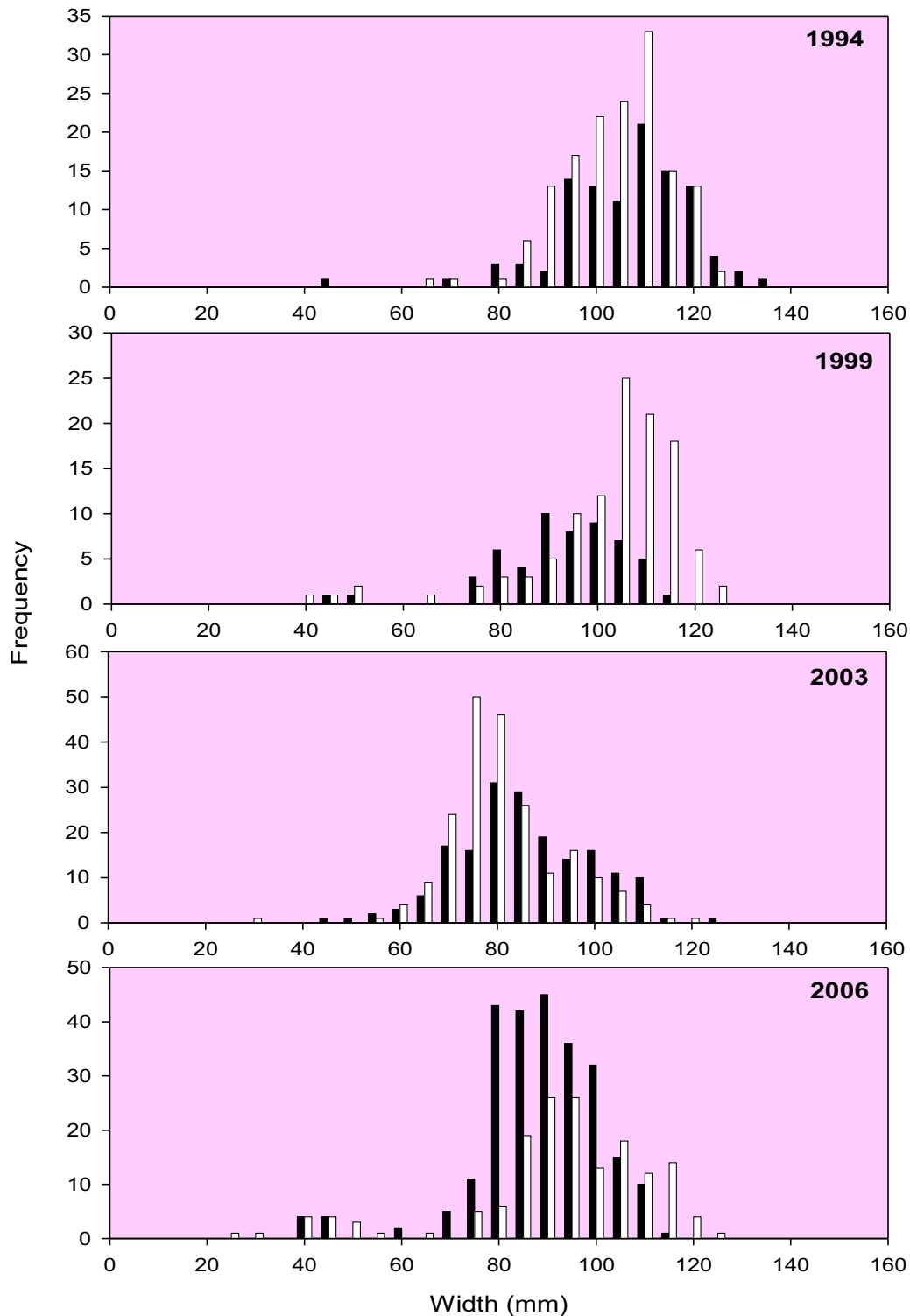


Figure 29. Size-frequency of scallops from pooled reserve (black bars) and control (open bars) treatments for each sample occasion (1994, 1999, 2003 and 2006).



6.0 DISCUSSION AND CONCLUSIONS

This report spans 15 years of monitoring from 1993 to 2007, although no data were collected from 1995 to 1998.

A growing body of studies that have shown changes in marine reserves in New Zealand (McCormick and Choat, 1987; Cole *et al.*, 1990; Creese and Jeffs, 1993; Jones *et al.*, 1993; MacDiarmid and Breen, 1993; Cole, 1994; Cole and Keuskamp, 1998; Kelly, 1999; Kelly *et al.*, 1999; Kelly *et al.*, 2000; Willis *et al.*, 2000; Cole *et al.*, 2000; Davidson, 2001; Davidson *et al.*, 2002, Shears *et al.*, 2006). These biological changes have generally been directly attributed to the cessation of fishing.

In general, the changes attributed to reservation have been relatively small and restricted to particular species. For example, Freeman (2005) reported little or no change as a result of reservation for species that had been previously harvested such as snapper, tarakihi, blue cod and blue moki in Te Tapuwae o Rongokako Marine Reserve, on the North Island's East Coast (2000 to 2004). Relatively few fish species within the Tonga Island Marine Reserve have increased in size or abundance since reservation, a period of 15 years. With the exception of blue cod and blue moki, no other reef fish appear to have responded to the exclusion of fishing. The increase in size and abundance of blue cod and blue moki has been relatively slow with few changes recorded prior to 2002. It is probable that this slow response to reservation at Tonga Island Marine Reserve is due to a variety of factors including:

- The biology of particular reef fish species (e.g. tarakihi being migratory),
- Heavy fishing pressure and low stocks prior to reservation,
- High levels of continued fishing pressure on the adjacent coast,
- Illegal fishing within the reserve,
- The isolation of the reef habitats of Abel Tasman coastline within Tasman and Golden Bays potentially limiting recruitment through migration from adjacent areas,
- The low energy and low productivity coastline (see Davidson and Chadderton 1994), and
- The relatively low fertility of adjacent the catchments.

It is possible that recovery of many species within the Tonga Island Marine Reserve will take a relatively long period compared to many marine reserves around New Zealand. It's sheltered, low productivity and



isolated situation may mean that some changes will be difficult to detect and may occur over a very long period.

6.1 Reef fish

Most New Zealand studies that have investigated the impact of reservation on reef fishes have adopted traditional diver strip counts to determine fish abundance (e.g. McCormick and Choat 1987, Cole *et al.*, 1992, Cole 1994, Cole *et al.* 2000, Davidson 2001, Freeman 2005). Willis *et al.* (2000), however, questioned the reliability of this method and reported that other methods such as baited video stations may be a more reliable method for determining fish abundance. The development of such new methods may thus be an important consideration for present and future marine reserve monitoring. For example, in areas subjected to consistently poor water visibility, alternative methods to diver visual counts will require development and testing.

The present study adopted traditional visual transect methodology to estimate fish density for a variety of reasons. Water visibility in this area was often unsuitable (i.e. < 4 m), however, on occasions was adequate for this method (i.e. > 4.5 m horizontal distance). The habitat was relatively homogeneous and the target species were relatively diver-neutral, providing divers did not disturb the benthos. As a precautionary approach, a large number of replicates (n = 12 per site) were collected at each site.

The present study demonstrated that the exclusion of fishing resulted in direct changes to particular species of reef fish in the Tonga Island Marine Reserve, especially blue cod. Blue cod population size structure, abundance and behaviour in the reserve all exhibited change. For example, in February 2005, large blue cod (> 300 mm TL) in the reserve represented 43.5% of the population compared to three individuals out of 90 blue cod (< 3.3%) from all control sites. Davidson and Richards (2004) measured blue cod using baited underwater methodology (BUV) in the Tonga Island Marine Reserve. The authors also reported that blue cod were larger from the marine reserve compared to the adjacent control sites. Cole *et al.* (2000) suggested that even relatively small marine reserves would protect blue cod as a proportion of the population would remain in the same physical area for long periods (i.e. several years).

Using BUV methods, Davidson and Richards (2004) reported that mean length values for blue cod were generally greater at Tonga Island Marine Reserve sites (i.e. often \geq 300 mm mean length) compared to Long Island-Kokomohua Reserve sites (i.e. usually < 300 mm mean length). Results collected during the present study have used underwater visual census methods (UVC) that are not directly comparable to size data collected using BUV methods due to differences in the methodology and diver estimate error. Regardless of these considerations blue cod were larger inside the both reserves compared to outside regardless of the methodology used.



At Long Island-Kokomohua Marine Reserve the proportion of very large blue cod > 330 mm in length, peaked in the reserve in April 1999 at 38.5% compared to the lowest value of < 1% for April 2000 at control sites (Davidson 2001). In the present study, large blue cod (> 330 mm TL) represented 46% of the total reserve population in the Tonga Island Marine Reserve compared to 0% for the control treatment. These proportions at Tonga Island Marine Reserve represented a dramatic difference between inside and outside the reserve.

The total abundance of blue cod has increased at both of the Nelson-Marlborough marine reserves but this increase has been more rapid and total abundance has reached higher levels at Long Island-Kokomohua Marine Reserve (Davidson 2004, Davidson and Richards 2004). For example, the mean density of all blue cod at Long Island has consistently been over six individuals per 60 m² since April 1999. In comparison, blue cod density for Tonga Island Marine Reserve has not risen above 1.2 individuals per 60 m² over the duration of the study. The dramatic difference in abundance between the two reserves is probably related to various factors including the proximity of Long Island to large areas of blue cod habitat in the Marlborough Sounds and Cook Strait. It is probable that recruitment into the reserve from movement and larval settlement is considerably higher at Long Island than for the Abel Tasman coast. Differences in the environmental variables operating at each marine reserve will also be important factors. It is unknown what density blue cod will reach in the Tonga Island Marine Reserve, however, it is certain that the abundance of this species has been on the rise.

Other reef fish species have also been slow to increase in abundance and size in Tonga Island Marine Reserve. Blue moki has shown relatively small but positive changes in the reserve. It is probable that blue moki recovery will occur over a relatively long time scale due to the isolation of the Abel Tasman coast from other reef habitats. Tarakihi numbers overall do not appear to be influenced by the absence of fishing. Tarakihi can be highly migratory and this will influence what changes occur in a marine reserve. Schools of very small individuals (6-14 cm) are regularly observed feeding on rock surfaces in the reserve and control areas, but large individuals > 30 cm are rare in both reserve and control areas.

Snapper were recorded from BUV stations located within the reserve but not outside the reserve by Davidson and Richards (2004). During the present study, snapper were not recorded from any transects, however, occasional individuals and one school of small snapper have been observed in the reserve since 2002 by divers. No snapper have been observed by divers outside the reserve during the study. As snapper usually avoid divers it is probable that their abundance has been underestimated using UVC methods.

6.2 Spiny lobsters

Spiny lobsters are intensively fished in many areas of New Zealand (Lipcius and Cobb, 1994). Several studies have shown abundance and size of spiny lobsters to be greater in protected areas than in nearby fished areas (e.g. MacDiarmid and Breen, 1993; Edgar and Barrett, 1999; Kelly *et al.*, 1999, 2000; Davidson *et al.*, 2002, Davidson 2004, Haggitt and Kelly 2004, Shears *et al.*, 2006). Those findings suggest



that some lobsters remain within non-fished areas, but there is also evidence of movement across reserve borders with some lobsters leaving reserves making them susceptible to capture (e.g. Kelly *et al.*, 2000; Kelly, 2001, Freeman in prep.). There is also evidence that egg production may be limited in intensively-fished populations that lack large males (MacDiarmid and Butler, 1999).

Kelly *et al.* (2000) estimated the increase in population abundance of spiny lobsters in northern marine reserves to be about 9% per year. Davidson (2004) reported that lobster abundance increased in the Long Island-Kokomohua Marine Reserve from 1.9 individuals per 100 m² in 1992 to 7.5 individuals per 100 m² in 2003 representing an increase in abundance of approximately 22% per year. Davidson *et al.* (2002) estimated a 4.4% per year increase for combined deep and shallow samples from Tonga Island Marine Reserve based on two sets of abundance data collected from 1994 and 1999.

Declines in lobster abundance has also been documented for marine reserve areas. Haggitt and Kelly (2003, 2004b) investigated the recovery of lobsters following a 80 % reduction in their abundance from Cape Rodney to Okakari Point Marine Reserve between 1995 and 2001. In Te Whanganui a Hei Marine Reserve, Haggitt and Kelly (2004a) reported a 24% decline in lobster abundance between 2003 and 2004. The authors stated that this decline was largely due to the loss of legal-sized lobsters within the reserve from intense fishing along the reserve boundaries.

Increased abundance and size of spiny lobster lobsters were the first documented biological change due to the cessation of fishing in the Tonga Island Marine Reserve. Over the life of the reserve, lobster densities have fluctuated between sample occasions but overall their density and size have increased compared to control areas. For example, lobster abundance doubled for deeper reserve samples from 1.01 individuals per 100 m² in 1994 to 2.17 individuals per 100 m² in 2000. From 2000 to 2007, the population increased more rapidly from 2.17 individuals per 100 m² to 7.3 individuals per 100 m². Similarly lobster density also increased in the shallow reserve sites over the same period, but this increase was less pronounced (0.55 individuals per 100 m² in 1994 to 3.56 individuals per 100 m² in 2007).

An increase in deep lobster density from 0.65 individuals per 100 m² in 1994 to 4.1 individuals per 100 m² in 2002 also occurred for the pooled control treatment. This increase was primarily due to the increase in abundance of lobsters at one control site (Separation Point). This natural increase in lobster abundance outside the reserve was also recorded by Davidson (2004) over the same period in the outer Queen Charlotte Sound, and was probably due to high natural settlement and subsequent growth of lobsters on the reef. This increase in lobster density was, however, reversed in subsequent years with numbers at Separation Point dropping dramatically after the peak in 2002 to the most recent sample in 2007.

It is unlikely that the changes in lobster methodology between samples collected in 1994 and those collected from 1998 onwards (i.e. a reduction in the sample area and increase in the number of replicates) would be responsible for changes in detected densities of lobsters in the Tonga Island Marine Reserve. MacDiarmid (1991) compared the precision of three different transect sizes (10 m x 10 m (n=20), 25 m x



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10 m (n=8) and 50 m x 10 m (n=4) each covering a total areas of 2000 m². MacDiarmid (1991) found that all transects provided a similar level of precision.

The density of lobsters recorded from the deep zone in Tonga Island Marine Reserve in 2007 some 15 years after reserve establishment was 7.3 individuals per 100 m². This density was comparable to that found at Long Island-Kokomohua Marine Reserve after only seven years following that reserves establishment in 1993 (i.e. approximately 7-10 individuals per 100m²). Densities of lobsters at Long Island have continued to increase reaching a mean of 13 individuals per 100 m² in 2007 (Davidson unpublished data).

The abundance of the different lobster size classes in Tonga Island Marine Reserve (i.e. juveniles, mature females, non-reproductive males and reproductive males) varied both over time and between and within reserve and control treatments. It should be noted that sex-size class densities represent underestimates of the real population as not all lobsters observed in quadrats could be both sexed and measured due to the habitat complexity.

The abundance of juveniles (< 85 mm CL) remained comparable between reserve and control treatments with a peak in December 2002 for both treatments. The similarity in the abundance of juveniles between reserve and control treatments was not unexpected as this size class is well below the legal size limit for lobsters and should not be removed from the control reefs by fishers. For the other size and sex classes the abundance remained relatively stable for the control treatment over the duration of the study. In contrast the removal of fishing from the reserve has resulted in an increase in the abundance of the larger male and female size classes. This was particularly obvious for mature females (> 86 mm CL) and reproductive males (> 139 mm CL) since 2004.

Davidson *et al.* (2002) estimated that approximately nine times as many eggs would be produced from females in the Tonga Island Marine Reserve compared to the equivalent length of non reserve coast. This estimate was based on the mean size of reproductive female lobsters, their density and known egg production. Since that time, the abundance of reproductive females has more than doubled in the reserve ensuring that egg production from the reserve will be dramatically higher when compared to the adjacent fished coast. Large male abundance in the Tonga Island Marine Reserve has increased over three-fold since Davidson *et al.* (2002) sampled lobsters. High numbers of large males (>139 mm CL) helps ensure high fertilisation rates compared to control areas where large males are relatively rare.

6.3 Scallops and horse mussels

Between 1994 and 2006 horse mussel density at the control site increased dramatically and then declined. In contrast, the abundance of horse mussels from the reserve sites was initially low (1994-1999) but increased in subsequent samples (2003 and 2006).



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Changes in horse mussel densities in the reserve and control treatments could be due to a variety of reasons including natural recruitment and mortality patterns as well as the effects of dredging. Further monitoring of the reserve and control sites may shed light in this issue.

Scallop density remained relatively low at the control treatment throughout the study, but increased at the reserve treatment in 2003 and again in 2006, relative to densities recorded in the first two samples. In contrast scallop mean size data showed no differences between reserve and control treatments. These data suggest that the exclusion of dredging from the reserve has allowed their abundance to increase.

6.4 Other species

Occasional monitoring of shore profiles and various benthic rocky species occurs as part of the present monitoring programme. These data have been collected in order to detect long term community changes and to assess effects on non-harvested species. For most of these species, monitoring has occurred on only two occasions, once in 1993 and once several years after reservation. At this stage insufficient temporal data exists to draw any conclusions regarding the impact of reservation on these species or community structure.

7.0 FUTURE BIOLOGICAL MONITORING

The current monitoring programme funded by the Department of Conservation, Nelson is carried out by Davidson Environmental Ltd. with assistance from Department staff of the Motueka Area. This study has spanned a period of 15 years and has detected changes that can be attributed to the establishment of the marine reserve (i.e. increased abundance and size). Some species, however, have shown no detectable response to reservation. It is important that the monitoring programme be continued in order to detect changes as they occur. Longer term monitoring will also detect community level changes that may take many more years to occur. Based on results collected as part of the present study and the sampling protocol produced by Davidson (2001), the following monitoring is recommended over the next three to four years.

Spiny lobsters

Lobsters should be counted, sexed and sized annually from reef sites outlined by Davidson (2001). Two additional sites to those outlined in Davidson (2001) have been monitored during this study (i.e. a reserve site located at Mosquito Reef and a new reserve site at Totara Rock). It is recommended that these sites continue to be monitored as part of the annual lobster programme.



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Fish

Fish densities using traditional visual underwater count methodology (UVC) should be collected annually from reef sites as outlined by Davidson (2001).

Macro-invertebrates

Kina size and density should be investigated every fifth year as outlined by Davidson (2001). Other key invertebrate densities should be sampled every fifth year as outlined by Davidson (2001).

Shore profiles

Particular shore profiles should be re-sampled once before 2010.

Scallops and horse mussels

Scallops and horse mussels should be sampled once every third year as outlined by Davidson (2001). The next sample is due in the summer of 2008-2009.



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REFERENCES

- Arrenguín – Sanchez, F. 1996. Catchability: a key parameter for fish stock assessment. *Review Fisheries Biology Fish* 6: 221-242.
- Bell, J.D. 1983. Effects of depth and marine reserve fishing restrictions on the structure of a rocky reef fish assemblage in the northwestern Mediterranean Sea. *Journal of Applied Ecology* 20: 357-369.
- Bennett, B.A. and Attwood, C.G. 1993. Shore-angling catches in the DeHoop Nature Reserve, South Africa. *South African Journal of Marine Science* 13: 213-222.
- Bennett, B.A. and Attwood, C.G. 1991. Evidence for recovery of a surf-zone fish assemblage following the establishment of a marine reserve on the southern coast of South Africa. *Marine Ecology Progress Series* 75: 173-181.
- Blackwell, R.G. 1998. Abundance, size and age composition, and yield-per-recruit of blue cod in the Marlborough Sounds, September 1996. NIWA Technical Report 30. 47p.
- Blackwell, R.G. 1997. Abundance, size composition, and sex ratio of blue cod in the Marlborough Sounds, September 1995. New Zealand Fisheries Data Report 88, NIWA, Wellington. 17p.
- Buxton, C.D. and Smale, M.J. 1989. Abundance and distribution patterns of three temperate fish in exploited and unexploited areas off the southern Cape coast. *Journal of Applied Ecology* 26: 441-451.
- Carbines, G. 1999. Determination of movement of blue cod in Southland. Final Research Report for Ministry of Fisheries. NIWA, Wellington. 20p.
- Carbines, G. 1998. Determination of movement of blue cod in Southland. Final Research Report for Ministry of Fisheries. NIWA, Wellington. 14p.
- Choat, J.H., Ayling, A.M., and Schiel, D.R. 1988. Temporal and spatial variation in an island fish fauna. *Journal of Experimental Marine Biology* 121: 91-111.
- Cole, R.G. 1994. Abundance, size structure, and diver-oriented behaviour of three large benthic carnivorous fishes in a marine reserve in northeastern New Zealand. *Biological Conservation* 70: 93-99.
- Cole, R.G., Villouta, E., and Davidson, R.J. 2000. Direct evidence of limited dispersal of the reef fish *Paraperis colias* (Pinguipedidae) within a marine reserve and adjacent fished areas. *Aquatic Conservation: Freshwater and Marine Ecosystems* 10: 421-436.
- Cole, R.G., Creese, R.G., Grace, R.V., Irving, P., and Jackson, B. 1992. Abundance patterns of subtidal benthic invertebrates and fishes at the Kermadec Islands. *Marine Ecology Progress Series* 82: 207-218.
- Cole, R.G., Ayling, T.M., and Creese, R.G. 1990. Effects of marine reserve protection at Goat Island, northern New Zealand. *New Zealand Journal of Marine and Freshwater Research* 24: 197-210.
- Cole, R.G. and Keuskamp, D. 1998. Indirect effects of protection from exploitation: patterns from populations of *Evechinus chloroticus* (Echinoidea) in northeastern New Zealand. *Marine Ecology Progress Series* 173: 215-226.
- Creese, R.G. and Jeffs, A. 1993. Biological research in New Zealand marine reserve. Pp. 15-22 in Battershill et al. (eds): Proceedings of the Second International Temperate Reef Symposium, 7-10 January 1992, Auckland, New Zealand. NIWA Marine, Wellington. 252p.



- Davidson, R.J. 2004. Long Island-Kokomohua Marine Reserve, Queen Charlotte Sound: 1992-2003. Prepared by Davidson Environmental Limited for Department of Conservation, Nelson. Survey and Monitoring Report No. 343.
- Davidson, R.J. 2001. Changes in population parameters and behaviour of blue cod (*Parapercis colias*) in Tonga Island Marine Reserve, Marlborough Sounds, New Zealand. *Aquatic Conservation: Marine and Freshwater Ecosystems* 11: 417-435.
- Davidson, R.J. 1992. A report on the intertidal and shallow subtidal ecology of the Abel Tasman National Park, Nelson. Department of Conservation, Nelson/Marlborough Conservancy Occasional Publication. No. 4, 161 p.
- Davidson, R.J. 1991. Tonga Island Marine Reserve: proposed protocol for ongoing subtidal biological monitoring. Prepared for the Department of Conservation by Davidson Environmental Ltd. Survey and Monitoring Report No. 316. 29p.
- Davidson, R.J. 1997. Biological monitoring of Long Island-Kokomohua Marine Reserve, Queen Charlotte Sound, Marlborough Sounds: update September 1993-April 1997. Prepared for the Department of Conservation by Davidson Environmental Ltd. Survey and Monitoring Report No. 150. 40p.
- Davidson, R. J. 2004. Long Island-Kokomohua Marine Reserve, Queen Charlotte Sound: 1992-2003 Prepared by Davidson Environmental Limited for Department of Conservation, Nelson. Survey and Monitoring Report No. 343.
- Davidson, R.J, Villouta, E., Cole, R.G., and Barrier, R.G.F. 2002. Effects of marine reserve protection on spiny lobster abundance and size at Tonga Island Marine Reserve, New Zealand. *Aquatic Conservation: Marine and Freshwater Ecosystems*. 12: 213-227.
- Davidson, R. J.; Richards L. A. 2004. Comparison of fish at reserve and control sites from two South Island marine reserves (Long Island-Kokomohua and Tonga Island) using baited underwater video. Prepared by Davidson Environmental Limited for Department of Conservation, Nelson. Survey and Monitoring Report No. 466.
- Davidson, R.J.; Chadderton, W.L., 1994. Marine reserve selection along the Abel Tasman National Park coast, New Zealand: consideration of subtidal rocky communities. *Aquatic Conservation: Freshwater and marine ecosystems* Vol. 4, 153-167.
- Deriso, R.B. and Parma, A.M. 1987. On the odds of catching fish with angling gear. *Transactions of American Fish Society* 116: 244-256.
- Dufour, V., Jouvenel, J., and Galzin, R. 1995. Study of a Mediterranean reef fish assemblage. Comparisons of population distributions between depths in protected and unprotected areas over one decade. *Aquatic Living Resources* 8: 17-25.
- Edgar, G.J. and Barrett, N.S. 1999. Effects of the declaration of marine reserves on Tasmanian reef fishes, invertebrates and plants. *Journal of Experimental Marine Biology and Ecology* 242: 107-144.
- Edgar, G.J. and Barrett, N.S. 1997. Short term biotic change in Tasmanian marine reserves. *Journal of Marine and Freshwater Research* 213: 261-279.
- Freeman, D.J. 2005. Reef fish monitoring Te Tapuwae o Rongokako Marine Reserve. Department of Conservation Technical Support Series Number 25.
- Freeman, D.J. and Duffy, C.A.J. 2003. Te Angiangi Marine Reserve: Reef fish monitoring 1995-2003. Department of Conservation, East Coast Hawke's Bay Conservancy, Technical Series No. 14.
- Garcia-Rubies, A. and Zabala, M. 1990. Effects of total fishing prohibition on the rocky fish assemblages of Medes Islands marine reserve (MW Mediterranean). *Scientifica Marina* 54: 317-218.



- Haggitt, T.R.; Kelly, S. 2003: Cape Rodney to Okakari Point Marine Reserve Lobster Monitoring Programme: 2003 Survey. Report to the Department of Conservation June 2003, 28 pp.
- Haggitt, T, Kelly, S. 2004. 2004a lobster survey of the Cathedral Cove (Te Whanganui A Hei) Marine Reserve. Prepared for Department of Conservation by Coastal and Aquatic Systems Ltd.
- Haggitt, T.R.; Kelly, S. 2004b: Cape Rodney to Okakari Point Marine Reserve Lobster Monitoring Programme: 2004 Survey. Report to the Department of Conservation June 2004, 31 pp.
- Haggitt, T.R.; Kelly, S. 2004c: Te Whanganui a Hei Marine Reserve Marine Reserve Biological Monitoring Plan. Report to the Department of Conservation June 2004, 36 pp
- Hilborn, R., Stokes, K., Maguire, Jean-Jacques, Smith, T., Botsford, W., Mangel, M., Orensanz, J., Parma, A., Rice, J., Bell, J., Cockrane, L., Garcia, S., Hall, S.J., Kirkwood, G.P., Sainsbury, K., Stefansson, G., Walters, C. 2004. When can marine reserves improve fisheries management? *Ocean and Coastal Management* 47: 197-205.
- Jennings, S. and Polunin, N.V.C. 1995. Biased underwater visual census biomass estimates for target-species in tropical reef fisheries. *Journal of Fish Biology* 47: 733-736.
- Jones, G.P., Cole, R.C. and Battershill, C.N. 1993. Marine reserves: do they work? Pp. 15-22 in Battershill *et al.* (eds): Proceedings of the Second International Temperate Reef Symposium, 7-10 January 1992, Auckland, New Zealand. NIWA Marine, Wellington. 252p.
- Kelly, S. 2001. Temporal variation in the movement of the spiny lobster *Jasus edwardsii*. *Marine and Freshwater Research* 52: 323-331.
- Kelly, S.; Scott, D.; MacDiarmid, A.B.; Babcock, R.C. 2000: Spiny lobster, *Jasus edwardsii*, recovery in marine reserves. *Biological Conservation* 92:359-369.
- Kelly, S. 1999. Marine reserves and the spiny lobster, *Jasus edwardsii*. Unpublished Ph.D thesis, University of Canterbury.
- Kelly, S., Scott, D., MacDiarmid, A.B., and Babcock, R.C. 2000. Spiny lobster, *Jasus edwardsii*, recovery in New Zealand marine reserves. *Biological Conservation* 92: 359-369.
- Kelly, S., MacDiarmid, A.B., and Babcock, R.C. 1999. Characteristics of spiny lobster, *Jasus edwardsii*, aggregations in exposed reef and sandy areas. *Marine and Freshwater Research* 50: 409-416.
- Kulbicki, M. 1998. How the acquired behaviour of commercial reef fishes may influence the results obtained from visual censuses. *Journal of Experimental Marine Biology and Ecology* 222: 11 - 30.
- Lipcius, R.N. and Cobb, J.S. 1994. Ecology and fishery biology of spiny lobsters. In: Spiny lobster management. Phillips *et al.* (eds.). Blackwell Scientific: London. Chapter 1.
- McCormick, M.I. and Choat, J.H. 1987. Estimating total abundance of a large temperate-reef fish using visual strip transects. *Marine Biology* 96: 469-478.
- McKnight, D.G. and Grange, K.R. 1991. Macrobenthos sediment-depth relationships in Marlborough Sounds. Report prepared for Department of Conservation by Oceanographic Institute, DSIR. No. P692. 19 p.
- MacDiarmid AB. 1989. Size at onset of maturity and size dependent reproductive output of female and male spiny lobsters *Jasus edwardsii* (Hutton) (Decapoda, Palinuridae) in northern New Zealand. *Journal of Experimental Marine Biology and Ecology* 89: 191-204.
- MacDiarmid, A.B. and Butler, M.J. 1999. Sperm economy and limitation in spiny lobsters. *Behavioral Ecology and Sociobiology* 46: 14-24.



- MacDiarmid, A.B. 1991: Seasonal changes in depth distribution, sex ratio and size frequency of spinylobster *Jasus edwardsii* on a coastal reef in northern New Zealand. *Marine Ecology Progressive Series* 70: 129-141.
- MacDiarmid, A.B. and Breen, P.A. 1993. Spiny lobster population change in a marine reserve. Pp. 15-22 in Battershill. *et al.* (eds): Proceedings of the Second International Temperate Reef Symposium, 7-10 January 1992, Auckland, New Zealand. NIWA Marine, Wellington. 252p.
- Mace, J.T. and Johnson, A.D. 1983. Tagging experiments on blue cod (*Parapercis colias*) in the Marlborough Sounds, New Zealand. *New Zealand Journal of Marine and Freshwater Research* 17: 207-211.
- Millar, R.B. and Fryer, R.J. 1999. Estimating the size-selection curves of towed gears, traps, nets and hooks. *Review Fisheries Biology Fish* 9: 89-116.
- Shears, N.T., Grace, R. V., Usmar, N. R., Kerr, V., and Babcock, R. 2006. Long-term trends in lobster populations in a partially protected vs. no-take Marine Park.. *Biological Conservation* 132: 222-231.
- Shears, N. T.; Babcock, R. C. 2003: Continuing trophic cascade effects after 25 years of no-take marine reserve protection. *Marine Ecology Progressive Series*. 246: 1-16
- Shears, N. T.; Babcock, R. C. 2002: Marine reserves demonstrate top-down control of community structure on temperate reefs. *Oecologia* 132: 131-142.
- Somerton, D.A. and Kikkawa, B.S. 1995. A stock survey technique using the time to capture individual fish on longlines. *Canadian Journal of Fisheries and Aquatic Sciences* 52: 260-267.
- Underwood, A.J. 1993. The mechanics of spatially replicated sampling programmes to detect environmental impacts in a variable world. *Australian Journal of Ecology* 18: 99-116.
- Villouta, E., Davidson, R.J., and Cole, R.G. (in prep.) Recruitment of blue cod (*Parapercis colias*) in the Marlborough Sounds, New Zealand.
- Willis, T.J. 2000. Te Whanganui A Hei fish monitoring Program: II. Change in Snapper and blue cod density. Report to the Department of Conservation, August 2000.
- Willis, T.J., Millar, R.B., Babcock, R.C., and Tolimieri, N. 2003. Burdens of evidence and the benefits of marine reserves for fisheries management: putting Descartes before des horse? *Environmental Conservation* 30: 97-103.
- Willis, T.J., Millar, R.B, and Babcock, R.C. 2000. Detection of spatial variability in relative density of fishes: comparison of visual census, angling, and baited underwater video. *Marine Ecology Progress Series* 198, 249-260.
- Wing, S. 2006. Baseline Ecological Monitoring of the Ulva Island / Te Wharawhara Marine Reserve. Report for Department of Conservation by Department of Marine Science, Otago.

Appendix 7. Visual fish data collected in February 2007 from boulder substrata from the Abel Tasman.

| Species | Site 1 Separation Point | | | | | | | | | | Site 2 Totaranui north | | | | | | | | | |
|-----------------------------------|-------------------------|----|---|---|---|----|----|---|----|----|------------------------|----|---|---|----|---|---|---|---|---|
| | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parapercis colias</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parapercis colias</i> | 2 | 1 | 1 | 0 | 0 | 2 | 0 | 3 | 4 | 0 | 0 | 0 | 4 | 5 | 0 | 4 | 5 | 0 | 0 | 0 |
| <i>Notolabrus celidotus</i> | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Notolabrus fucicola</i> | 1 | 10 | 0 | 2 | 9 | 21 | 4 | 6 | 19 | 4 | 2 | 15 | 4 | 2 | 15 | 0 | 0 | 0 | 0 | 0 |
| <i>Caesioperca lepidoptera</i> | 5 | 3 | 1 | 3 | 2 | 4 | 1 | 1 | 0 | 1 | 2 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pseudolabrus miles</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Odax pullus</i> | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheilodactylus nigripes</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Latridopsis ciliaris</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Nemadactylus macropterus</i> | 0 | 0 | 0 | 1 | 0 | 0 | 20 | 1 | 0 | 15 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheilodactylus spectabilis</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Aplodactylus arcidens</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Scorpius lineolatus</i> | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parika scaber</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicolenus percoides</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Upeneichthys lineatus</i> | 1 | 1 | 0 | 0 | 0 | 2 | 2 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Habitat | Boulder | | | | | | | | | | Bedrock | | | | | | | | | |
| Depth (m) | 6 - 12.5 | | | | | | | | | | 6 - 10 | | | | | | | | | |

Appendix 7. Visual fish data collected in February 2007 from boulder substrata from the Abel Tasman.

| Species | Site 3 Totaranui Reef | | | | | | | | | | Site 4 Awaroa Head | | | | | | | | | | Site 5 Canoe Bay | | | | | | | | | | | | | | | | | | | |
|-----------------------------------|-----------------------|---|---|---|---|----|---|---|---|---|--------------------|---|---|---|---|---|---|---|---|---|------------------|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|
| | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parapercis colias</i> | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parapercis colias</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Notolabrus celidotus</i> | 5 | 8 | 6 | 7 | 9 | 12 | 9 | 7 | 8 | 3 | 4 | 1 | 3 | 3 | 5 | 4 | 2 | 3 | 3 | 4 | 6 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Notolabrus fucicola</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Caesioperca lepidoptera</i> | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Pseudolabrus miles</i> | 2 | 2 | 0 | 1 | 0 | 2 | 1 | 2 | 0 | 1 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Odax pullus</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheilodactylus nigripes</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Latridopsis ciliaris</i> | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Nemadactylus macropterus</i> | 1 | 1 | 0 | 5 | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Cheilodactylus spectabilis</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Aplodactylus arcidens</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Scorpius lineolatus</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Parika scaber</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Helicolenus percoides</i> | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| <i>Upeneichthys lineatus</i> | 7 | 4 | 0 | 4 | 2 | 0 | 0 | 0 | 1 | 1 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Bedrock | 7 - 12 | | | | | | | | | | 7 - 12 | | | | | | | | | | Boulder | | | | | | | | | | | | | | | | | | | |
| Depth (m) | 7 - 12 | | | | | | | | | | 7 - 12 | | | | | | | | | | 7 - 12 | | | | | | | | | | | | | | | | | | | |

Appendix 8. Lobster sizes from December 2002.

Shallow

| Sample site | Treatment | Number | Depth | Total | | | | | Total |
|------------------|-----------|--------|-------|-------|---|---|---|----|-------|
| | | | | 1 | 2 | 3 | 4 | 5 | |
| Separation Point | Control | 1 | 7 | 3 | 0 | 1 | 1 | 3 | 8 |
| Awaroa Head | Control | 2 | 7 | 0 | 0 | 0 | 0 | 7 | 7 |
| Cottage Loaf | Reserve | 3 | 7 | 0 | 0 | 2 | 3 | 7 | 12 |
| Tonga Island | Reserve | 4 | 7 | 1 | 0 | 0 | 4 | 2 | 7 |
| Foul Point | Reserve | 5 | 6 | 26 | 8 | 5 | 4 | 13 | 56 |
| Whale Rock | Reserve | 6 | 7 | 0 | 0 | 0 | 1 | 3 | 4 |
| Mosquito Bay | Reserve | 7 | 7 | 3 | 6 | 9 | 5 | 7 | 30 |
| Bark Bay Reef | Control | 8 | 7 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pitt Head | Control | 10 | 7 | 3 | 0 | 3 | 0 | 0 | 6 |
| | | | | | | | | | 21 |
| | | | | | | | | | 109 |
| | | | | | | | | | 130 |

Deep

| Sample site | Treatment | Number | Depth | Total | | | | | Total |
|------------------|-----------|--------|-------|-------|----|----|----|----|-------|
| | | | | 1 | 2 | 3 | 4 | 5 | |
| Separation Point | Control | 1 | 10 | 0 | 13 | 20 | 31 | 9 | 73 |
| Awaroa Head | Control | 2 | 11 | 0 | 0 | 1 | 0 | 0 | 1 |
| Cottage Loaf | Reserve | 3 | 11 | 3 | 0 | 0 | 3 | 9 | 15 |
| Tonga Island | Reserve | 4 | 10 | 7 | 3 | 1 | 7 | 11 | 29 |
| Foul Point | Reserve | 5 | 10 | 3 | 11 | 0 | 0 | 2 | 16 |
| Whale Rock | Reserve | 6 | 10 | 0 | 0 | 1 | 0 | 3 | 4 |
| Mosquito Bay | Reserve | 7 | 9 | 13 | 0 | 0 | 1 | 8 | 22 |
| Bark Bay Reef | Control | 8 | 11 | 0 | 0 | 1 | 0 | 3 | 4 |
| Pitt Head | Control | 10 | 10 | 0 | 0 | 3 | 1 | 0 | 4 |
| | | | | | | | | | 82 |
| | | | | | | | | | 86 |
| | | | | | | | | | 168 |

Appendix 9. Lobster sizes from February 2004.

Shallow

| Sample site | Treatment | Number | Total | | | | | Total All sites |
|------------------|-----------|--------|-------|----|---|---|----|-----------------|
| | | | 1 | 2 | 3 | 4 | 5 | |
| Separation Point | Control | 1 | 1 | 0 | 0 | 0 | 8 | 9 |
| Awaroa Head | Control | 2 | 0 | 0 | 1 | 0 | 1 | 2 |
| Cottage Loaf | Reserve | 3 | 0 | 0 | 0 | 3 | 4 | 7 |
| Tonga Island | Reserve | 4 | 0 | 0 | 1 | 1 | 5 | 7 |
| Foul Point | Reserve | 5 | 1 | 12 | 2 | 0 | 0 | 15 |
| Whale Rock | Reserve | 6 | 3 | 0 | 0 | 4 | 0 | 7 |
| Mosquito Bay | Reserve | 7 | 0 | 2 | 3 | 4 | 11 | 20 |
| Bark Bay Reef | Control | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
| Totara Rock | Control | 9 | 0 | 2 | 0 | 0 | 0 | 2 |
| Pitt Head | Control | 10 | 3 | 0 | 0 | 3 | 0 | 6 |
| | | | | | | | 56 | 19 |
| | | | | | | | | 75 |

Deep

| Sample site | Treatment | Number | Total | | | | | Total All sites |
|------------------|-----------|--------|-------|----|----|---|----|-----------------|
| | | | 1 | 2 | 3 | 4 | 5 | |
| Separation Point | Control | 1 | 23 | 15 | 12 | 6 | 6 | 62 |
| Awaroa Head | Control | 2 | 0 | 1 | 0 | 1 | 1 | 3 |
| Cottage Loaf | Reserve | 3 | 0 | 0 | 0 | 1 | 16 | 17 |
| Tonga Island | Reserve | 4 | 1 | 0 | 7 | 7 | 2 | 17 |
| Foul Point | Reserve | 5 | 12 | 9 | 0 | 0 | 0 | 21 |
| Whale Rock | Reserve | 6 | 1 | 3 | 0 | 3 | 1 | 8 |
| Mosquito Bay | Reserve | 7 | 2 | 7 | 0 | 7 | 4 | 20 |
| Bark Bay Reef | Control | 8 | 0 | 0 | 1 | 3 | 1 | 5 |
| Totara Rock | Control | 9 | 8 | 4 | 0 | 0 | 0 | 12 |
| Pitt Head | Control | 10 | 0 | 0 | 2 | 0 | 0 | 2 |
| | | | | | | | 83 | 84 |
| | | | | | | | | 167 |

Appendix 10. Lobster sizes and density from shallow quadrats in April 2006.

| Site name | Site no. | Treatment | N quadrats | Depth (m) | Habitat | Number per quadrat | | | | | Total n | Sex (includes extras outside quadrats) | | | Density (100m ²) | SD | SE | |
|------------------|----------|-----------|------------|-----------|---------|--------------------|----|---|----|---|-----------------------------------|--|--------|-----------|------------------------------|------|------|---------|
| | | | | | | 1 | 2 | 3 | 4 | 5 | | Male | Female | Juvenile | | | | Unknown |
| SHALLOW | | | | | | | | | | | | | | | | | | |
| Cottage Loaf | RL1 | Reserve | 10 | 5-7 m | Boulder | 2 | 0 | 6 | 7 | 0 | 15 | 11 | 3 | 1 | 0 | 3.00 | 3.32 | 1.05 |
| Tonga Island | RL2 | Reserve | 10 | 5-7 m | Boulder | 9 | 6 | 2 | 5 | 1 | 23 | 16 | 5 | 1 | 1 | 4.60 | 3.21 | 1.01 |
| Foul Point | RL3 | Reserve | 10 | 5-6 m | Boulder | 9 | 4 | 9 | 4 | | 26 | 10 | 8 | 6 | 2 | 6.50 | 2.89 | 0.91 |
| Whale Rock | RL4 | Reserve | 10 | 5-7 m | Boulder | 18 | 12 | 8 | 3 | 6 | 47 | 24 | 21 | 2 | 0 | 9.40 | 5.81 | 1.84 |
| Mosquito Bay | RL5 | Reserve | 10 | 7-6 m | Boulder | 1 | 5 | 0 | 0 | 0 | 6 | 3 | 3 | 0 | 0 | 1.20 | 2.17 | 0.69 |
| Separation Point | CL1 | Control | 10 | 5-6 m | Boulder | 4 | 2 | 2 | 10 | 3 | 21 | 9 | 4 | 3 | 5 | 4.20 | 3.35 | 1.06 |
| Awaroa Head | CL2 | Control | 10 | 5-7 m | Boulder | 0 | 3 | 2 | 0 | 0 | 5 | 0 | 3 | 1 | 1 | 1.00 | 1.41 | 0.45 |
| Bark Bay Reef | CL3 | Control | 10 | 5-7 m | Boulder | 0 | 1 | 2 | 0 | 0 | 3 | 1 | 0 | 2 | 0 | 0.60 | 0.89 | 0.28 |
| Totara Rock | CL4 | Control | 10 | 5-6 m | Boulder | 2 | 1 | 3 | 1 | 3 | 10 | 1 | 3 | 4 | 2 | 2.00 | 1.00 | 0.32 |
| Pitt Head | CL5 | Control | 10 | 5-6 m | Boulder | 0 | 0 | 2 | 0 | 2 | 4 | 2 | 2 | 0 | 0 | 0.80 | 1.10 | 0.35 |
| Shallow | | | | | | Number | | | | | Density (100m²) | | | SD | SE | | | |
| ALL RESERVE | | | | | | 117 | | | | | 4.88 | | | 4.48 | 0.90 | | | |
| ALL CONTROL | | | | | | 43 | | | | | 1.72 | | | 2.13 | 0.43 | | | |

Appendix 11. Lobster sizes and density from deep quadrats in April 2006.

| Site name | Site no. | Treatment | N quadrats | Depth (m) | Habitat | Number per quadrat | | | | | Total n | Sex (includes extras outside quadrats) | | | Density (100m ²) | SD | SE |
|------------------|----------|-----------|------------|-----------|---------|--------------------|----|----|----|----|---------|--|--------|----------|------------------------------|------|------|
| | | | | | | 1 | 2 | 3 | 4 | 5 | | Male | Female | Juvenile | | | |
| DEEP | | | | | | | | | | | | | | | | | |
| Cottage Loaf | RL1 | Reserve | 10 | 5-7 m | Boulder | 1 | 2 | 4 | 2 | 10 | 8 | 2 | 0 | 0 | 2.00 | 1.22 | 0.39 |
| Tonga Island | RL2 | Reserve | 10 | 5-7 m | Boulder | 10 | 3 | 6 | 15 | 37 | 16 | 17 | 2 | 2 | 7.40 | 5.13 | 1.62 |
| Foul Point | RL3 | Reserve | 10 | 5-6 m | Boulder | 4 | 8 | 11 | 9 | 40 | 20 | 12 | 5 | 3 | 8.00 | 2.55 | 0.81 |
| Whale Rock | RL4 | Reserve | 10 | 5-7 m | Boulder | 10 | 4 | 8 | 3 | 36 | 23 | 11 | 0 | 2 | 7.20 | 3.56 | 1.13 |
| Mosquito Bay | RL5 | Reserve | 10 | 7-6 m | Boulder | 10 | 12 | 18 | 3 | 43 | 6 | 33 | 0 | 4 | 8.60 | 7.20 | 2.28 |
| Separation Point | CL1 | Control | 10 | 5-6 m | Boulder | 6 | 8 | 8 | 0 | 35 | 11 | 7 | 12 | 5 | 7.00 | 4.69 | 1.48 |
| Awaroa Head | CL2 | Control | 10 | 5-7 m | Boulder | 0 | 0 | 1 | 1 | 3 | 2 | 0 | 1 | 0 | 0.60 | 0.55 | 0.17 |
| Bark Bay Reef | CL3 | Control | 10 | 5-7 m | Boulder | 1 | 3 | 1 | 0 | 5 | 2 | 1 | 2 | 0 | 1.00 | 1.22 | 0.39 |
| Totara Rock | CL4 | Control | 10 | 5-6 m | Boulder | 0 | 0 | 2 | 4 | 7 | 2 | 3 | 2 | 0 | 1.40 | 1.67 | 0.53 |
| Pitt Head | CL5 | Control | 10 | 5-6 m | Boulder | 1 | 0 | 3 | 2 | 6 | 3 | 1 | 1 | 1 | 1.20 | 1.30 | 0.41 |

| Deep | Number | Density (100m ²) | SD | SE |
|-------------|--------|------------------------------|------|------|
| ALL RESERVE | 166 | 6.64 | 4.72 | 0.94 |
| ALL CONTROL | 56 | 2.24 | 3.27 | 0.65 |

Appendix 12. Lobster sizes and density from shallow quadrats in February 2007.

| Site name | Site no. | Treatment | N quadrats | Depth (m) | Habitat | Number per quadrat | | | | | Total n | Sex (includes extras outside quadrats) | | | Density (100m ²) | SD | SE | |
|------------------|----------|-----------|------------|-----------|---------|--------------------|---|----|----|----|---------|--|--------|----------|------------------------------|-------|------|---------|
| | | | | | | 1 | 2 | 3 | 4 | 5 | | Male | Female | Juvenile | | | | Unknown |
| SHALLOW | | | | | | | | | | | | | | | | | | |
| Cottage Loaf | RL1 | Reserve | 10 | 5-7 m | Boulder | 4 | 2 | 2 | 1 | 0 | 9 | 3 | 5 | 1 | 0 | 1.80 | 1.48 | 0.47 |
| Tonga Island | RL2 | Reserve | 10 | 5-7 m | Boulder | 4 | 7 | 1 | 3 | 3 | 18 | 9 | 7 | 1 | 1 | 3.60 | 2.19 | 0.69 |
| Foul Point | RL3 | Reserve | 10 | 5-6 m | Boulder | 3 | 2 | 0 | 0 | 4 | 9 | 6 | 2 | 3 | 1 | 1.80 | 1.79 | 0.57 |
| Whale Rock | RL4 | Reserve | 10 | 5-7 m | Boulder | 1 | 0 | 0 | 1 | 1 | 3 | 1 | 1 | 1 | 0 | 0.60 | 0.55 | 0.17 |
| Mosquito Bay | RL5 | Reserve | 10 | 7-6 m | Boulder | 4 | 6 | 10 | 14 | 16 | 50 | 13 | 18 | 12 | 8 | 10.00 | 5.10 | 1.61 |
| Separation Point | CL1 | Control | 10 | 5-6 m | Boulder | 3 | 1 | 0 | 17 | 6 | 27 | 6 | 2 | 7 | 12 | 5.40 | 6.88 | 2.17 |
| Awaroa Head | CL2 | Control | 10 | 5-7 m | Boulder | 1 | 0 | 0 | 1 | 1 | 3 | 1 | 0 | 0 | 2 | 0.60 | 0.55 | 0.17 |
| Bark Bay Reef | CL3 | Control | 10 | 5-7 m | Boulder | 0 | 0 | 0 | 0 | 4 | 4 | 1 | 0 | 1 | 2 | 0.80 | 1.79 | 0.57 |
| Totara Rock | CL4 | Control | 10 | 5-6 m | Boulder | 4 | 2 | 0 | 2 | 1 | 9 | 6 | 1 | 1 | 1 | 1.80 | 1.48 | 0.47 |
| Pitt Head | CL5 | Control | 10 | 5-6 m | Boulder | 0 | 4 | 0 | 0 | 0 | 4 | 0 | 2 | 0 | 2 | 0.80 | 1.79 | 0.57 |

| Shallow | | Number | Density (100m ²) | SD | SE |
|---------|---------|--------|------------------------------|------|------|
| ALL | RESERVE | 89 | 3.56 | 4.22 | 0.84 |
| ALL | CONTROL | 47 | 1.88 | 3.57 | 0.71 |

Appendix 13. Lobster sizes and density from deep quadrats in February 2007.

| Site name | Site no. | Treatment | N quadrats | Depth (m) | Habitat | Number per quadrat | | | | | Total n | Sex (includes extras outside quadrats) | | | Density (100m ²) | SD | SE | |
|------------------|----------|-----------|------------|-----------|---------|--------------------|----|----|----|----|---------|--|--------|----------|------------------------------|-------|------|---------|
| | | | | | | 1 | 2 | 3 | 4 | 5 | | Male | Female | Juvenile | | | | Unknown |
| DEEP | | | | | | | | | | | | | | | | | | |
| Cottage Loaf | RL1 | Reserve | 10 | 5-7 m | Boulder | 4 | 2 | 3 | 4 | 0 | 13 | 9 | 2 | 2 | 0 | 2.60 | 1.67 | 0.53 |
| Tonga Island | RL2 | Reserve | 10 | 5-7 m | Boulder | 0 | 23 | 15 | 3 | 13 | 54 | 25 | 26 | 3 | 0 | 10.80 | 9.34 | 2.95 |
| Foul Point | RL3 | Reserve | 10 | 5-6 m | Boulder | 0 | 0 | 2 | 16 | 0 | 18 | 5 | 3 | 9 | 1 | 3.60 | 6.99 | 2.21 |
| Whale Rock | RL4 | Reserve | 10 | 5-7 m | Boulder | 3 | 2 | 3 | 4 | 12 | 24 | 14 | 8 | 1 | 1 | 4.80 | 4.09 | 1.29 |
| Mosquito Bay | RL5 | Reserve | 10 | 7-6 m | Boulder | 5 | 23 | 11 | 12 | 23 | 74 | 35 | 30 | 8 | 1 | 14.80 | 7.95 | 2.51 |
| Separation Point | CL1 | Control | 10 | 5-6 m | Boulder | 0 | 2 | 10 | 9 | 0 | 21 | 9 | 3 | 7 | 2 | 4.20 | 4.92 | 1.56 |
| Awaroa Head | CL2 | Control | 10 | 5-7 m | Boulder | 0 | 0 | 2 | 0 | 3 | 5 | 2 | 2 | 0 | 1 | 1.00 | 1.41 | 0.45 |
| Bark Bay Reef | CL3 | Control | 10 | 5-7 m | Boulder | 0 | 0 | 1 | 1 | 0 | 2 | 1 | 0 | 1 | 0 | 0.40 | 0.55 | 0.17 |
| Totara Rock | CL4 | Control | 10 | 5-6 m | Boulder | 2 | 1 | 2 | 1 | 3 | 9 | 4 | 1 | 3 | 1 | 1.80 | 0.84 | 0.26 |
| Pitt Head | CL5 | Control | 10 | 5-6 m | Boulder | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0.20 | 0.45 | 0.14 |

| Deep | Number | Density (100m ²) | SD | SE |
|-------------|--------|------------------------------|------|------|
| ALL RESERVE | 183 | 7.32 | 7.71 | 1.54 |
| ALL CONTROL | 38 | 1.52 | 2.60 | 0.52 |

Appendix 14 Shore profile data for ATNP, March 2001.

| SITE | 1 | | | 2 | | | 3 | | | |
|------|--------------|-----------|-----------------|-----------------------------|-----------|----------------|------------------------------|-----------|-----------------|---------|
| | Distance (m) | Depth (m) | Substratum | % cover algae | Depth (m) | Substratum | % cover algae | Depth (m) | Substratum | % algae |
| 0 | | 0 | bedrock | 0 | 0 | | 2% <i>C. maschalocarpum</i> | 0 | bedrock | 0 |
| 5 | | 0.1 | bedrock | 0 | 0.2 | boulder | 3% <i>C. maschalocarpum</i> | 1.4 | bedrock | 0 |
| 10 | | 1.6 | bedrock | 2% <i>C. maschalocarpum</i> | 0.3 | boulder | 1% <i>C. maschalocarpum</i> | 3.9 | bedrock/sand | 0 |
| 15 | | 3.5 | boulder | 0 | 1.4 | boulder | 0 | 3.7 | boulder/cobble | 0 |
| 20 | | 3.9 | boulder | 0 | 1.4 | boulder | 5% <i>C. flexuosum</i> | 3.7 | bedrock/cobble | 0 |
| 25 | | 4.9 | boulder/sand | 0 | 2.1 | boulder | 0 | 4 | bedrock/sand | 0 |
| 30 | | 4.8 | boulder/cobble | 0 | 2.8 | boulder | 0 | 4.1 | bedrock/sand | 0 |
| 35 | | 5.4 | boulder | 0 | 3.2 | boulder | 0 | 4.5 | bedrock/sand | 0 |
| 40 | | 5.5 | boulder/bedrock | 0 | 3.6 | boulder/gravel | 0 | 4.7 | bedrock/sand | 0 |
| 45 | | 5.8 | boulder/bedrock | 0 | 4 | boulder/cobble | 0 | 4.7 | bedrock/sand | 0 |
| 50 | | 6.8 | boulder/bedrock | 0 | 4.2 | boulder/cobble | 0 | 5 | bedrock/sand | 0 |
| 55 | | 7.9 | bedrock/sand | 0 | 4.2 | boulder/cobble | 0 | 5.3 | bedrock/sand | 0 |
| 60 | | 8.2 | boulder/sand | 0 | 4.6 | boulder/cobble | 1% <i>C. maschalocarpum</i> | 5.3 | bedrock/boulder | 0 |
| 65 | | 8.5 | boulder | 0 | 5.3 | boulder/cobble | 10% <i>C. maschalocarpum</i> | 6.3 | bedrock/boulder | 0 |
| 70 | | 9.7 | boulder/sand | 0 | 6.6 | boulder | 5% <i>C. maschalocarpum</i> | 6.6 | bedrock/boulder | 0 |
| 75 | | 10.6 | boulder/sand | 0 | 8.5 | boulder | 0 | 7.3 | boulder/sand | 0 |
| 80 | | 11.6 | boulder | 0 | 10.8 | boulder | 0 | 7.5 | boulder/sand | 0 |
| 85 | | 12.9 | boulder/sand | 0 | 12.2 | sand/shell | 0 | 8 | sand/cobble | 0 |
| 90 | | 14.6 | boulder/sand | 0 | 12.6 | sand/shell | 0 | 8.3 | sand/shell | 0 |
| 95 | | 15.7 | shell/sand | 0 | 12.7 | sand/shell | 0 | 8.3 | sand/shell | 0 |
| 100 | | 15.8 | shell/sand | 0 | 12.7 | sand/shell | 0 | 8.5 | sand/shell | 0 |
| 105 | | | | | 12.9 | sand/shell | | | | |
| 110 | | | | | 13.1 | sand/shell | | | | |
| 115 | | | | | 13.3 | sand/shell | | | | |
| 120 | | | | | 13.6 | sand/shell | | | | |
| 125 | | | | | 13.9 | sand/shell | | | | |
| 130 | | | | | 14.2 | sand/shell | | | | |
| 135 | | | | | 14.5 | sand/shell | | | | |
| 140 | | | | | 14.7 | sand/shell | | | | |
| 145 | | | | | 14.9 | sand/shell | | | | |
| 150 | | | | | 15.2 | sand/shell | | | | |

Appendix 14 (continued). Shore profile data for ATNP, March 2001.

| 4 | | 4 | | 5 | | 5 | | 6 | | 6 | | | |
|-----------|--------------------|-------------------------|-----------|-----------------|---|-----------|-----------------|-----------|-----------------|------------------------------|-----------|-----------------|------------------------------|
| Depth (m) | Substratum | % cover algae | Depth (m) | Substratum | % cover algae | Depth (m) | Substratum | Depth (m) | Substratum | % cover algae | Depth (m) | Substratum | % cover algae |
| 0 | bedrock/boulder | 50% <i>C. flexuosum</i> | 0 | boulder/gravel | 5% <i>Cystophora</i> sp. 5% <i>C. flexuosum</i> | 0 | boulder | 0 | boulder | 10% <i>C. maschalocarpum</i> | 0 | boulder | 10% <i>C. maschalocarpum</i> |
| 2.8 | bedrock/boulder | 60% <i>C. flexuosum</i> | 0.7 | boulder/sand | 2% <i>C. flexuosum</i> | 1.3 | boulder/sand | 1.3 | boulder/cobble | 5% <i>C. flexuosum</i> | 1.3 | boulder/cobble | 5% <i>C. flexuosum</i> |
| 5 | boulder/sand | 40% <i>C. flexuosum</i> | 1.3 | boulder/sand | 1% <i>C. flexuosum</i> | 1.8 | boulder/sand | 2.3 | boulder/bedrock | 5% <i>C. flexuosum</i> | 2.3 | boulder/bedrock | 5% <i>C. flexuosum</i> |
| 4.5 | boulder/sand | 40% <i>C. flexuosum</i> | 1.8 | boulder/sand | 0.5% <i>C. flexuosum</i> | 2 | boulder/sand | 3.4 | boulder | 0 | 3.4 | boulder | 0 |
| 5 | boulder/silt | 30% <i>C. flexuosum</i> | 2 | boulder/sand | 0 | 2.3 | sand/shell | 3.7 | boulder | 0 | 3.7 | boulder | 0 |
| 5.4 | boulder/silt/shell | 0 | 2.3 | sand/shell | 0 | 2.4 | boulder/shell | 5.7 | boulder | 0 | 5.7 | boulder | 0 |
| 5.1 | boulder/silt/shell | 0 | 2.4 | boulder/shell | 0 | 0.2 | boulder | 7.9 | boulder/sand | 0 | 7.9 | boulder/sand | 0 |
| 6.5 | boulder/silt/shell | 0 | 0.2 | boulder | 0 | 1.6 | boulder/sand | 8.5 | sand/shell | 0 | 8.5 | sand/shell | 0 |
| 6.7 | boulder/silt/shell | 0 | 1.6 | boulder/sand | 0 | 2.5 | bedrock | 8.8 | sand/shell | 0 | 8.8 | sand/shell | 0 |
| 7 | boulder/sand/shell | 0 | 2.5 | bedrock | 0 | 4.4 | bedrock/boulder | 9.3 | sand/boulder | 0 | 9.3 | sand/boulder | 0 |
| 7.2 | boulder/sand/shell | 0 | 4.4 | bedrock/boulder | 0 | 6 | gravel/sand | 8.2 | boulder | 0 | 8.2 | boulder | 0 |
| 8.2 | boulder/sand/shell | 0 | 6 | gravel/sand | 0 | 6.7 | cobble/sand | 10.9 | sand/shell | 0 | 10.9 | sand/shell | 0 |
| 8.2 | sand/shell | 0 | 6.7 | cobble/sand | 0 | 7.5 | sand/silt | 11.4 | sand/shell | 0 | 11.4 | sand/shell | 0 |
| 8.2 | sand/shell | 0 | 7.5 | sand/silt | 0 | 7.8 | sand/shell | 11.8 | silt/shell | 0 | 11.8 | silt/shell | 0 |
| 8.3 | sand/shell | 0 | 7.8 | sand/shell | 0 | 8 | sand/shell | 12.1 | silt/shell | 0 | 12.1 | silt/shell | 0 |
| 8.4 | sand/shell | 0 | 8 | sand/shell | 0 | 8.1 | sand/shell | 12.3 | silt/shell | 0 | 12.3 | silt/shell | 0 |
| 8.5 | sand/shell | 0 | 8.1 | sand/shell | 0 | 8.4 | sand/shell | 12.7 | silt/shell | 0 | 12.7 | silt/shell | 0 |
| 8.5 | sand/shell | 0 | 8.4 | sand/shell | 0 | 8.6 | sand/shell | 13.1 | silt/shell | 0 | 13.1 | silt/shell | 0 |
| 8.6 | sand/shell | 0 | 8.6 | sand/shell | 0 | 8.8 | sand/shell | 13.7 | shell/sand | 0 | 13.7 | shell/sand | 0 |
| 8.8 | sand/shell | 0 | 8.8 | sand/shell | 0 | 9 | sand/shell | 14.1 | shell/sand | 0 | 14.1 | shell/sand | 0 |
| 8.9 | sand/shell | 0 | 9 | sand/shell | 0 | 9.2 | sand/shell | 14.3 | shell/sand | 0 | 14.3 | shell/sand | 0 |
| 9 | sand/shell | 0 | 9.2 | sand/shell | 0 | 9.5 | sand/shell | 15.3 | shell/sand | 0 | 15.3 | shell/sand | 0 |
| 9 | sand/shell | 0 | 9.5 | sand/shell | 0 | 9.6 | sand/shell | 15.7 | shell/sand | 0 | 15.7 | shell/sand | 0 |
| 9 | sand/shell | 0 | 9.6 | sand/shell | 0 | 9.8 | sand/shell | 15.9 | shell/sand | 0 | 15.9 | shell/sand | 0 |
| 9.1 | sand/shell | 0 | 9.8 | sand/shell | 0 | 10 | shell/silt | 16.2 | shell/sand | 0 | 16.2 | shell/sand | 0 |
| 9.1 | sand/shell | 0 | 10 | shell/silt | 0 | 10.2 | shell/silt | 17.1 | shell/sand | 0 | 17.1 | shell/sand | 0 |
| 9.1 | sand/shell | 0 | 10.2 | shell/silt | 0 | 10.3 | shell/silt | 17.3 | shell/sand | 0 | 17.3 | shell/sand | 0 |
| 9.2 | sand/shell | 0 | 10.3 | shell/silt | 0 | 10.4 | shell/silt | 17.5 | shell/sand | 0 | 17.5 | shell/sand | 0 |
| 9.2 | sand/shell | 0 | 10.4 | shell/silt | 0 | 10.5 | shell/silt | 17.7 | shell/sand | 0 | 17.7 | shell/sand | 0 |
| 9.2 | sand/shell | 0 | 10.5 | shell/silt | 0 | 10.6 | shell/silt | 17.9 | shell/sand | 0 | 17.9 | shell/sand | 0 |
| 9.2 | sand/shell | 0 | 10.6 | shell/silt | 0 | | | 18 | shell/sand | 0 | 18 | shell/sand | 0 |

Appendix 15. Raw benthic quadrat data collected in March 2001 from the ATNP coastline.

| SITES | 1 | | | | | | | | 2 | | | | | | | |
|--|----|----|----|---|----|----|----|---|----|-----|----|----|----|----|----|---|
| Species | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| <i>Aptos aptos</i> | | | | | | | | | | | | | | | | |
| <i>Ancorina alata</i> | | | | | | 2 | | 1 | | | | | | | | |
| <i>Aplysilla sulphurea</i> (%) | | | | | | | | | | | | 1 | | | | |
| <i>Crella incrustans</i> (%) | | | | | | | | | | | | | | | | |
| <i>Callyspongia</i> sp | | | | | | | | | | | | | | | | |
| <i>Chelonaplysilla violacea</i> (%) | | | 1 | 1 | | | | | | | | | | | | |
| <i>Celleporaria agglutinans</i> (%) | | | | | | | | | | | | | | | | |
| <i>Watersipora cucullata</i> (%) | | | | 2 | 2 | | | | | | | | | | | |
| <i>Actinothoe albocincta</i> | | | | | | | | | | | | | | | | |
| <i>Alcyonium aurantiacum</i> | | | | | | | | 2 | | | | | | | | |
| <i>Corynactis haddonni</i> | | | | | | | | | | | | | | | | |
| <i>Culicia rubeola</i> | | | | | 35 | | 55 | 3 | | | | 15 | | | 15 | |
| <i>Magasella sanguinea</i> | | | | | | | | | | | | | | | | |
| <i>Cantharadus purpureus</i> | | | | | | | | | | | | | | | | |
| <i>Cellana</i> spp. | | | | | | | | 1 | 16 | 115 | 12 | | 2 | 3 | 12 | 4 |
| <i>Cheidothaeus albidus</i> | | 1 | 2 | | | | | | | | | | | | | |
| <i>Chitons</i> spp. | | | | | | | | | | | | | | | | |
| <i>Cookia sulcata</i> | | | 2 | 2 | | 3 | 1 | | | 1 | 3 | 2 | 1 | 4 | 4 | |
| <i>Chromodoris amoena</i> | | | | | | | | | | | | | | | | |
| <i>Hauistorum hauitorium</i> | | | | | | | | | | | | | | | | |
| <i>Haliotis iris</i> | | | | | | | | | | | | | | | | |
| <i>Maoricolpus roseus</i> | | | | | | | | | | | | | 1 | | | |
| <i>Mauria</i> spp. | | | | | | | | 1 | | | | | | | | |
| <i>Modiolarca impacta</i> | | | | | | | | | | | | | | | | |
| <i>Monia zelandica</i> | | | | | | 2 | 3 | | | 1 | | 4 | | | 2 | 1 |
| <i>Pecten novaezelandiae</i> | | | | | | | | | | | | | | | | |
| <i>Trochus</i> spp. | | 1 | 5 | 9 | | 2 | | 2 | | 1 | 1 | | 2 | | 2 | 8 |
| Unidentified whelk | 1 | 1 | | 2 | 1 | | | | 1 | | | | | | | |
| <i>Turbo granosus</i> | | | | | | | | | | | | | | | | |
| <i>Turbo smaragdus</i> | 2 | | | | | | | | 3 | 7 | | 11 | 12 | 11 | 3 | 7 |
| <i>Galeolaria hystrix</i> | | | | | | | | | | 1 | 6 | | | 1 | 3 | |
| <i>Allostichaster insignis</i> | | | | | | | | | | | | | | | | |
| <i>Coscinasterias calamaria</i> | | | | | | | | | | | | | | | | |
| <i>Evechinus chloroticus</i> | 4 | | | 1 | | | | 2 | 4 | 5 | | 1 | 6 | 6 | 2 | 4 |
| <i>Patiriella regularis</i> | | | | | | | | | 1 | 1 | | 1 | 1 | 1 | | |
| <i>Pectinura maculata</i> | | | | | | | | | | | | | | | | |
| <i>Pentagonaster pulchellus</i> | | | | | | | | | | | | | | | | |
| <i>Pseudechinus albocinctus</i> | | | | | | | | | | | | | | | | |
| <i>Stichaster australis</i> | | | | | | | | | | | | | | | | |
| <i>Stichopus mollis</i> | | | | | | | | | | | | | 1 | | | |
| <i>Aplidium</i> sp. (%) | | | | | | | | | | | | | | | | |
| <i>Aplidium benham</i> | | | | | | | | | | | | | | | | |
| <i>Leptoclinides</i> sp. (%) | | | | | | | | | | | | | | | | |
| <i>Cnemidocarpa bicornuata</i> | 3 | 1 | | | | 1 | 2 | | | | | | | | | |
| Unidentified warty squirt | 1 | | | 1 | 1 | 4 | 3 | 5 | | | | | | | | |
| <i>Didemnum</i> sp. (%) | | | 1 | 1 | 1 | 1 | | 1 | | | | | | | | |
| <i>Carpophyllum flexuosum</i> (%) | 1 | 1 | 2 | 1 | 5 | 2 | 2 | 1 | | | | | | | | |
| <i>Carpophyllum flexuosum</i> (stipes) | 45 | 47 | 7 | 1 | 12 | 13 | 22 | 4 | | | | | | | | |
| <i>Corallina</i> (%) | 3 | 25 | 35 | 4 | 35 | 15 | 15 | 2 | 25 | 75 | 7 | 75 | 8 | 6 | 85 | 8 |

Appendix 15 (continued). Raw benthic quadrat data collected in March 2001 from the ATNP.

| 6 | | | | | | | | | | | 7 | | | | | | | | 8 | | | | | | | |
|----|----|----|----|----|----|----|----|----|----|---|---|----|----|---|---|----|----|---|----|---|----|----|----|----|----|---|
| 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 1 | 1 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| | | | | | | | | | 1 | | | | | | | | | 1 | | | | | | | | |
| | | 1 | | | 1 | 1 | | 1 | 1 | | | | | | | | | | | | | | | | | 1 |
| | | | | | | | | | | | 1 | 5 | 1 | 1 | | 1 | | 2 | 1 | | | | | | | 1 |
| | | | | | 35 | | | | | | 5 | 45 | 13 | 2 | 4 | 1 | 45 | 3 | 15 | 8 | 35 | 3 | 45 | 5 | 25 | |
| | | | 1 | | | | | | | | | | | | | | | 1 | 2 | 2 | | | 2 | 1 | 1 | |
| 2 | 1 | 3 | 4 | 1 | 1 | 3 | 2 | | 2 | | 1 | 2 | 2 | | | | | 1 | 1 | 1 | | | 3 | 2 | | |
| | | | | | | | | | | | | | | | 1 | | | | | 2 | | | | | | |
| | | 1 | | | 3 | | | 1 | | | 2 | 1 | | | 7 | | 1 | 2 | 8 | 2 | | | 2 | 1 | 2 | 1 |
| 1 | 2 | 18 | 4 | 2 | 3 | 6 | 5 | 5 | 7 | | 1 | | 4 | | | | | | | 1 | 2 | 3 | 1 | | | |
| | | | | | | | | | | | | | 1 | 1 | | | | | | 1 | 3 | | | 18 | 3 | |
| | | 1 | 1 | | | 2 | | 2 | 3 | 1 | | | 2 | | | 1 | 1 | | 7 | 1 | 5 | 4 | 4 | 4 | 1 | 5 |
| | | | | | | | | | | | | | | | | | | | | 3 | | | | | | |
| | | | 1 | | | | | | | | | | 1 | | | | | | 2 | | 1 | 3 | | | | 1 |
| | | | | | | 3 | 1 | | 1 | | 1 | 1 | | 5 | 5 | 1 | | 2 | 1 | 3 | | | 3 | | | 1 |
| | | | | | | | | | | | | | | | | | | | | | | | | | | |
| | | 1 | | | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | | | | | | | |
| | | | | | | | 1 | | | | 1 | | | | | | | | | 2 | | | 1 | | | 1 |
| | | | | | | | | | | | 1 | | | | | | | | | | | | | | | 1 |
| | | | | | 2 | 1 | | | | | 3 | 4 | | | 1 | 5 | | 3 | | | | | 1 | 1 | 3 | 1 |
| 8 | 6 | 35 | 95 | 6 | 1 | 1 | 1 | 5 | 2 | | | | | | | 1 | 1 | | 1 | | | | | | | |
| 56 | 31 | 38 | 48 | 38 | 35 | 27 | 32 | 15 | 13 | | | | | | 2 | 3 | | | 2 | 1 | 5 | 1 | 1 | 15 | 4 | 1 |
| 7 | 9 | 8 | 9 | 92 | 95 | 9 | 8 | 9 | 9 | | 1 | 3 | 55 | 8 | 8 | 85 | 85 | 9 | 15 | 4 | 24 | 1 | 1 | 35 | 4 | 6 |
| | | | | | | | | | | | | | | | | | | | 7 | 8 | 5 | 45 | 8 | 55 | 58 | 6 |

Appendix 15 (continued). Raw benthic quadrat data collected in March 2001 from the ATNP.

| SITES | 9 | | | | | | | | Mean | SD |
|--|----|----|----|----|---|----|---|----|-------|-------|
| | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | | |
| Species | | | | | | | | | | |
| <i>Aptos aptos</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Ancorina alata</i> | | | | | | | | | 1.33 | 0.58 |
| <i>Aplysilla sulphurea</i> (%) | | | | | | | | | 0.00 | 0.00 |
| <i>Crella incrustans</i> (%) | 1 | | | | | | | | 1.40 | 0.55 |
| <i>Callyspongia</i> sp | | | | | | | | | 0.00 | 0.00 |
| <i>Chelonaplysilla violacea</i> (%) | | | | | 1 | | | | 1.17 | 0.38 |
| <i>Celleporaria agglutinans</i> (%) | | | | | | | | | 0.00 | 0.00 |
| <i>Watersipora cucullata</i> (%) | | | 1 | | | | | | 1.47 | 1.06 |
| <i>Actinothoe albocincta</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Alcyonium aurantiacum</i> | | | | | | | | | 2.00 | 0.00 |
| <i>Corynactis haddonni</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Culicia rubeola</i> | | | 25 | | 1 | 35 | 3 | | 15.33 | 16.42 |
| <i>Magasella sanguinea</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Cantharadus purpureus</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Cellana</i> spp. | 6 | 11 | 3 | 5 | 9 | 5 | 4 | 19 | 7.13 | 18.49 |
| <i>Cheidothaerus albidus</i> | | | | | | | | | 1.50 | 0.71 |
| <i>Chitons</i> spp. | | | | | | | | 1 | 1.00 | 0.00 |
| <i>Cookia sulcata</i> | | | | | | 1 | | 2 | 2.03 | 1.27 |
| <i>Chromodoris amoena</i> | | | | | | | | | 1.00 | 0.00 |
| <i>Haustorium haustorium</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Haliotis iris</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Maoricolpus roseus</i> | | | | 12 | | | | | 4.00 | 4.07 |
| <i>Mauria</i> spp. | | | | | | | | | 1.00 | 0.00 |
| <i>Modiolarca impacta</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Monia zelandica</i> | 11 | 3 | 4 | | 1 | 3 | 3 | 7 | 3.51 | 4.04 |
| <i>Pecten novaezelandiae</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Trochus</i> spp. | | | | 1 | 2 | 4 | | 2 | 5.30 | 5.50 |
| Unidentified whelk | | | | | | | | 1 | 2.59 | 4.06 |
| <i>Turbo granosus</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Turbo smaragdus</i> | 5 | 8 | 5 | 8 | 7 | 7 | 5 | 7 | 4.64 | 3.17 |
| <i>Galeolaria hystrix</i> | | 1 | 1 | 2 | | 2 | 3 | | 2.59 | 2.04 |
| <i>Allostichaster insignis</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Coscinasterias calamaria</i> | | 1 | 2 | | | | | | 1.33 | 0.58 |
| <i>Evechinus chloroticus</i> | | | 2 | 1 | 3 | 3 | | 1 | 2.28 | 1.56 |
| <i>Patiriella regularis</i> | 1 | | 1 | | | 1 | 1 | | 1.50 | 0.90 |
| <i>Pectinura maculata</i> | 1 | | | | | | | | 1.00 | 0.00 |
| <i>Pentagonaster pulchellus</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Pseudechinus albocinctus</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Stichaster australis</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Stichopus mollis</i> | | | 1 | | | | | | 1.00 | 0.00 |
| <i>Aplidium</i> sp. (%) | | | | | | | | 1 | 1.00 | 0.00 |
| <i>Aplidium benham</i> | | | | | | | | | 0.00 | 0.00 |
| <i>Leptoclinides</i> sp. (%) | | | | | | | | | 1.14 | 0.38 |
| <i>Cnemidocarpa bicornuata</i> | | | | | | | | | 1.45 | 0.69 |
| Unidentified warty squirt | 1 | | | | | | | | 2.38 | 1.54 |
| <i>Didemnum</i> sp. (%) | | | | | | | | | 1.10 | 0.31 |
| <i>Carpophyllum flexuosum</i> (%) | | | | | | | | | 8.68 | 20.50 |
| <i>Carpophyllum flexuosum</i> (stipes) | | | | | | | | | 21.12 | 17.41 |
| <i>Corallina</i> (%) | 6 | 65 | 45 | 3 | 3 | 4 | 6 | 65 | 36.12 | 36.63 |

Appendix 16. Kina (*E. chloroticus*) densities collected from 1m² quadrats (April 2002). 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Depth | 1 | Depth | 2 | Depth | 3 | Depth | 4 | Depth | 5 | Depth | 6 | Depth | 7 |
|-------|---|-------|---|-------|---|-------|----|-------|---|-------|---|-------|---|
| 6 | 2 | 7 | 2 | 5.5 | 0 | 6 | 3 | 6.5 | 0 | 6 | 5 | 4.5 | 5 |
| 6 | 1 | 7 | 0 | 7 | 1 | 6.5 | 2 | 6.5 | 1 | 6.5 | 4 | 4.5 | 1 |
| 6 | 4 | 7.5 | 3 | 9 | 2 | 6 | 3 | 7.5 | 1 | 8.5 | 0 | 4.5 | 2 |
| 5.5 | 0 | 7.5 | 0 | 10 | 2 | 6 | 10 | 6.5 | 1 | 8.5 | 1 | 5 | 2 |
| 5.5 | 0 | 8 | 2 | 10 | 1 | 6 | 1 | 7 | 0 | 8 | 2 | 5 | 0 |
| 6 | 5 | 8 | 8 | 10 | 0 | 5.5 | 6 | 7.5 | 0 | 9 | 2 | 5 | 0 |
| 6.5 | 5 | 7.5 | 0 | 10 | 1 | 6 | 2 | 7.5 | 2 | 9.5 | 1 | 5 | 0 |
| 6.5 | 2 | 7 | 2 | 7.5 | 1 | 6.5 | 2 | 7 | 2 | 10 | 3 | 5 | 0 |
| 7 | 2 | 7 | 2 | 7 | 0 | 6 | 6 | 8.5 | 1 | 10 | 4 | 5 | 3 |
| 7.5 | 0 | 9 | 1 | 6.5 | 1 | 6 | 0 | 8.5 | 1 | 9.5 | 0 | 5 | 2 |
| 9.5 | 1 | 9 | 0 | 6.5 | 0 | 6 | 1 | 8 | 3 | 8.5 | 1 | 5 | 2 |
| 10 | 0 | 9 | 1 | 6 | 0 | 6 | 2 | 8.5 | 0 | 9 | 1 | 4 | 3 |
| 10 | 0 | 9 | 0 | 6 | 0 | 5.5 | 2 | 9 | 4 | 8 | 0 | 4 | 1 |
| 10 | 0 | 11 | 0 | 5 | 0 | 5.5 | 0 | 9.5 | 4 | 7 | 0 | 4 | 2 |
| 10 | 2 | 11 | 0 | 5.5 | 0 | 5.5 | 0 | 10 | 1 | 6.5 | 0 | 4 | 3 |
| 10 | 0 | 11 | 1 | 6.5 | 0 | 5 | 1 | 10 | 1 | 6.5 | 0 | 4 | 0 |
| 9.5 | 0 | 12 | 2 | 7 | 0 | 5.5 | 1 | 10 | 2 | 5.5 | 4 | 4 | 8 |
| 9.5 | 0 | 13 | 1 | 12.5 | 0 | 6 | 1 | 10 | 0 | 5 | 3 | 4 | 4 |
| 6 | 0 | 13 | 0 | 11 | 0 | 6 | 2 | 10 | 0 | 5.5 | 3 | 4 | 0 |
| 7.5 | 1 | 12 | 0 | 12 | 0 | 6 | 2 | 7 | 1 | 5 | 0 | 3.5 | 2 |
| 8 | 0 | 11 | 0 | 10.5 | 0 | 6 | 0 | 5.5 | 4 | 6.5 | 3 | 4 | 4 |
| 8 | 0 | 9.5 | 0 | 10.5 | 0 | 6 | 0 | 5 | 1 | 7 | 3 | 4 | 0 |
| 9.5 | 0 | 10 | 1 | 10 | 0 | 6 | 3 | 5.5 | 3 | 7 | 1 | 4 | 1 |
| 9.5 | 0 | 9 | 0 | 10 | 3 | 5.5 | 0 | 6 | 1 | 7.5 | 1 | 5 | 3 |
| 7 | 0 | 10 | 1 | 8 | 0 | 6.5 | 0 | 7.5 | 0 | 8 | 1 | 5 | 2 |
| 5.5 | 0 | 10 | 1 | 5 | 2 | 6.5 | 0 | 8 | 0 | 8 | 1 | 6 | 4 |
| 5 | 5 | 9 | 1 | 6 | 0 | 6.5 | 0 | 8 | 1 | 10 | 1 | 6 | 2 |
| 5 | 2 | 9 | 1 | 6 | 0 | 6.5 | 0 | 8.5 | 2 | 10 | 0 | 6 | 4 |
| 5 | 2 | 9.5 | 0 | 6.5 | 1 | 6 | 0 | 8 | 3 | 8.5 | 0 | 6 | 0 |
| 5 | 0 | 8.2 | 0 | 8 | 2 | 5.5 | 0 | 8 | 2 | 9 | 4 | 6 | 1 |
| 6.9 | 2 | 8 | 1 | 5 | 0 | 5 | 0 | 6.9 | 0 | 9 | 0 | 3.7 | 4 |
| 6.7 | 0 | 8.9 | 3 | 5.5 | 1 | 5 | 1 | 6.5 | 0 | 10 | 1 | 3.6 | 0 |
| 6.5 | 4 | 9.4 | 0 | 6 | 2 | 5 | 0 | 7 | 0 | 9 | 0 | 3.4 | 4 |
| 6.8 | 6 | 9.1 | 0 | 6.1 | 2 | 6 | 2 | 7.1 | 0 | 9 | 1 | 4.3 | 1 |
| 6.9 | 4 | 10.6 | 2 | 5.9 | 2 | 6 | 4 | 7.3 | 4 | 9 | 2 | 4.3 | 4 |
| 6.8 | 1 | 10.2 | 2 | 5.6 | 0 | 6 | 0 | 6.9 | 1 | 9 | 0 | 4.5 | 2 |
| 7 | 1 | 10.2 | 0 | 6.8 | 0 | 6 | 0 | 8.2 | 3 | 9 | 1 | 4.6 | 4 |
| 7 | 0 | 10.9 | 0 | 6.5 | 1 | 6 | 0 | 8.3 | 5 | 9 | 2 | 4.7 | 0 |
| 7.2 | 6 | 10.6 | 0 | 7.9 | 2 | 6 | 4 | 9.9 | 2 | 8 | 0 | 4.2 | 5 |
| 7.4 | 4 | 9.7 | 1 | 8.8 | 0 | 6 | 5 | 9.9 | 1 | 7 | 1 | 4.3 | 2 |
| 7.5 | 2 | 11.2 | 1 | 9.5 | 2 | 6 | 1 | 9.8 | 5 | 7 | 3 | 4.3 | 0 |
| 7.1 | 4 | 12.1 | 1 | 10 | 0 | 5 | 0 | 9.2 | 2 | 7 | 2 | 4.2 | 4 |
| 7.1 | 2 | 12.6 | 0 | 9.8 | 2 | 3 | 0 | 10.1 | 1 | 8 | 1 | 4.5 | 2 |
| 7.3 | 0 | 11.9 | 1 | 9.9 | 3 | 6 | 0 | 10.4 | 1 | 7 | 1 | 4.5 | 6 |
| 7.5 | 0 | 12.8 | 0 | 9.3 | 2 | 5 | 3 | 8.7 | 4 | 7 | 1 | 4.3 | 1 |
| 8.9 | 0 | 13 | 1 | 9.3 | 1 | 5 | 0 | 8.3 | 2 | 6 | 1 | 4.4 | 2 |
| 9.5 | 1 | 11.2 | 0 | 9.1 | 2 | 5 | 1 | 7.6 | 0 | 6 | 2 | 4.7 | 3 |
| 10.3 | 0 | 11.1 | 1 | 8.5 | 4 | 5 | 1 | 7.5 | 0 | 6 | 2 | 4.6 | 0 |
| 140.2 | 0 | 10.9 | 0 | 7.5 | 1 | 6 | 2 | 7.7 | 0 | 7 | 0 | 5.3 | 3 |
| 10 | 0 | 10.9 | 0 | 7.7 | 0 | 6 | 2 | 7.7 | 4 | 6 | 3 | 5.5 | 8 |
| 9.4 | 0 | 9.9 | 0 | 6.6 | 0 | 6 | 1 | 7 | 1 | 7 | 1 | 5.4 | 2 |
| 8 | 1 | 9.8 | 2 | 5.8 | 0 | 6 | 2 | 6.4 | 0 | 7 | 1 | 5.5 | 2 |
| 7.9 | 0 | 9.8 | 0 | 5.7 | 2 | 5 | 0 | 5.2 | 0 | 5 | 3 | 5.6 | 2 |
| 7 | 0 | 8.9 | 2 | 5.9 | 0 | 5 | 6 | 5 | 1 | 6 | 3 | 5.7 | 1 |
| 6.5 | 3 | 8.3 | 1 | 5.2 | 6 | 5 | 0 | 6.6 | 0 | 5 | 2 | 5.3 | 1 |
| 6.4 | 2 | 7.5 | 0 | 5 | 0 | 5 | 0 | 7.2 | 0 | 5 | 1 | 5.3 | 3 |
| 6.3 | 1 | 7 | 0 | 5 | 1 | 5 | 2 | 7.3 | 1 | 5 | 7 | 5.4 | 2 |
| 5.5 | 0 | 7 | 0 | 5.2 | 0 | 5 | 1 | 7.4 | 2 | 5 | 1 | 5.8 | 0 |
| 6 | 0 | 7 | 0 | 5.6 | 0 | 5 | 1 | 6.9 | 0 | 6 | 5 | 5.4 | 1 |
| 5.9 | 3 | | | 5.6 | 0 | 5 | 1 | 6.1 | 0 | 5 | 4 | 4.6 | 1 |
| | | | | | | 5 | 2 | | | | | | |

Appendix 17. Cook's turban (*Cookia sulcata*) densities collected in April 2002. 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Depth | 1 | Depth | 2 | Depth | 3 | Depth | 4 | Depth | 5 | Depth | 6 | Depth | 7 |
|-------|---|-------|----|-------|----|-------|---|-------|----|-------|---|-------|---|
| 6 | 3 | 7 | 17 | 5.5 | 1 | 6 | 0 | 6.5 | 4 | 6 | 0 | 4.5 | 0 |
| 6 | 1 | 7 | 3 | 7 | 4 | 6.5 | 0 | 6.5 | 0 | 6.5 | 2 | 4.5 | 1 |
| 6 | 2 | 7.5 | 4 | 9 | 6 | 6 | 0 | 7.5 | 1 | 8.5 | 0 | 4.5 | 0 |
| 5.5 | 0 | 7.5 | 8 | 10 | 3 | 6 | 0 | 6.5 | 0 | 8.5 | 0 | 5 | 0 |
| 5.5 | 1 | 8 | 1 | 10 | 1 | 6 | 0 | 7 | 1 | 8 | 0 | 5 | 4 |
| 6 | 1 | 8 | 1 | 10 | 0 | 5.5 | 1 | 7.5 | 1 | 9 | 2 | 5 | 0 |
| 6.5 | 3 | 7.5 | 3 | 10 | 0 | 6 | 1 | 7.5 | 1 | 9.5 | 0 | 5 | 0 |
| 6.5 | 3 | 7 | 4 | 7.5 | 2 | 6.5 | 0 | 7 | 0 | 10 | 1 | 5 | 2 |
| 7 | 4 | 7 | 4 | 7 | 2 | 6 | 1 | 8.5 | 0 | 10 | 0 | 5 | 0 |
| 7.5 | 1 | 9 | 2 | 6.5 | 4 | 6 | 0 | 8.5 | 9 | 9.5 | 1 | 5 | 2 |
| 9.5 | 2 | 9 | 3 | 6.5 | 3 | 6 | 2 | 8 | 4 | 8.5 | 3 | 5 | 0 |
| 10 | 1 | 9 | 1 | 6 | 10 | 6 | 0 | 8.5 | 2 | 9 | 0 | 4 | 2 |
| 10 | 3 | 9 | 2 | 6 | 7 | 5.5 | 0 | 9 | 0 | 8 | 3 | 4 | 1 |
| 10 | 1 | 11 | 1 | 5 | 9 | 5.5 | 0 | 9.5 | 0 | 7 | 5 | 4 | 0 |
| 10 | 0 | 11 | 1 | 5.5 | 5 | 5.5 | 0 | 10 | 2 | 6.5 | 2 | 4 | 3 |
| 10 | 0 | 11 | 4 | 6.5 | 11 | 5 | 1 | 10 | 0 | 6.5 | 7 | 4 | 0 |
| 9.5 | 1 | 12 | 0 | 7 | 0 | 5.5 | 0 | 10 | 0 | 5.5 | 0 | 4 | 0 |
| 9.5 | 5 | 13 | 0 | 12.5 | 0 | 6 | 0 | 10 | 7 | 5 | 0 | 4 | 4 |
| 6 | 4 | 13 | 0 | 11 | 0 | 6 | 0 | 10 | 1 | 5.5 | 0 | 4 | 0 |
| 7.5 | 3 | 12 | 0 | 12 | 0 | 6 | 0 | 7 | 7 | 5 | 0 | 3.5 | 0 |
| 8 | 3 | 11 | 0 | 10.5 | 1 | 6 | 0 | 5.5 | 3 | 6.5 | 1 | 4 | 0 |
| 8 | 6 | 9.5 | 1 | 10.5 | 2 | 6 | 1 | 5 | 5 | 7 | 8 | 4 | 0 |
| 9.5 | 7 | 10 | 0 | 10 | 0 | 6 | 1 | 5.5 | 2 | 7 | 5 | 4 | 0 |
| 9.5 | 5 | 9 | 1 | 10 | 0 | 5.5 | 2 | 6 | 10 | 7.5 | 2 | 5 | 0 |
| 7 | 0 | 10 | 0 | 8 | 1 | 6.5 | 3 | 7.5 | 4 | 8 | 0 | 5 | 0 |
| 5.5 | 2 | 10 | 1 | 5 | 2 | 6.5 | 2 | 8 | 0 | 8 | 1 | 6 | 0 |
| 5 | 0 | 9 | 0 | 6 | 0 | 6.5 | 0 | 8 | 0 | 10 | 0 | 6 | 0 |
| 5 | 0 | 9 | 0 | 6 | 1 | 6.5 | 2 | 8.5 | 1 | 10 | 0 | 6 | 4 |
| 5 | 0 | 9.5 | 1 | 6.5 | 3 | 6 | 3 | 8 | 0 | 8.5 | 3 | 6 | 0 |
| 5 | 1 | 8.2 | 6 | 8 | 4 | 5.5 | 0 | 8 | 2 | 9 | 0 | 6 | 0 |
| 6.9 | 2 | 8 | 2 | 5 | 1 | 5 | 9 | 6.9 | 1 | 9 | 1 | 3.7 | 0 |
| 6.7 | 0 | 8.9 | 1 | 5.5 | 2 | 5 | 1 | 6.5 | 6 | 10 | 1 | 3.6 | 1 |
| 6.5 | 4 | 9.4 | 1 | 6 | 0 | 5 | 1 | 7 | 2 | 9 | 1 | 3.4 | 0 |
| 6.8 | 6 | 9.1 | 7 | 6.1 | 0 | 6 | 3 | 7.1 | 0 | 9 | 0 | 4.3 | 1 |
| 6.9 | 4 | 10.6 | 0 | 5.9 | 1 | 6 | 0 | 7.3 | 2 | 9 | 0 | 4.3 | 0 |
| 6.8 | 1 | 10.2 | 1 | 5.6 | 7 | 6 | 1 | 6.9 | 0 | 9 | 0 | 4.5 | 0 |
| 7 | 1 | 10.2 | 3 | 6.8 | 1 | 6 | 4 | 8.2 | 0 | 9 | 0 | 4.6 | 0 |
| 7 | 0 | 10.9 | 2 | 6.5 | 1 | 6 | 1 | 8.3 | 0 | 9 | 2 | 4.7 | 0 |
| 7.2 | 6 | 10.6 | 1 | 7.9 | 0 | 6 | 1 | 9.9 | 2 | 8 | 1 | 4.2 | 0 |
| 7.4 | 4 | 9.7 | 6 | 8.8 | 3 | 6 | 0 | 9.9 | 0 | 7 | 3 | 4.3 | 2 |
| 7.5 | 2 | 11.2 | 0 | 9.5 | 2 | 6 | 1 | 9.8 | 2 | 7 | 0 | 4.3 | 0 |
| 7.1 | 4 | 12.1 | 0 | 10 | 1 | 5 | 1 | 9.2 | 3 | 7 | 1 | 4.2 | 0 |
| 7.1 | 2 | 12.6 | 0 | 9.8 | 1 | 3 | 9 | 10.1 | 1 | 8 | 0 | 4.5 | 0 |
| 7.3 | 0 | 11.9 | 0 | 9.9 | 0 | 6 | 0 | 10.4 | 0 | 7 | 1 | 4.5 | 0 |
| 7.5 | 0 | 12.8 | 1 | 9.3 | 1 | 5 | 0 | 8.7 | 0 | 7 | 0 | 4.3 | 3 |
| 8.9 | 0 | 13 | 0 | 9.3 | 0 | 5 | 1 | 8.3 | 0 | 6 | 0 | 4.4 | 8 |
| 9.5 | 1 | 11.2 | 4 | 9.1 | 5 | 5 | 0 | 7.6 | 1 | 6 | 0 | 4.7 | 1 |
| 10.3 | 0 | 11.1 | 3 | 8.5 | 1 | 5 | 0 | 7.5 | 2 | 6 | 2 | 4.6 | 1 |
| 140.2 | 0 | 10.9 | 0 | 7.5 | 3 | 6 | 0 | 7.7 | 2 | 7 | 7 | 5.3 | 0 |
| 10 | 0 | 10.9 | 1 | 7.7 | 3 | 6 | 2 | 7.7 | 3 | 6 | 0 | 5.5 | 1 |
| 9.4 | 0 | 9.9 | 2 | 6.6 | 1 | 6 | 3 | 7 | 0 | 7 | 0 | 5.4 | 0 |
| 8 | 1 | 9.8 | 0 | 5.8 | 1 | 6 | 0 | 6.4 | 1 | 7 | 3 | 5.5 | 0 |
| 7.9 | 0 | 9.8 | 4 | 5.7 | 0 | 5 | 0 | 5.2 | 1 | 5 | 2 | 5.6 | 1 |
| 7 | 0 | 8.9 | 2 | 5.9 | 0 | 5 | 0 | 5 | 0 | 6 | 1 | 5.7 | 0 |
| 6.5 | 3 | 8.3 | 2 | 5.2 | 0 | 5 | 2 | 6.6 | 0 | 5 | 1 | 5.3 | 0 |
| 6.4 | 2 | 7.5 | 4 | 5 | 0 | 5 | 3 | 7.2 | 3 | 5 | 0 | 5.3 | 0 |
| 6.3 | 1 | 7 | 0 | 5 | 0 | 5 | 0 | 7.3 | 0 | 5 | 3 | 5.4 | 0 |
| 5.5 | 0 | 7 | 0 | 5.2 | 0 | 5 | 4 | 7.4 | 1 | 5 | 2 | 5.8 | 0 |
| 6 | 0 | 7 | 4 | 5.6 | 0 | 5 | 0 | 6.9 | 1 | 6 | 0 | 5.4 | 0 |
| 5.9 | 3 | 0 | 0 | 5.6 | 0 | 5 | 4 | 6.1 | 0 | 5 | 0 | 4.6 | 0 |
| | | | | | | 5 | 0 | | | | | | |

Appendix 18. Density data for topshell (*Trochus* sp.) collected in April 2002. 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Depth | 1 | Depth | 2 | Depth | 3 | Depth | 4 | Depth | 5 | Depth | 6 | Depth | 7 |
|-------|---|-------|----|-------|----|-------|----|-------|----|-------|---|-------|---|
| 6 | 1 | 7 | 2 | 5.5 | 13 | 6 | 0 | 6.5 | 0 | 4.5 | 0 | 6 | 0 |
| 6 | 0 | 7 | 0 | 7 | 7 | 6.5 | 0 | 6.5 | 0 | 4.5 | 1 | 6.5 | 0 |
| 6 | 0 | 7.5 | 2 | 9 | 7 | 6 | 0 | 7.5 | 0 | 4.5 | 1 | 8.5 | 0 |
| 5.5 | 0 | 7.5 | 3 | 10 | 6 | 6 | 0 | 6.5 | 0 | 5 | 0 | 8.5 | 0 |
| 5.5 | 0 | 8 | 4 | 10 | 0 | 6 | 7 | 7 | 0 | 5 | 3 | 8 | 0 |
| 6 | 0 | 8 | 0 | 10 | 0 | 5.5 | 2 | 7.5 | 0 | 5 | 0 | 9 | 0 |
| 6.5 | 0 | 7.5 | 5 | 10 | 5 | 6 | 1 | 7.5 | 2 | 5 | 0 | 9.5 | 0 |
| 6.5 | 0 | 7 | 9 | 7.5 | 2 | 6.5 | 2 | 7 | 3 | 5 | 0 | 10 | 0 |
| 7 | 2 | 7 | 3 | 7 | 6 | 6 | 0 | 8.5 | 4 | 5 | 0 | 10 | 0 |
| 7.5 | 0 | 9 | 0 | 6.5 | 1 | 6 | 0 | 8.5 | 2 | 5 | 0 | 9.5 | 0 |
| 9.5 | 0 | 9 | 0 | 6.5 | 8 | 6 | 2 | 8 | 0 | 5 | 0 | 8.5 | 0 |
| 10 | 0 | 9 | 0 | 6 | 12 | 6 | 14 | 8.5 | 1 | 4 | 0 | 9 | 0 |
| 10 | 0 | 9 | 8 | 6 | 0 | 5.5 | 4 | 9 | 0 | 4 | 0 | 8 | 3 |
| 10 | 0 | 11 | 1 | 5 | 5 | 5.5 | 2 | 9.5 | 1 | 4 | 0 | 7 | 0 |
| 10 | 0 | 11 | 4 | 5.5 | 6 | 5.5 | 3 | 10 | 3 | 4 | 0 | 6.5 | 0 |
| 10 | 0 | 11 | 5 | 6.5 | 5 | 5 | 1 | 10 | 0 | 4 | 0 | 6.5 | 0 |
| 9.5 | 1 | 12 | 2 | 7 | 1 | 5.5 | 3 | 10 | 0 | 4 | 0 | 5.5 | 0 |
| 9.5 | 0 | 13 | 1 | 12.5 | 0 | 6 | 0 | 10 | 2 | 4 | 0 | 5 | 0 |
| 6 | 0 | 13 | 0 | 11 | 0 | 6 | 0 | 10 | 0 | 4 | 0 | 5.5 | 0 |
| 7.5 | 0 | 12 | 0 | 12 | 0 | 6 | 4 | 7 | 0 | 3.5 | 0 | 5 | 0 |
| 8 | 0 | 11 | 0 | 10.5 | 0 | 6 | 0 | 5.5 | 0 | 4 | 0 | 6.5 | 0 |
| 8 | 3 | 9.5 | 2 | 10.5 | 0 | 6 | 0 | 5 | 1 | 4 | 0 | 7 | 0 |
| 9.5 | 6 | 10 | 0 | 10 | 1 | 6 | 10 | 5.5 | 0 | 4 | 0 | 7 | 0 |
| 9.5 | 1 | 9 | 3 | 10 | 3 | 5.5 | 5 | 6 | 1 | 5 | 0 | 7.5 | 0 |
| 7 | 0 | 10 | 0 | 8 | 0 | 6.5 | 2 | 7.5 | 6 | 5 | 0 | 8 | 0 |
| 5.5 | 2 | 10 | 1 | 5 | 3 | 6.5 | 2 | 8 | 0 | 6 | 0 | 8 | 0 |
| 5 | 1 | 9 | 0 | 6 | 0 | 6.5 | 1 | 8 | 2 | 6 | 0 | 10 | 0 |
| 5 | 0 | 9 | 1 | 6 | 11 | 6.5 | 2 | 8.5 | 16 | 6 | 0 | 10 | 0 |
| 5 | 2 | 9.5 | 0 | 6.5 | 12 | 6 | 2 | 8 | 0 | 6 | 0 | 8.5 | 0 |
| 5 | 1 | 8.2 | 3 | 8 | 12 | 5.5 | 4 | 8 | 0 | 6 | 0 | 9 | 0 |
| 6.9 | 1 | 8 | 0 | 5 | 1 | 5 | 1 | 6.9 | 1 | 3.7 | 0 | 9 | 0 |
| 6.7 | 0 | 8.9 | 1 | 5.5 | 1 | 5 | 0 | 6.5 | 2 | 3.6 | 0 | 10 | 0 |
| 6.5 | 1 | 9.4 | 2 | 6 | 2 | 5 | 0 | 7 | 0 | 3.4 | 0 | 9 | 0 |
| 6.8 | 1 | 9.1 | 0 | 6.1 | 0 | 6 | 0 | 7.1 | 0 | 4.3 | 0 | 9 | 0 |
| 6.9 | 0 | 10.6 | 4 | 5.9 | 0 | 6 | 0 | 7.3 | 0 | 4.3 | 0 | 9 | 0 |
| 6.8 | 0 | 10.2 | 2 | 5.6 | 1 | 6 | 7 | 6.9 | 0 | 4.5 | 0 | 9 | 0 |
| 7 | 0 | 10.2 | 2 | 6.8 | 0 | 6 | 0 | 8.2 | 1 | 4.6 | 0 | 9 | 0 |
| 7 | 0 | 10.9 | 4 | 6.5 | 1 | 6 | 1 | 8.3 | 12 | 4.7 | 0 | 9 | 0 |
| 7.2 | 0 | 10.6 | 1 | 7.9 | 4 | 6 | 0 | 9.9 | 17 | 4.2 | 0 | 8 | 0 |
| 7.4 | 0 | 9.7 | 8 | 8.8 | 5 | 6 | 0 | 9.9 | 6 | 4.3 | 0 | 7 | 0 |
| 7.5 | 2 | 11.2 | 1 | 9.5 | 2 | 6 | 4 | 9.8 | 7 | 4.3 | 0 | 7 | 0 |
| 7.1 | 0 | 12.1 | 0 | 10 | 4 | 5 | 0 | 9.2 | 4 | 4.2 | 0 | 7 | 0 |
| 7.1 | 1 | 12.6 | 0 | 9.8 | 12 | 3 | 0 | 10.1 | 5 | 4.5 | 0 | 8 | 1 |
| 7.3 | 1 | 11.9 | 2 | 9.9 | 1 | 6 | 1 | 10.4 | 0 | 4.5 | 6 | 7 | 0 |
| 7.5 | 0 | 12.8 | 0 | 9.3 | 2 | 5 | 4 | 8.7 | 3 | 4.3 | 0 | 7 | 0 |
| 8.9 | 0 | 13 | 0 | 9.3 | 1 | 5 | 2 | 8.3 | 2 | 4.4 | 2 | 6 | 0 |
| 9.5 | 0 | 11.2 | 0 | 9.1 | 2 | 5 | 0 | 7.6 | 0 | 4.7 | 1 | 6 | 0 |
| 10.3 | 0 | 11.1 | 5 | 8.5 | 3 | 5 | 2 | 7.5 | 2 | 4.6 | 0 | 6 | 0 |
| 140.2 | 0 | 10.9 | 0 | 7.5 | 1 | 6 | 0 | 7.7 | 0 | 5.3 | 0 | 7 | 0 |
| 10 | 0 | 10.9 | 0 | 7.7 | 2 | 6 | 0 | 7.7 | 0 | 5.5 | 3 | 6 | 0 |
| 9.4 | 1 | 9.9 | 0 | 6.6 | 0 | 6 | 0 | 7 | 0 | 5.4 | 0 | 7 | 0 |
| 8 | 0 | 9.8 | 0 | 5.8 | 0 | 6 | 0 | 6.4 | 0 | 5.5 | 0 | 7 | 0 |
| 7.9 | 1 | 9.8 | 0 | 5.7 | 0 | 5 | 0 | 5.2 | 0 | 5.6 | 0 | 5 | 0 |
| 7 | 0 | 8.9 | 2 | 5.9 | 0 | 5 | 2 | 5 | 0 | 5.7 | 0 | 6 | 0 |
| 6.5 | 0 | 8.3 | 13 | 5.2 | 0 | 5 | 2 | 6.6 | 1 | 5.3 | 0 | 5 | 0 |
| 6.4 | 0 | 7.5 | 6 | 5 | 0 | 5 | 0 | 7.2 | 1 | 5.3 | 0 | 5 | 0 |
| 6.3 | 0 | 7 | 3 | 5 | 0 | 5 | 4 | 7.3 | 2 | 5.4 | 0 | 5 | 0 |
| 5.5 | 0 | 7 | 1 | 5.2 | 1 | 5 | 8 | 7.4 | 2 | 5.8 | 0 | 5 | 0 |
| 6 | 0 | 7 | 0 | 5.6 | 0 | 5 | 0 | 6.9 | 0 | 5.4 | 0 | 6 | 0 |
| 5.9 | 1 | 0 | 0 | 5.6 | 2 | 5 | 0 | 6.1 | 0 | 4.6 | 0 | 5 | 1 |
| | | | | | | 5 | 0 | | | | | | |

Appendix 19. Density data for cats eye snail (*Turbo smaragdus*) collected in April 2002. 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Depth | 1 | Depth | 2 | Depth | 3 | Depth | 4 | Depth | 5 | Depth | 6 | Depth | 7 |
|-------|---|-------|---|-------|----|-------|----|-------|-----|-------|----|-------|---|
| 6 | 0 | 7 | 0 | 5.5 | 6 | 6 | 0 | 6.5 | 9 | 4.5 | 19 | 6 | 0 |
| 6 | 0 | 7 | 0 | 7 | 2 | 6.5 | 4 | 6.5 | 9.2 | 4.5 | 7 | 6.5 | 0 |
| 6 | 1 | 7.5 | 0 | 9 | 0 | 6 | 6 | 7.5 | 14 | 4.5 | 19 | 8.5 | 0 |
| 5.5 | 0 | 7.5 | 0 | 10 | 0 | 6 | 6 | 6.5 | 9 | 5 | 14 | 8.5 | 0 |
| 5.5 | 0 | 8 | 0 | 10 | 0 | 6 | 6 | 7 | 14 | 5 | 15 | 8 | 0 |
| 6 | 1 | 8 | 0 | 10 | 0 | 5.5 | 9 | 7.5 | 5 | 5 | 24 | 9 | 0 |
| 6.5 | 0 | 7.5 | 0 | 10 | 0 | 6 | 5 | 7.5 | 8 | 5 | 21 | 9.5 | 0 |
| 6.5 | 1 | 7 | 0 | 7.5 | 0 | 6.5 | 6 | 7 | 6 | 5 | 27 | 10 | 0 |
| 7 | 0 | 7 | 0 | 7 | 0 | 6 | 0 | 8.5 | 2 | 5 | 26 | 10 | 0 |
| 7.5 | 0 | 9 | 0 | 6.5 | 3 | 6 | 0 | 8.5 | 1 | 5 | 39 | 9.5 | 0 |
| 9.5 | 0 | 9 | 0 | 6.5 | 0 | 6 | 2 | 8 | 3 | 5 | 11 | 8.5 | 0 |
| 10 | 0 | 9 | 0 | 6 | 1 | 6 | 5 | 8.5 | 1 | 4 | 24 | 9 | 0 |
| 10 | 0 | 9 | 0 | 6 | 4 | 5.5 | 2 | 9 | 1 | 4 | 15 | 8 | 0 |
| 10 | 0 | 11 | 0 | 5 | 0 | 5.5 | 5 | 9.5 | 0 | 4 | 19 | 7 | 0 |
| 10 | 0 | 11 | 0 | 5.5 | 0 | 5.5 | 0 | 10 | 0 | 4 | 9 | 6.5 | 0 |
| 10 | 0 | 11 | 0 | 6.5 | 2 | 5 | 0 | 10 | 0 | 4 | 29 | 6.5 | 0 |
| 9.5 | 0 | 12 | 0 | 7 | 0 | 5.5 | 0 | 10 | 1 | 4 | 17 | 5.5 | 0 |
| 9.5 | 0 | 13 | 0 | 12.5 | 0 | 6 | 4 | 10 | 1 | 4 | 10 | 5 | 0 |
| 6 | 0 | 13 | 0 | 11 | 0 | 6 | 6 | 10 | 0 | 4 | 21 | 5.5 | 0 |
| 7.5 | 0 | 12 | 0 | 12 | 0 | 6 | 0 | 7 | 1 | 3.5 | 26 | 5 | 0 |
| 8 | 0 | 11 | 0 | 10.5 | 0 | 6 | 3 | 5.5 | 21 | 4 | 36 | 6.5 | 0 |
| 8 | 0 | 9.5 | 0 | 10.5 | 0 | 6 | 1 | 5 | 4 | 4 | 21 | 7 | 0 |
| 9.5 | 0 | 10 | 0 | 10 | 0 | 6 | 1 | 5.5 | 5 | 4 | 34 | 7 | 0 |
| 9.5 | 0 | 9 | 0 | 10 | 7 | 5.5 | 1 | 6 | 4 | 5 | 25 | 7.5 | 0 |
| 7 | 0 | 10 | 0 | 8 | 0 | 6.5 | 1 | 7.5 | 0 | 5 | 26 | 8 | 0 |
| 5.5 | 0 | 10 | 1 | 5 | 4 | 6.5 | 0 | 8 | 0 | 6 | 7 | 8 | 0 |
| 5 | 0 | 9 | 1 | 6 | 6 | 6.5 | 0 | 8 | 1 | 6 | 7 | 10 | 0 |
| 5 | 0 | 9 | 1 | 6 | 4 | 6.5 | 2 | 8.5 | 1 | 6 | 8 | 10 | 0 |
| 5 | 0 | 9.5 | 0 | 6.5 | 5 | 6 | 0 | 8 | 10 | 6 | 0 | 8.5 | 0 |
| 5 | 0 | 8.2 | 0 | 8 | 0 | 5.5 | 0 | 8 | 5 | 6 | 20 | 9 | 0 |
| 6.9 | 1 | 8 | 0 | 5 | 21 | 5 | 0 | 6.9 | 9 | 3.7 | 22 | 9 | 0 |
| 6.7 | 0 | 8.9 | 0 | 5.5 | 3 | 5 | 2 | 6.5 | 14 | 3.6 | 55 | 10 | 0 |
| 6.5 | 2 | 9.4 | 0 | 6 | 8 | 5 | 13 | 7 | 13 | 3.4 | 32 | 9 | 0 |
| 6.8 | 0 | 9.1 | 0 | 6.1 | 1 | 6 | 8 | 7.1 | 22 | 4.3 | 15 | 9 | 0 |
| 6.9 | 0 | 10.6 | 0 | 5.9 | 8 | 6 | 0 | 7.3 | 7 | 4.3 | 22 | 9 | 0 |
| 6.8 | 0 | 10.2 | 0 | 5.6 | 0 | 6 | 0 | 6.9 | 10 | 4.5 | 15 | 9 | 0 |
| 7 | 0 | 10.2 | 0 | 6.8 | 15 | 6 | 6 | 8.2 | 13 | 4.6 | 8 | 9 | 0 |
| 7 | 0 | 10.9 | 0 | 6.5 | 20 | 6 | 8 | 8.3 | 2 | 4.7 | 4 | 9 | 0 |
| 7.2 | 0 | 10.6 | 0 | 7.9 | 3 | 6 | 4 | 9.9 | 0 | 4.2 | 4 | 8 | 0 |
| 7.4 | 0 | 9.7 | 0 | 8.8 | 0 | 6 | 8 | 9.9 | 1 | 4.3 | 0 | 7 | 0 |
| 7.5 | 0 | 11.2 | 0 | 9.5 | 3 | 6 | 10 | 9.8 | 3 | 4.3 | 8 | 7 | 0 |
| 7.1 | 0 | 12.1 | 0 | 10 | 0 | 5 | 1 | 9.2 | 0 | 4.2 | 9 | 7 | 0 |
| 7.1 | 0 | 12.6 | 0 | 9.8 | 0 | 3 | 5 | 10.1 | 0 | 4.5 | 5 | 8 | 0 |
| 7.3 | 0 | 11.9 | 0 | 9.9 | 2 | 6 | 12 | 10.4 | 0 | 4.5 | 3 | 7 | 0 |
| 7.5 | 0 | 12.8 | 0 | 9.3 | 0 | 5 | 9 | 8.7 | 5 | 4.3 | 6 | 7 | 0 |
| 8.9 | 0 | 13 | 0 | 9.3 | 0 | 5 | 4 | 8.3 | 10 | 4.4 | 3 | 6 | 0 |
| 9.5 | 0 | 11.2 | 0 | 9.1 | 0 | 5 | 4 | 7.6 | 13 | 4.7 | 7 | 6 | 0 |
| 10.3 | 0 | 11.1 | 0 | 8.5 | 0 | 5 | 0 | 7.5 | 14 | 4.6 | 13 | 6 | 0 |
| 140 | 0 | 10.9 | 0 | 7.5 | 1 | 6 | 1 | 7.7 | 11 | 5.3 | 10 | 7 | 0 |
| 10 | 0 | 10.9 | 0 | 7.7 | 0 | 6 | 3 | 7.7 | 4 | 5.5 | 0 | 6 | 0 |
| 9.4 | 0 | 9.9 | 0 | 6.6 | 10 | 6 | 3 | 7 | 3 | 5.4 | 5 | 7 | 0 |
| 8 | 0 | 9.8 | 0 | 5.8 | 20 | 6 | 8 | 6.4 | 16 | 5.5 | 5 | 7 | 0 |
| 7.9 | 0 | 9.8 | 0 | 5.7 | 13 | 5 | 4 | 5.2 | 25 | 5.6 | 6 | 5 | 0 |
| 7 | 0 | 8.9 | 0 | 5.9 | 17 | 5 | 2 | 5 | 23 | 5.7 | 1 | 6 | 1 |
| 6.5 | 0 | 8.3 | 0 | 5.2 | 9 | 5 | 4 | 6.6 | 13 | 5.3 | 8 | 5 | 0 |
| 6.4 | 0 | 7.5 | 0 | 5 | 20 | 5 | 7 | 7.2 | 12 | 5.3 | 8 | 5 | 0 |
| 6.3 | 0 | 7 | 0 | 5 | 20 | 5 | 4 | 7.3 | 7 | 5.4 | 3 | 5 | 0 |
| 5.5 | 0 | 7 | 0 | 5.2 | 11 | 5 | 0 | 7.4 | 5 | 5.8 | 12 | 5 | 0 |
| 6 | 0 | 7 | 0 | 5.6 | 2 | 5 | 3 | 6.9 | 8 | 5.4 | 19 | 6 | 0 |
| 5.9 | 0 | 0 | 0 | 5.6 | 11 | 5 | 8 | 6.1 | 10 | 4.6 | 22 | 5 | 0 |
| | | | | | | 5 | 3 | | | | | | |

Appendix 20. Density data for limpets (*Cellana* sp.) collected in April 2002. 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Depth | 1 | Depth | 2 | Depth | 3 | Depth | 4 | Depth | 5 | Depth | 6 | Depth | 7 |
|-------|---|-------|---|-------|---|-------|---|-------|---|-------|----|-------|---|
| 6 | 0 | 7 | 8 | 5.5 | 0 | 6 | 0 | 6.5 | 0 | 4.5 | 8 | 6 | 0 |
| 6 | 2 | 7 | 3 | 7 | 0 | 6.5 | 0 | 6.5 | 0 | 4.5 | 10 | 6.5 | 0 |
| 6 | 3 | 7.5 | 6 | 9 | 0 | 6 | 0 | 7.5 | 0 | 4.5 | 7 | 8.5 | 0 |
| 5.5 | 1 | 7.5 | 9 | 10 | 0 | 6 | 0 | 6.5 | 0 | 5 | 6 | 8.5 | 0 |
| 5.5 | 0 | 8 | 6 | 10 | 0 | 6 | 0 | 7 | 0 | 5 | 5 | 8 | 0 |
| 6 | 0 | 8 | 2 | 10 | 0 | 5.5 | 0 | 7.5 | 0 | 5 | 4 | 9 | 0 |
| 6.5 | 0 | 7.5 | 5 | 10 | 0 | 6 | 0 | 7.5 | 0 | 5 | 2 | 9.5 | 0 |
| 6.5 | 0 | 7 | 6 | 7.5 | 0 | 6.5 | 0 | 7 | 0 | 5 | 1 | 10 | 0 |
| 7 | 0 | 7 | 5 | 7 | 0 | 6 | 1 | 8.5 | 0 | 5 | 9 | 10 | 0 |
| 7.5 | 0 | 9 | 3 | 6.5 | 0 | 6 | 0 | 8.5 | 0 | 5 | 4 | 9.5 | 0 |
| 9.5 | 0 | 9 | 2 | 6.5 | 0 | 6 | 0 | 8 | 0 | 5 | 0 | 8.5 | 0 |
| 10 | 0 | 9 | 0 | 6 | 0 | 6 | 0 | 8.5 | 0 | 4 | 0 | 9 | 0 |
| 10 | 0 | 9 | 1 | 6 | 0 | 5.5 | 0 | 9 | 0 | 4 | 1 | 8 | 0 |
| 10 | 0 | 11 | 0 | 5 | 0 | 5.5 | 0 | 9.5 | 0 | 4 | 8 | 7 | 0 |
| 10 | 0 | 11 | 0 | 5.5 | 0 | 5.5 | 0 | 10 | 0 | 4 | 0 | 6.5 | 0 |
| 10 | 0 | 11 | 0 | 6.5 | 0 | 5 | 0 | 10 | 0 | 4 | 6 | 6.5 | 0 |
| 9.5 | 0 | 12 | 0 | 7 | 0 | 5.5 | 0 | 10 | 0 | 4 | 3 | 5.5 | 0 |
| 9.5 | 0 | 13 | 0 | 12.5 | 0 | 6 | 0 | 10 | 0 | 4 | 13 | 5 | 0 |
| 6 | 0 | 13 | 0 | 11 | 0 | 6 | 0 | 10 | 0 | 4 | 6 | 5.5 | 0 |
| 7.5 | 0 | 12 | 0 | 12 | 0 | 6 | 0 | 7 | 0 | 3.5 | 5 | 5 | 0 |
| 8 | 0 | 11 | 0 | 10.5 | 0 | 6 | 0 | 5.5 | 0 | 4 | 4 | 6.5 | 0 |
| 8 | 0 | 9.5 | 0 | 10.5 | 0 | 6 | 0 | 5 | 1 | 4 | 5 | 7 | 0 |
| 9.5 | 0 | 10 | 0 | 10 | 0 | 6 | 0 | 5.5 | 0 | 4 | 2 | 7 | 0 |
| 9.5 | 0 | 9 | 3 | 10 | 0 | 5.5 | 0 | 6 | 0 | 5 | 2 | 7.5 | 0 |
| 7 | 0 | 10 | 0 | 8 | 0 | 6.5 | 0 | 7.5 | 0 | 5 | 5 | 8 | 0 |
| 5.5 | 0 | 10 | 1 | 5 | 0 | 6.5 | 0 | 8 | 0 | 6 | 5 | 8 | 0 |
| 5 | 0 | 9 | 0 | 6 | 0 | 6.5 | 0 | 8 | 0 | 6 | 1 | 10 | 0 |
| 5 | 1 | 9 | 2 | 6 | 0 | 6.5 | 0 | 8.5 | 1 | 6 | 1 | 10 | 0 |
| 5 | 1 | 9.5 | 2 | 6.5 | 0 | 6 | 0 | 8 | 0 | 6 | 0 | 8.5 | 0 |
| 5 | 6 | 8.2 | 0 | 8 | 0 | 5.5 | 0 | 8 | 0 | 6 | 4 | 9 | 0 |
| 6.9 | 0 | 8 | 0 | 5 | 2 | 5 | 0 | 6.9 | 0 | 3.7 | 0 | 9 | 0 |
| 6.7 | 0 | 8.9 | 0 | 5.5 | 1 | 5 | 0 | 6.5 | 0 | 3.6 | 2 | 10 | 0 |
| 6.5 | 0 | 9.4 | 0 | 6 | 0 | 5 | 0 | 7 | 0 | 3.4 | 2 | 9 | 0 |
| 6.8 | 0 | 9.1 | 0 | 6.1 | 0 | 6 | 0 | 7.1 | 0 | 4.3 | 1 | 9 | 0 |
| 6.9 | 0 | 10.6 | 0 | 5.9 | 0 | 6 | 0 | 7.3 | 0 | 4.3 | 3 | 9 | 0 |
| 6.8 | 1 | 10.2 | 0 | 5.6 | 0 | 6 | 0 | 6.9 | 0 | 4.5 | 1 | 9 | 0 |
| 7 | 0 | 10.2 | 0 | 6.8 | 0 | 6 | 0 | 8.2 | 0 | 4.6 | 3 | 9 | 0 |
| 7 | 0 | 10.9 | 0 | 6.5 | 0 | 6 | 0 | 8.3 | 0 | 4.7 | 0 | 9 | 0 |
| 7.2 | 0 | 10.6 | 0 | 7.9 | 0 | 6 | 0 | 9.9 | 0 | 4.2 | 0 | 8 | 0 |
| 7.4 | 0 | 9.7 | 0 | 8.8 | 0 | 6 | 0 | 9.9 | 0 | 4.3 | 2 | 7 | 0 |
| 7.5 | 0 | 11.2 | 0 | 9.5 | 0 | 6 | 0 | 9.8 | 0 | 4.3 | 2 | 7 | 0 |
| 7.1 | 0 | 12.1 | 0 | 10 | 1 | 5 | 0 | 9.2 | 0 | 4.2 | 4 | 7 | 0 |
| 7.1 | 0 | 12.6 | 0 | 9.8 | 0 | 3 | 0 | 10.1 | 0 | 4.5 | 3 | 8 | 0 |
| 7.3 | 0 | 11.9 | 0 | 9.9 | 0 | 6 | 0 | 10.4 | 0 | 4.5 | 2 | 7 | 0 |
| 7.5 | 0 | 12.8 | 0 | 9.3 | 0 | 5 | 0 | 8.7 | 0 | 4.3 | 6 | 7 | 0 |
| 8.9 | 0 | 13 | 0 | 9.3 | 0 | 5 | 0 | 8.3 | 0 | 4.4 | 0 | 6 | 0 |
| 9.5 | 0 | 11.2 | 0 | 9.1 | 0 | 5 | 0 | 7.6 | 0 | 4.7 | 1 | 6 | 0 |
| 10.3 | 0 | 11.1 | 0 | 8.5 | 0 | 5 | 1 | 7.5 | 0 | 4.6 | 5 | 6 | 0 |
| 140.2 | 0 | 10.9 | 0 | 7.5 | 0 | 6 | 0 | 7.7 | 0 | 5.3 | 2 | 7 | 0 |
| 10 | 0 | 10.9 | 2 | 7.7 | 0 | 6 | 0 | 7.7 | 0 | 5.5 | 0 | 6 | 0 |
| 9.4 | 0 | 9.9 | 0 | 6.6 | 0 | 6 | 0 | 7 | 0 | 5.4 | 3 | 7 | 0 |
| 8 | 0 | 9.8 | 0 | 5.8 | 0 | 6 | 0 | 6.4 | 0 | 5.5 | 8 | 7 | 0 |
| 7.9 | 0 | 9.8 | 0 | 5.7 | 0 | 5 | 0 | 5.2 | 1 | 5.6 | 6 | 5 | 0 |
| 7 | 0 | 8.9 | 0 | 5.9 | 2 | 5 | 0 | 5 | 0 | 5.7 | 6 | 6 | 0 |
| 6.5 | 0 | 8.3 | 0 | 5.2 | 2 | 5 | 0 | 6.6 | 0 | 5.3 | 4 | 5 | 0 |
| 6.4 | 0 | 7.5 | 0 | 5 | 1 | 5 | 1 | 7.2 | 0 | 5.3 | 3 | 5 | 0 |
| 6.3 | 0 | 7 | 0 | 5 | 0 | 5 | 0 | 7.3 | 0 | 5.4 | 3 | 5 | 0 |
| 5.5 | 0 | 7 | 0 | 5.2 | 1 | 5 | 0 | 7.4 | 0 | 5.8 | 6 | 5 | 0 |
| 6 | 0 | 7 | 0 | 5.6 | 1 | 5 | 0 | 6.9 | 0 | 5.4 | 2 | 6 | 0 |
| 5.9 | 0 | 0 | 0 | 5.6 | 0 | 5 | 0 | 6.1 | 0 | 4.6 | 2 | 5 | 0 |
| | | | | | | 5 | 0 | | | | | | |

Appendix 21. Kina size class data collected in April 2002 from the ATNP coast. 1=Separation Pt., 2=Awaroa Head, 3=Canoe Bay, 4=Arch Point, 5=Foul Point, 6=Whale Rock, 7=Totara Rock.

| Size | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
|------|---|---|----|----|---|----|----|
| 30 | | | | | | | |
| 31 | | | | | | 1 | |
| 32 | | | 1 | | | | |
| 33 | | | | | | | |
| 34 | | | | | | | |
| 35 | | | | | | | 2 |
| 36 | | 1 | | | | | 3 |
| 37 | 1 | | | | | | 2 |
| 38 | | | | | | | 6 |
| 39 | 9 | | | | 1 | 1 | 5 |
| 40 | 1 | | | 1 | | 4 | 7 |
| 41 | | | | 2 | 1 | | 8 |
| 42 | 2 | 1 | 1 | | 1 | 1 | 9 |
| 43 | 3 | | 2 | | | 1 | 13 |
| 44 | 1 | | 2 | | 1 | 1 | 8 |
| 45 | 2 | | 2 | 5 | | 2 | 10 |
| 46 | 2 | 2 | 2 | 2 | 2 | 2 | 8 |
| 47 | | 1 | 1 | | 1 | | 3 |
| 48 | 2 | 5 | | | 2 | 2 | 8 |
| 49 | 2 | 7 | 3 | 1 | 1 | 2 | 2 |
| 50 | 2 | | 5 | 9 | 1 | 4 | 9 |
| 51 | 4 | 1 | 5 | | 3 | | 5 |
| 52 | 8 | 2 | 1 | 1 | 4 | | 3 |
| 53 | 9 | 2 | 2 | 1 | 4 | 2 | 4 |
| 54 | 7 | 7 | 3 | 4 | 6 | 2 | 4 |
| 55 | 8 | 4 | 4 | 8 | 5 | 14 | 1 |
| 56 | 5 | 4 | 6 | 1 | 2 | 3 | 2 |
| 57 | 2 | | | 3 | 5 | | 1 |
| 58 | 3 | 1 | 5 | 2 | 8 | 3 | |
| 59 | 7 | 3 | 8 | 1 | 1 | 1 | 1 |
| 60 | 1 | 3 | 8 | 14 | 7 | 15 | |
| 61 | 3 | 7 | 10 | 4 | 6 | | |
| 62 | 3 | 3 | 9 | 8 | 3 | 2 | |
| 63 | 2 | 2 | 5 | 5 | 3 | 3 | |
| 64 | 1 | 4 | 4 | 3 | 6 | 6 | |
| 65 | | 5 | 3 | 7 | 4 | 7 | |
| 66 | | 2 | 3 | 2 | 1 | 2 | |
| 67 | 1 | 3 | 2 | 3 | 3 | | 1 |
| 68 | | 1 | 2 | | 2 | 2 | |
| 69 | | 3 | 2 | 1 | 3 | | |
| 70 | | 2 | 3 | 2 | 2 | 8 | |
| 71 | 1 | 1 | 1 | 2 | 1 | 3 | |
| 72 | | | 2 | 1 | | 4 | |
| 73 | | 1 | 2 | | | | |
| 74 | 1 | | 4 | 1 | | | |
| 75 | | | 2 | 1 | 2 | 2 | |
| 76 | | | | 1 | 1 | | |
| 77 | | | | 1 | | 1 | |
| 78 | | | | | | | |
| 79 | 1 | | | 1 | | | |
| 80 | | | 1 | | | 2 | |
| 81 | | | 1 | | | | |
| 82 | | | 1 | | | | |
| 83 | | | | | | | |
| 84 | | | 1 | | | | |
| 85 | | | 1 | | | 1 | |
| 86 | | 1 | | | 1 | | |
| 87 | | | | | 1 | | |
| 88 | | | | | | | |
| 89 | | | | | | | |
| 90 | | | | | | | |
| 91 | | | | 1 | | | |
| 92 | | | | | | | |
| 93 | | | | | | | |
| 94 | | | | | | | |
| 95 | | | | | | | |
| 96 | | | 1 | 1 | | | |

Appendix 22. Cook's turban size class data collected in April 2002 from the ATNP coast.

| Reserve | | | | | | Control | | | | | | |
|---------|----|----|----|----|----|---------|----|----|----|----|----|----|
| 26 | 46 | 53 | 28 | 32 | 38 | 29 | 53 | 32 | 40 | 53 | 34 | 46 |
| 27 | 46 | 53 | 28 | 32 | | 31 | 54 | 32 | 40 | 54 | 34 | 46 |
| 27 | 46 | 53 | 28 | 32 | | 31 | 54 | 32 | 41 | 55 | 34 | 46 |
| 27 | 47 | 53 | 28 | 32 | | 31 | 54 | 32 | 41 | 55 | 35 | 46 |
| 28 | 47 | 53 | 28 | 32 | | 31 | 54 | 33 | 41 | 58 | 35 | 46 |
| 28 | 47 | 53 | 28 | 32 | | 34 | 54 | 33 | 41 | 58 | 35 | 46 |
| 29 | 48 | 54 | 28 | 32 | | 34 | 55 | 33 | 42 | 59 | 35 | 46 |
| 30 | 48 | 54 | 28 | 32 | | 34 | 55 | 33 | 43 | 60 | 35 | 46 |
| 30 | 48 | 54 | 29 | 32 | | 35 | 55 | 33 | 43 | 60 | 36 | 47 |
| 30 | 48 | 54 | 29 | 33 | | 35 | 55 | 33 | 43 | 60 | 36 | 47 |
| 31 | 48 | 54 | 29 | 33 | | 35 | 55 | 34 | 43 | 22 | 36 | 47 |
| 31 | 48 | 54 | 29 | 33 | | 36 | 56 | 34 | 43 | 24 | 36 | 47 |
| 31 | 48 | 54 | 29 | 33 | | 36 | 56 | 35 | 43 | 25 | 36 | 47 |
| 33 | 48 | 55 | 29 | 33 | | 37 | 56 | 35 | 43 | 26 | 36 | 47 |
| 34 | 49 | 55 | 29 | 33 | | 37 | 56 | 35 | 44 | 27 | 37 | 47 |
| 35 | 49 | 55 | 29 | 34 | | 37 | 56 | 35 | 45 | 27 | 37 | 48 |
| 35 | 49 | 55 | 30 | 34 | | 37 | 56 | 35 | 45 | 27 | 37 | 48 |
| 35 | 49 | 55 | 30 | 34 | | 40 | 56 | 35 | 45 | 27 | 38 | 48 |
| 36 | 49 | 56 | 30 | 34 | | 40 | 56 | 36 | 45 | 27 | 38 | 48 |
| 36 | 49 | 56 | 30 | 34 | | 42 | 56 | 36 | 45 | 27 | 38 | 48 |
| 37 | 49 | 56 | 30 | 34 | | 42 | 57 | 36 | 45 | 28 | 38 | 48 |
| 37 | 50 | 56 | 30 | 34 | | 43 | 57 | 36 | 46 | 28 | 38 | 48 |
| 37 | 50 | 57 | 30 | 34 | | 43 | 57 | 36 | 46 | 29 | 38 | 48 |
| 38 | 50 | 57 | 30 | 34 | | 44 | 57 | 36 | 46 | 29 | 38 | 48 |
| 38 | 50 | 57 | 30 | 34 | | 44 | 57 | 36 | 47 | 29 | 38 | 49 |
| 38 | 50 | 57 | 30 | 34 | | 44 | 57 | 36 | 47 | 29 | 38 | 50 |
| 38 | 50 | 57 | 30 | 34 | | 44 | 57 | 36 | 47 | 29 | 38 | 50 |
| 38 | 50 | 58 | 30 | 35 | | 45 | 57 | 37 | 48 | 30 | 39 | 50 |
| 38 | 50 | 59 | 30 | 35 | | 45 | 58 | 37 | 48 | 30 | 40 | 50 |
| 38 | 50 | 60 | 30 | 35 | | 45 | 58 | 37 | 48 | 30 | 40 | 50 |
| 39 | 50 | 60 | 30 | 35 | | 46 | 58 | 37 | 48 | 31 | 40 | 51 |
| 39 | 50 | 61 | 30 | 35 | | 46 | 59 | 37 | 48 | 31 | 40 | 51 |
| 40 | 50 | 18 | 30 | 35 | | 46 | 60 | 37 | 48 | 31 | 40 | 52 |
| 40 | 50 | 19 | 30 | 35 | | 46 | 60 | 37 | 48 | 31 | 40 | 53 |
| 41 | 51 | 20 | 30 | 35 | | 47 | 60 | 37 | 49 | 31 | 41 | 55 |
| 42 | 51 | 20 | 31 | 35 | | 47 | 61 | 38 | 49 | 32 | 41 | 62 |
| 42 | 51 | 21 | 31 | 36 | | 47 | 61 | 38 | 49 | 32 | 41 | 63 |
| 42 | 51 | 24 | 31 | 36 | | 48 | 61 | 38 | 49 | 32 | 41 | 32 |
| 42 | 51 | 24 | 31 | 36 | | 48 | 64 | 38 | 50 | 32 | 42 | |
| 42 | 51 | 25 | 31 | 36 | | 48 | 64 | 38 | 50 | 32 | 42 | |
| 43 | 51 | 25 | 31 | 36 | | 49 | 64 | 38 | 50 | 32 | 42 | |
| 43 | 51 | 25 | 31 | 36 | | 50 | 65 | 38 | 50 | 33 | 42 | |
| 43 | 51 | 25 | 31 | 36 | | 50 | 66 | 38 | 50 | 33 | 42 | |
| 43 | 51 | 26 | 31 | 36 | | 50 | 67 | 38 | 50 | 33 | 43 | |
| 44 | 52 | 26 | 31 | 36 | | 50 | 20 | 39 | 50 | 33 | 43 | |
| 44 | 52 | 26 | 31 | 36 | | 51 | 25 | 39 | 51 | 33 | 43 | |
| 44 | 52 | 26 | 31 | 37 | | 51 | 26 | 39 | 51 | 33 | 44 | |
| 45 | 52 | 27 | 31 | 37 | | 51 | 28 | 39 | 51 | 33 | 44 | |
| 45 | 52 | 27 | 31 | 37 | | 52 | 29 | 39 | 51 | 33 | 44 | |
| 45 | 52 | 27 | 32 | 37 | | 52 | 30 | 40 | 52 | 34 | 44 | |
| 46 | 52 | 27 | 32 | 37 | | 52 | 30 | 40 | 52 | 34 | 44 | |
| 46 | 52 | 27 | 32 | 37 | | 53 | 31 | 40 | 52 | 34 | 44 | |
| 53 | 52 | 27 | 32 | 38 | | 53 | 32 | 40 | 52 | 34 | 44 | |
| 38 | 53 | 27 | 32 | 38 | | 53 | 32 | 40 | 53 | 34 | 45 | |

Appendix 23. Scallop size data collected in March 2003 from the ATNP coast.

| SITES | 1 | 2 | 3 | 4 | 4 (continued) | 5 |
|--------------------|-------------|-------------|-------------|-----|---------------|-------------|
| | 70 | 106 | 102 | 104 | 73 | 82 |
| | 79 | 104 | 62 | 94 | 74 | 70 |
| | 75 | 83 | 105 | 104 | 74 | 87 |
| | 92 | 76 | 102 | 94 | 74 | 93 |
| | 87 | 80 | 66 | 75 | 55 | 100 |
| | 80 | 75 | 81 | 93 | 78 | 84 |
| | 69 | 100 | 55 | 76 | 98 | 75 |
| | 71 | 102 | 93 | 79 | 82 | 92 |
| | 81 | 79 | 109 | 73 | 77 | 98 |
| | 86 | 90 | 99 | 76 | 76 | 70 |
| | 84 | 93 | 62 | 82 | 72 | 78 |
| | 58 | 108 | 49 | 83 | 69 | 84 |
| | 45 | 86 | 77 | 75 | 67 | 89 |
| | 87 | 93 | 60 | 72 | 74 | 101 |
| | 74 | 80 | 97 | 116 | 65 | 73 |
| | 86 | 90 | 61 | 102 | 84 | 108 |
| | 85 | 82 | 67 | 113 | 86 | 73 |
| | 66 | 82 | 99 | 76 | 69 | 84 |
| | 78 | 82 | 104 | 75 | 76 | 84 |
| | 79 | 82 | 96 | 80 | 72 | 98 |
| | 86 | 68 | 81 | 101 | 70 | 63 |
| | 83 | 84 | 101 | 100 | 71 | 74 |
| | 81 | 78 | 109 | 79 | 67 | 77 |
| | 69 | 96 | 100 | 75 | 77 | 77 |
| | 71 | 92 | 69 | 78 | 70 | 92 |
| | 76 | 106 | 79 | 73 | 88 | 73 |
| | 76 | 72 | 74 | 74 | 79 | 82 |
| | 67 | 80 | 96 | 30 | 80 | 98 |
| | 85 | 76 | 91 | 71 | 75 | 77 |
| | 82 | 98 | 56 | 80 | 70 | 98 |
| | 69 | 103 | | 75 | 79 | 84 |
| | 84 | 98 | | 64 | 77 | 85 |
| | 74 | 68 | | 83 | 78 | 82 |
| | 84 | 106 | | 86 | 72 | 70 |
| | 80 | 87 | | 75 | 78 | 90 |
| | 72 | 96 | | 78 | 66 | 80 |
| | 68 | 82 | | 80 | 73 | 82 |
| | 73 | 90 | | 80 | 71 | 66 |
| | 80 | 100 | | 90 | 85 | 68 |
| | 90 | 94 | | 76 | 75 | 66 |
| | 82 | 96 | | 84 | 76 | 87 |
| | 82 | 89 | | 79 | 74 | 94 |
| | 51 | 82 | | 73 | 65 | 99 |
| | 74 | 88 | | 70 | 72 | 71 |
| | 77 | 86 | | 75 | 81 | 106 |
| | 65 | 107 | | 73 | 70 | 69 |
| | 62 | 86 | | 91 | 102 | 70 |
| | 78 | 85 | | 80 | 93 | 59 |
| | 81 | 109 | | 67 | 78 | 85 |
| | 91 | 112 | | 71 | 99 | 92 |
| | 71 | 101 | | 75 | 95 | 75 |
| | 76 | 94 | | 68 | 92 | 78 |
| | 69 | 101 | | 80 | 106 | 95 |
| | 85 | 97 | | 72 | 91 | 98 |
| | 88 | 76 | | 90 | 79 | 93 |
| | 76 | 71 | | 69 | 83 | 94 |
| | 83 | 122 | | 68 | 77 | 75 |
| | 66 | 87 | | 101 | 84 | 61 |
| | 68 | 90 | | 80 | 74 | 59 |
| | 79 | 106 | | 64 | 82 | 66 |
| | 78 | 90 | | 65 | 76 | 60 |
| | 84 | 85 | | 85 | 82 | 89 |
| | 82 | 99 | | 73 | 74 | 87 |
| | 80 | 98 | | 74 | 78 | 57 |
| | 78 | 65 | | 71 | 73 | |
| | 71 | 93 | | 80 | 62 | |
| | 78 | 104 | | 82 | 81 | |
| | 82 | 93 | | 78 | 65 | |
| | 70 | 92 | | 80 | 78 | |
| | 77 | 107 | | 75 | 76 | |
| | 73 | 95 | | 80 | 73 | |
| | 75 | 70 | | 72 | 108 | |
| | 77 | 80 | | 80 | | |
| | | 78 | | 72 | | |
| | | 95 | | 70 | | |
| Number of scallops | 73 | 75 | 30 | | 147 | 64 |
| Mean size | 76.5890411 | 90.34666667 | 83.4 | | 79.0472973 | 81.65625 |
| SD | 8.62463464 | 11.84299694 | 18.81232171 | | 11.24316751 | 12.68447213 |
| 95% CI | 1.978459216 | 2.680271641 | 6.731775056 | | 1.817514312 | 3.10763667 |

Appendix 24. Scallop size data collected in April 2006 from the ATNP.

| SITES | Reserve | | Control | |
|-------|---------|-----|---------|-----|
| | S1 | S3 | S4 | S5 |
| | 97 | 108 | 39 | 23 |
| | 80 | 108 | 92 | 76 |
| | 80 | 101 | 98 | 47 |
| | 82 | 89 | 93 | 76 |
| | 86 | 91 | 112 | 101 |
| | 83 | 107 | 86 | 106 |
| | 74 | 98 | 112 | 108 |
| | 86 | 76 | 93 | 113 |
| | 76 | 87 | 86 | 84 |
| | 90 | 46 | 87 | 74 |
| | 107 | 66 | 101 | 104 |
| | 93 | 76 | 119 | 86 |
| | 106 | 86 | 96 | 79 |
| | 97 | 87 | 98 | 107 |
| | 96 | 76 | 96 | 92 |
| | 91 | 107 | 92 | 86 |
| | 88 | 94 | 106 | 96 |
| | 86 | 97 | 119 | 106 |
| | 100 | 84 | 97 | 88 |
| | 77 | 98 | 88 | 97 |
| | 78 | 86 | 91 | 94 |
| | 86 | 99 | 89 | 89 |
| | 100 | 88 | 116 | 106 |
| | 77 | 102 | 96 | 101 |
| | 81 | 114 | 92 | 100 |
| | 78 | 88 | 94 | 42 |
| | 99 | 99 | 106 | 30 |
| | 72 | 82 | 86 | 48 |
| | 90 | 96 | 99 | 42 |
| | 84 | 88 | 87 | 76 |
| | 44 | 94 | 102 | 82 |
| | 79 | 98 | 39 | 78 |
| | 72 | 60 | 126 | 47 |
| | 94 | 77 | 116 | 102 |
| | 101 | 103 | 108 | 88 |
| | 83 | 97 | 112 | 91 |
| | 74 | 101 | 112 | 98 |
| | 94 | 82 | 109 | 102 |
| | 89 | 86 | 111 | 90 |
| | 91 | 97 | 90 | 84 |
| | 89 | 90 | 96 | 91 |
| | 89 | 96 | 90 | 92 |
| | 100 | 86 | 102 | 76 |
| | 103 | 96 | 98 | 90 |
| | 93 | 90 | 110 | 86 |
| | 92 | 82 | 117 | 104 |
| | 103 | 73 | 107 | 87 |
| | 80 | 90 | 87 | 100 |
| | 78 | 96 | 89 | 40 |
| | 94 | 101 | 102 | 46 |
| | 79 | 90 | 93 | 114 |
| | 97 | 86 | 82 | 111 |
| | 70 | 96 | 91 | 112 |
| | 87 | 71 | 84 | 89 |
| | 83 | 88 | 112 | 86 |
| | 82 | 93 | 90 | 87 |
| | 96 | 90 | 102 | 81 |
| | 86 | 82 | 109 | 94 |
| | 94 | 83 | 114 | 116 |
| | 93 | 93 | 104 | 104 |
| | 73 | 81 | 109 | 113 |
| | 88 | 83 | 94 | 106 |
| | 98 | 89 | 98 | 86 |
| | 80 | 100 | 79 | 83 |
| | 94 | 97 | | 96 |
| | 98 | 99 | | 86 |
| | 84 | 81 | | 81 |
| | 80 | 99 | | 82 |
| | 81 | 84 | | 102 |
| | 79 | 88 | | 90 |
| | 93 | 91 | | 94 |
| | 84 | 87 | | 91 |

Appendix 24 (cont). Scallop size data collected in April 2006 from the ATNP.

| | | | | |
|--|-----|-----|-----|-----|
| | 73 | 81 | 109 | 113 |
| | 88 | 83 | 94 | 106 |
| | 98 | 89 | 98 | 85 |
| | 80 | 100 | 79 | 83 |
| | 94 | 97 | | 95 |
| | 98 | 99 | | 85 |
| | 84 | 81 | | 81 |
| | 80 | 99 | | 82 |
| | 81 | 84 | | 102 |
| | 79 | 88 | | 90 |
| | 93 | 91 | | 94 |
| | 84 | 87 | | 91 |
| | 101 | 73 | | 77 |
| | 90 | 84 | | 92 |
| | 39 | 86 | | 74 |
| | 41 | 81 | | 90 |
| | 60 | 101 | | 83 |
| | 82 | 79 | | 89 |
| | 76 | 78 | | 97 |
| | 101 | 68 | | 87 |
| | 84 | 93 | | 89 |
| | 82 | 99 | | 97 |
| | 80 | 90 | | 97 |
| | 88 | 95 | | 45 |
| | 80 | 96 | | 36 |
| | 93 | 78 | | 51 |
| | 81 | 84 | | 101 |
| | 80 | 76 | | 94 |
| | 82 | 92 | | 86 |
| | 80 | 105 | | 82 |
| | 43 | 87 | | 84 |
| | 40 | 86 | | 110 |
| | 38 | 90 | | 82 |
| | 80 | 79 | | 75 |
| | 76 | 95 | | 65 |
| | 77 | 97 | | |
| | 39 | 108 | | |
| | 101 | 107 | | |
| | 80 | 90 | | |
| | 96 | 81 | | |
| | 80 | 88 | | |
| | 83 | 91 | | |
| | 80 | 96 | | |
| | 92 | 108 | | |
| | 76 | 98 | | |
| | 73 | 86 | | |
| | 71 | 89 | | |
| | 87 | 95 | | |
| | 81 | 79 | | |

Appendix 24 (cont.). Scallop size data collected in April 2006 from the ATNP.

| | | | | |
|---------------------------|--------|--------|-------|-------|
| | | 87 | | |
| | | 81 | | |
| | | 87 | | |
| | | 80 | | |
| | | 80 | | |
| | | 106 | | |
| | | 95 | | |
| | | 102 | | |
| | | 82 | | |
| | | 81 | | |
| | | 105 | | |
| | | 96 | | |
| | | 82 | | |
| | | 82 | | |
| | | 78 | | |
| | | 100 | | |
| | | 90 | | |
| | | 85 | | |
| | | 76 | | |
| | | 79 | | |
| | | 84 | | |
| | | 93 | | |
| | | 90 | | |
| | | 80 | | |
| | | 76 | | |
| | | 87 | | |
| | | 96 | | |
| | | 94 | | |
| | | 92 | | |
| | | 75 | | |
| | | 86 | | |
| | | 92 | | |
| <i>Number of scallops</i> | 109.00 | 141.00 | 64.00 | 95.00 |
| <i>Mean size</i> | 83.03 | 89.01 | 97.52 | 85.54 |
| <i>SD</i> | 14.43 | 10.54 | 15.07 | 20.06 |
| <i>SE</i> | 2.71 | 1.74 | 3.69 | 4.03 |

Appendix 25. Horse mussel and scallop density data collected from the Abel Tasman from 50 m² quadrats in March 2003.

| BarkBay (middle) | | | Bark Bay (north) | | |
|--------------------|--------------|---------|------------------|--------------|---------|
| Depth (m) | Horse mussel | Scallop | Depth (m) | Horse mussel | Scallop |
| 5 | 34 | 3 | 3 | 26 | 15 |
| 5 | 50 | 2 | 3 | 8 | 7 |
| 4.5 | 108 | 1 | 4 | 17 | 8 |
| 3.5 | 99 | 2 | 4 | 12 | 2 |
| 4 | 24 | 6 | 4 | 4 | 4 |
| 5 | 21 | 4 | 3.5 | 18 | 1 |
| 6 | 20 | 2 | 4 | 5 | 1 |
| 6 | 8 | 3 | 4 | 26 | 2 |
| 5 | 13 | 9 | 4 | 13 | 3 |
| 4 | 49 | 0 | 4 | 23 | 3 |
| Mean density | 42.6 | 3.2 | | 15.2 | 4.6 |
| Standard deviation | 35.00 | 2.62 | | 8.18 | 4.35 |
| Standard error | 11.07 | 0.83 | | 2.59 | 1.38 |

| Tonga (south) | | | Tonga (north) | | | Tonga (north-west) | | |
|--------------------|--------------|---------|---------------|--------------|---------|--------------------|--------------|---------|
| Depth (m) | Horse mussel | Scallop | Depth (m) | Horse mussel | Scallop | Depth (m) | Horse mussel | Scallop |
| 8 | 13 | 39 | 12 | 0 | 4 | 6 | 37 | 6 |
| 7 | 6 | 63 | 9 | 4 | 14 | 4 | 35 | 5 |
| 5 | 0 | 29 | 9 | 1 | 12 | 5 | 8 | 0 |
| 7 | 6 | 49 | 8 | 3 | 5 | 5 | 24 | 1 |
| 7 | 2 | 40 | 12 | 2 | 8 | 5 | 2 | 1 |
| 7 | 0 | 56 | 11 | 2 | 5 | 5 | 4 | 5 |
| 8 | 0 | 77 | 12 | 6 | 10 | 6 | 11 | 6 |
| 7 | 1 | 9 | 9 | 3 | 19 | 5.5 | 6 | 9 |
| 8 | 0 | 7 | 10 | 3 | 4 | 5.5 | 9 | 0 |
| 8 | 0 | 9 | | | | 6 | 11 | 10 |
| Mean density | 2.8 | 37.8 | | 2.67 | 9.00 | | 14.7 | 4.3 |
| Standard deviation | 4.32 | 24.31 | | 1.73 | 5.22 | | 12.70 | 3.65 |
| Standard error | 1.36 | 7.69 | | 0.55 | 1.65 | | 4.02 | 1.16 |

Appendix 26. Horse mussel and scallop density data collected from the Abel Tasman from 50 m² quadrats in April 2006.

| BarkBay (middle) | | | Bark Bay (north) | | |
|--------------------------|--------------|---------|--------------------------|--------------|---------|
| Depth (m) | Horse mussel | Scallop | Depth (m) | Horse mussel | Scallop |
| 5.5 | 3 | 5 | 5 | 4 | 4 |
| 6 | 2 | 16 | 5 | 3 | 2 |
| 6 | 12 | 6 | 5.5 | 6 | 0 |
| 7 | 5 | 6 | 6 | 2 | 0 |
| 6 | 3 | 9 | 6 | 1 | 2 |
| 7 | 8 | 4 | 6 | 3 | 5 |
| 7 | 3 | 4 | 5.5 | 2 | 3 |
| 6 | 5 | 9 | 5.5 | 0 | 2 |
| 5.5 | 5 | 12 | 5 | 4 | 3 |
| 6 | 3 | 10 | 5 | 2 | 5 |
| Mean (50m ²) | 4.90 | 8.10 | Mean (50m ²) | 2.70 | 2.60 |
| SD | 3.03 | 3.87 | SD | 1.70 | 1.78 |
| SE | 0.96 | 1.22 | SE | 0.54 | 0.56 |

| Tonga (south) | | | Tonga (north-west) | | |
|--------------------------|--------------|---------|--------------------------|--------------|---------|
| Depth (m) | Horse mussel | Scallop | Depth (m) | Horse mussel | Scallop |
| 8 | 13 | 33 | 10 | 19 | 15 |
| 8 | 8 | 27 | 10 | 20 | 22 |
| 8 | 5 | 39 | 11 | 6 | 18 |
| 9 | 15 | 11 | 11 | 10 | 11 |
| 8 | 3 | 38 | 11 | 16 | 18 |
| 8.5 | 18 | 20 | 11 | 1 | 6 |
| 7.5 | 4 | 31 | 10.5 | 2 | 4 |
| 7 | 5 | 42 | 11 | 13 | 8 |
| 8 | 4 | 47 | 10.5 | 1 | 3 |
| 8.5 | 11 | 28 | 10.5 | 7 | 5 |
| Mean (50m ²) | 8.60 | 31.60 | Mean (50m ²) | 9.50 | 11.00 |
| SD | 5.32 | 10.73 | SD | 7.26 | 6.82 |
| SE | 1.68 | 3.39 | SE | 2.30 | 2.16 |